Research Article

Sorption of As(V) from aqueous solution using in situ growth MgAl–NO₃ layered double hydroxide thin film developed on AA6082



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Abstract

In this work, the MgAl–NO₃ layered double hydroxide (LDH) developed by the single-step in situ growth method is used as a robust sorbent to remove arsenic from aqueous solution. The MgAl-LDH exhibiting two different distinct morphologies (platelet structure and cauliflower-shaped structure) was developed on the AA6082 substrate with the variation in synthesis parameters, where AA6082 specimen acts as both the reactant and support. The structural characterizations were investigated through scanning electron microscopy, X-ray diffraction analysis, and energy dispersion spectroscopy, while the adsorption of arsenic on MgAl-LDH was studied through Langmuir and Freundlich models. The Langmuir isotherms have shown a maximum adsorption capacity of around 213 and 239 mg/g for platelet and cauliflower-like MgAl-LDH, respectively. The pseudo-first-order and pseudo-second-order Lagergren kinetic models were studied for the understanding of the adsorption kinetics. The results depicted that anion exchange and the electrostatic interaction are the possible reasons of arsenic sorption on MgAl-LDH, but the ion exchange mechanism is found to be the dominant mechanism. The maximum adsorption capacity of cauliflower-shaped MgAl-LDH was found to be slightly higher than platelet structure, but overall maximum arsenic adsorption uptake values of both in situ growth structures have found to be exceeded the mostly reported MgAl-LDH maximum adsorption capacities.

Keywords In situ growth MgAI-LDH · AA6082 · Arsenic removal · Kinetics · Adsorption isotherm

1 Introduction

Arsenic contaminations in groundwater are considered a serious hazard for human health as well as for ecosystem due to their toxicity and carcinogenic nature [1]. According to the World Health Organization and the United States Environmental Protection Agency (USEPA) provided guideline, the maximum arsenic level should be less than 10 μ g/L in the drinking water [1, 2]. In that scenario, it is very alarming to found a much higher concentration of arsenic in drinking water in different regions throughout the world due to the presence of geological resources of arsenic in that areas and furthermore also due to anthropogenic pollution from different industrial processes. It is

estimated that 150 million people in 70 different countries are at the risk of arsenic contaminations [3], while the rapid increase in industrial pollutants worldwide is causing more discharge of non-biodegradable heavy metals [4]. Therefore, it is indispensable to develop low-cost efficient water purification technologies to purify drinking water from heavy metals.

A number of approaches have been reported to remove arsenic, for instance, solvent extraction electrodialysis, membrane technologies, adsorption, coagulation, and ion exchange with certain limitations. Adsorption techniques among them are considered more promising due to their regeneration capability, design simplicity, operational easiness, and small-scale treatment plants applicability

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[5]. Recently, layered double hydroxide has got prominent attention as adsorbents due to their low cost, high anion exchange capability, greater surface area, nontoxic nature of composing elements, and also the availability of a wide range of cationic and anionic combinations [6-8]. Layered double hydroxide, a promising type of chemical conversion material, also known as hydrotalcite-like compounds or anionic clays, has got prominent attention in the fields of biomedical science, applied chemistry, coatings, and recently as environmental pollutants purifiers, to remove heavy metals from the water [9–11]. Generally, two different methods are employed to fabricate layered double hydroxide: single-step in situ approach [8, 12] and colloidal assembly technique [13, 14]. This brucite-like structure contains positively charged layers being bonded together via charge-compensating anions and water molecules and thus acts as an organized structural material to facilitate effective charge transfer and also yields a comprehensive ion exchange area, which makes it enable for anion exchange with various active metal units [8]. The general process involved during LDH adsorption is electrostatic interaction, anion exchange, and physical adsorption [15, 16]. Among various types of developed layered double hydroxides, MgAI-LDH intercalated with NO₃⁻¹ is known as effective adsorbents for the removal of As(V) from the aqueous solutions [17]. Although many studies have been done on the MgAI-LDH adsorptive properties [18-20], yet to the best of our knowledge, there has been no report on the in situ growth MgAl-LDH adsorptive behaviors for the arsenic removal and role of surface morphology on LDH ion exchange capabilities. We studied, in this work, two distinct LDH morphologies, i.e., platelet structure and cauliflower-shaped structure, to investigate arsenic adsorption behavior of in situ growth LDH.

In this contribution, MgAI-LDHs intercalated with nitrate anions were synthesized by the in situ growth method and their effectiveness on the arsenic adsorption behavior is investigated. The effectiveness of the arsenic removal is investigated by the adsorption isotherms and kinetics studies, while the structural characterization was investigated by XRD, SEM, and EDS analyses. The LDH film thickness, surface morphologies, and synthesis parameters have found to be an influential effect on the LDH adsorption process and their ion exchange capability.

2 Experimental

2.1 Materials

The AA6082 extruded bar was purchased from Meal Center (Italy), which mainly consists of magnesium (0.60-1.20%), iron (0.50%), silicon (0.70-1.30%), copper (0.10%),

manganese (0.40–1.00%), and of balance percentage of aluminum. The Mg(NO₃)₂·6H₂O (purity 98%) NH₄NO₃ (purity 95%), and NaOH (purity \geq 98%) were purchased from Sigma-Aldrich Corporation.

2.2 Synthesis of MgAl-layered double hydroxide film

The in situ growth layered double hydroxide film is developed on the AA6082 substrate (3.14 cm²), based on our previous work [8]. The specimens were initially ground with silicon carbon paper, started from 500-grit paper to 1000-, 2400-, and 4000-grit paper, respectively. The samples were cleaned with the deionized water and further with pure ethanol ultrasonically for 15 min. Lastly, the specimens were immersed in 0.1 M aqueous NaOH solution for 1 min to etch the oxide layer on the surface of the alloy. From our previous study [8], we selected two different synthesis conditions to develop distinct MgAI-LDH morphologies, i.e., cauliflower-like and platelet-shaped structure. The following combinations were used for that purpose: 0.008 MgNO₃ & 0.048 NH₄NO₃ and 0.028 MgNO₃ & 0.084 NH_4NO_3 at 60 °C and 80 °C, respectively, at a pH of 10 for 24 h of crystallization time. After that, the developed coated specimens were dried at room temperature.

3 Characterization

The structural characterization of MgAl-LDH film is carried out by the SEM (JEOL-IT300 microscope equipped with an EDS detector), while XRD patterns of the calcined and uncalcined LDH-coated samples were recorded by X-ray diffraction (XRD) (X'Pert HighScore diffractometer, Rigaku, Japan) by using cobalt K- α (λ = 1.789 Å⁻¹) emission source at 10 mA and 30 kV conditions.

The arsenic concentrations in the solution were determined by the atomic absorption spectrophotometer [AAS vario 6, Analytic Jena (Germany)]. To investigate the adsorption behavior, MgAI-LDH specimens were dipped in 20 mL model solutions of various arsenic concentrations (0.01–0.1 M) for 24 h. To determine adsorption kinetics, 0.1 M arsenic solution (500 mL) was used to investigate adsorption behavior as a function time at room temperature. The Lagergren pseudo-first-order and pseudo-second-order fitting models were employed to understand the mechanism of As(V) sorption on MgAI-LDHs.

4 Results and Discussion

Figure 1 shows the XRD patterns of MgAl-LDHs before and after the As(V) adsorption. It can be seen that as-prepared as well as treated specimens have shown the characteristic

peaks of layered double hydroxide structure, as described in the literature [21]. The characteristic peaks of (003), (006), (009) crystal planes are attributed to the presence of MgAl–NO₃ LDH structure. The sharp and symmetric characteristic peaks demonstrated well-crystalline and ordered structure of MgAI-LDH. The LDH structural peaks after As(V) adsorption have shown the similar original LDH characteristic peaks, which describe the intactness of layered double hydroxide structure which was further confirmed by the SEM analysis. The intense reflection peaks of (003) at low 2θ value exhibited the interlayer distance (d_{003}) of \approx 0.90 nm and \approx 0.89 nm for cauliflower and platelet LDH structure, respectively. The comparatively high interlayer thickness of cauliflower LDH structure may be attributed to strong NO_3^{-1} intercalation inside the galleries of brucite-like structure. However, due to the formation of carbonate anions, a small absorption peak of d_{003} for MgAl-LDH platelet structure can also be observed. The intensity and broadness of the reflection peaks vary after the As(V) intercalations, which indicates the anion exchange interaction between As(V) in the solution and nitrates in the LDH galleries. After the anion exchange, the d_{003} value decreased, possibly due to the new orientation of As(V) inside the layers, the water molecules reorientation during the anion exchange process, and also due to the HAsO₄⁻² ionic radius; moreover, reduced interlayer distance described the strong electrostatic interaction between the positive layers and the intercalated anions [22]. As revealed from the SEM observations (Figs. 2 and 3), compact and uniform interviewed LDH structures were obtained, which

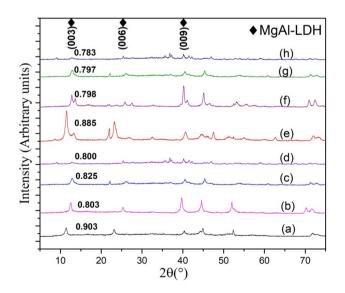


Fig. 1 XRD patterns of cauliflower and platelet MgAl-LDH before and after treatment with different sodium arsenate concentrated solutions. (a) Cauliflower MgAl-LDH as prepared, (b) 0.06 M, (c) 0.08 M, (d) 0.1 M, (e) platelet MgAl-LDH as prepared, (f) 0.06 M, (g) 0.08 M, (h) 0.1 M. (Interlayer distance (d_{003}) is listed with each peak)

retained almost their original shape after the adsorption, but there can also be seen a few aggregations of particles on the MgAl-LDH surface. So we can assume that during the adsorption, two possible interactions took place: (1) physical interaction of LDH surface with the As(V) and (2) anion exchange process between the nitrates and As(V). Figure 4 shows the EDS findings after treatment of MgAl-LDH with a 0.1 M arsenic solution, and it is clear that after treatment MgAl-LDH structure mainly contains Mg, Al, As, and oxygen, while the small amount of carbon was also evident in EDS study, caused by the contaminations, which results in the formation of carbonate anions.

The arsenic adsorption isotherms (adsorption capacity) were measured by varying the initial arsenic concentrations in the model solution for a contact period of 24 h, as shown in Fig. 5. The amount of adsorption is found to increase with the increase in arsenic concentration in the solution (from 0.02 to 0.1 M) at a pH of 7. It is also interesting to note that cauliflower-like LDH structure has shown the highest arsenic adsorption value (239 mg/g) compared to that of platelet structure (206 mg/g), possibly due to its high film thickness and of greater contoured nanosheets which relatively increase its surface area. But overall, both structures have effectively removed the arsenate from the water. The equilibrium arsenic adsorption amount is calculated by Eq. (1) [23]:

$$q = \frac{V(C_o - C_t)}{m},\tag{1}$$

where "V" is the volume of the solution in L, C_o is the initial arsenic concentration, while C_t is the concentration at time t in mg/L. q (mg/g) is the amount of arsenic sorbed (mg/g) at a time "t" on MgAl-LDH, and "m" is the sorbent mass (MgAl-LDH film) in grams calculated by the mass difference of before and after MgAl-LDH formation on AA6082 substrate. (The specimens were dried at 60 °C for 40 min before mass measurement.) Figure 5 shows the graphical relationship between q_e and $C_{e'}$ and it is clear that the adsorption capacity of MgAl-LDH structure increases with the increase in amount of As(V) in the model solution, because of larger concentration gradient which causes an increase in mass transfer driving force and leads to the availability of more arsenate molecules on the MgAl-LDH active sites.

Further, the results of adsorption experiments were fitted by using the Langmuir model and Freundlich model to analyze the adsorption behavior of arsenate on MgAl-LDH. The general assumption of the Langmuir model is the uniform adsorption process with the context that each adsorption surface consisted of identical binding sites with

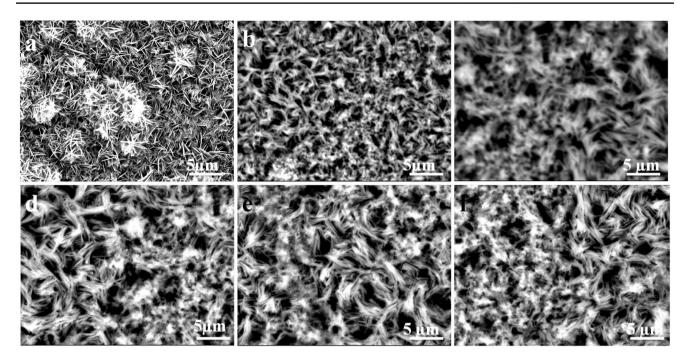


Fig. 2 SEM images of platelet MgAl-LDH film with adsorption in different (Na₃AsO₄) concentration solutions for 24 h, **a** as-prepared LDH, **b** 0.02 M, **c** 0.04 M, **d** 0.06 M, **e** 0.08 M, **f** 0.1 M

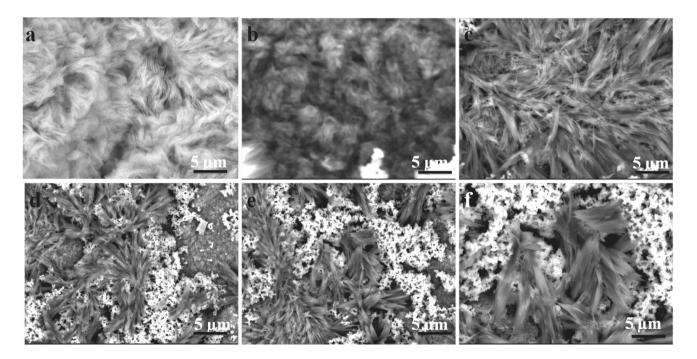


Fig. 3 SEM images of cauliflower MgAI-LDH film with adsorption in different (Na₃AsO₄) concentration solutions for 24 h, **a** as-prepared LDH, **b** 0.02 M, **c** 0.04 M, **d** 0.06 M, **e** 0.08 M, **f** 0.1 M

monolayer adsorption behavior. The linear form of the Langmuir model can be written as follows:

$$\frac{C_e}{q_e} = \frac{1}{b \times Q_0} + \frac{C_e}{Q_0},\tag{2}$$

where C_e/q_e (mg/g) is the equilibrium ratio between the concentration of adsorbate ions and the number of adsorbate ions adsorbed on per unit weight of MgAl-LDH, Q_0 is associated with the maximum adsorption capacity, while "b" is termed as Langmuir constant. If C_e/q_e is plotted

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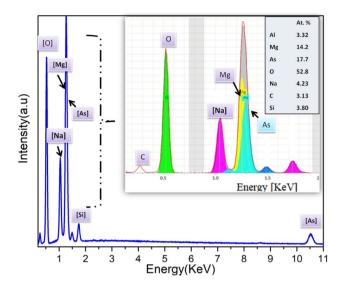


Fig. 4 EDS spectrum of MgAI-LDH (cauliflower) after treatment with 0.1 M sodium arsenate concentrated solution

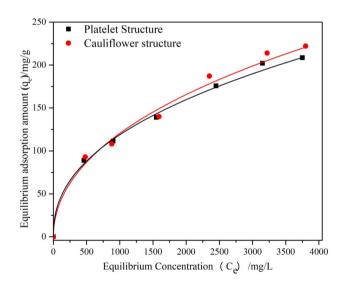


Fig. 5 Equilibrium adsorption isotherms of MgAl-LDHs by varying the initial As(V) concentration in the arsenic solution

against $C_{e^{i}}$ then from the slope and intercept of this plot, Q_{0} and b can be determined, respectively (Fig. 6).

On the other hand, Freundlich model is another empirical description of non-ideal adsorption on the heterogeneous surface and provided an idea about the multilayer sorption with the description that the binding surface energy reduces with the increase in solute adsorbed on the surface. The Freundlich model is expressed as:

$$\log q_e = \log k + \frac{1}{n} \log C_e. \tag{3}$$

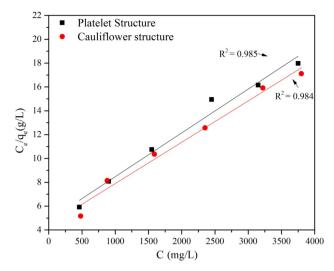


Fig. 6 A highly linear Langmuir adsorption isotherm of As(V) by MgAl-LDHs

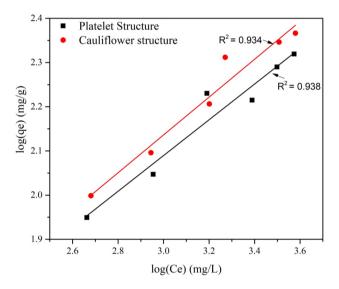


Fig. 7 Freundlich isotherm of As(V) by MgAl-LDHs

By plotting log q_e versus log C_e , the coefficients "k" and "n" can be determined if a straight line is obtained. Freundlich isotherms for As(V) adsorption on MgAl–NO₃ LDHs are shown in Fig. 7. The correlation coefficient values and isotherm constants of Langmuir and Freundlich isotherms are listed in Table 1.

From the obtained fitting results, it can be seen that the maximum adsorption capacity of As(V) was found by the Langmuir model with high linearity (Table 1). The maximum adsorption capacities of cauliflower and platelet structure for Langmuir As(V) adsorption isotherms are found to be 239 and 212 mg/g, respectively. The results obtained by the Langmuir model are also in

Table 1	Langmuir and Freundlich	parameters for As(V) removal by MgAI-LDHs	
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Langmuir model					
R^2	Q _o (mg/g)	(L/mg)	Slope		
0.985	212.8	0.0010	0.0047		
0.984	238.1	0.0008	0.0042		
Freundlich model					
R^2	К	1/ <i>n</i>	Slope		
0.994	6.622165	0.42	0.416		
0.998	7.063176	0.42	0.421		
	$ \frac{1}{R^2} $ 0.985 0.984 Freundlich model $ \frac{1}{R^2} $ 0.994	R^2 $Q_o (mg/g)$ 0.985 212.8 0.984 238.1 Freundlich model R^2 R^2 K 0.994 6.622165	R^2 $Q_o (mg/g)$ (L/mg) 0.985 212.8 0.0010 0.984 238.1 0.0008 Freundlich model R^2 K $1/n$ 0.994 6.622165 0.42		

Table 2A comparison studywith other reports to adsorbAs(V) by MgAI-LDHs

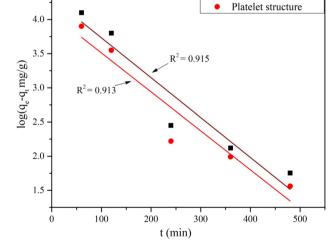
Layered double hydroxide (adsorbents)	Synthesis method	Adsorption capacity (mg/g)		
		As(V)	References	
MgAl–NO ₃ -LDH (cauliflower structure)	In situ	239	This work	
MgAl–NO ₃ -LDH (platelet structure)	In situ	212	This work	
MgAINO ₃ -LDH	Co-precipitation	143	Rahman et al. [18]	
MgAINO ₃ -LDH	Co-precipitation	77	Rahman et al. [18].	
MgAINO ₃ -LDH	Co-precipitation	100	Doušová et al. [24]	
MgAICO ₃ -LDH	Co-precipitation	45	Wu et al. [25]	
MgAICI-LDH	Co-precipitation	88	Wu et al. [25]	
MgAINO ₃ -LDH (calcined)	Co-precipitation	216	Yu et al. [<mark>26</mark>]	
MgAICO ₃ -LDH	Co-precipitation	121 Huang et al. [20		

4.5

close agreement with the results obtained by Eq. (1). So we can say that the cauliflower-like MgAl-LDH structure has shown slightly better performance than the platelet-like structure. The results obtained here are also relatively higher than the MgAl-LDH synthesized by co-precipitation method, from the literature studies (Table 2) [18, 20, 24–26].

5 Kinetics of As(V) removal

Figures 8 and 9 show the adsorption of arsenate by MgAl–NO₃ LDHs as a function of time to understand the kinetics of As(V) removal. With the passage of time, As(V) concentration decreases in the solution which caused the decline in mass transfer due to lowering of static driving forces. Therefore, we can say that the adsorption is rapid initially at the start of the process when there was a high gradient of concentration and then it decreased with the passage of time until the system approaches the equilibrium. For the kinetics modeling and understanding of the mechanism of arsenate adsorption, Lagergren pseudo-first-order and pseudo-second-order models were studied. Lagergren equation is used [(Eq. (4)] to understand the adsorption kinetics, and in the case of a pseudo-first-order process, the Lagergren equation after integration with



Cauliflower structure

Fig. 8 Pseudo-first-order model for As(V) removal by MgAl-LDHs

the conditions $(q_t = 0 \text{ at } t = 0, q_t = q_t \text{ at } t = t)$ is expressed in Eq. (5), where q_t is the amount of arsenate adsorbed in mg/g at time t, k_1 is the time constant (min⁻¹), and q_e is the equilibrium adsorption value in mg/g at time t (calculated by the extrapolation of the experimental data). Equation 6 expresses the pseudo-second-order Lagergren equation, where k_2 is the second-order rate constant in g/mg × min.

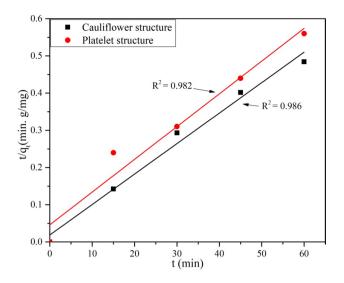


Fig.9 Pseudo-second-order model for $\mathsf{As}(\mathsf{V})$ removal by MgAl-LDHs

The pseudo-first-order kinetic model provides a general idea about the rate of change in arsenate concentration with the passage of time and stated that the change in the concentration of the adsorbate and the amount of adsorbent over time is logarithmically proportional [27], while a pseudo-second-order kinetic model suggests that the number of active sites occupied on the adsorbent is directly proportional to the adsorption capacity. The pseudo-first-order [28] [Eq. (5)] and pseudo-second-order [Eq. (6)] [29] equations are described as:

$$\frac{dq_t}{d_t} = k_1 (q_e - q_t) \tag{4}$$

$$\ln\left(q_e - q_t\right) = \ln\left(q_e\right) - k_1 t \tag{5}$$

$$\frac{1}{q_t} = \frac{1}{k_2(q_e^2)} + \frac{t}{q_e}.$$
(6)

The correlation coefficients (R^2 values), rate constants, and other parameters for these two kinetic models were calculated from the respective slope and intercept and are tabulated in Table 3. The pseudo-first-order model shows a lower correlation (low linearity) than

Table 3Kinetic parameters forAs(V) removal by MgAI-LDHs

the pseudo-second-order model, which indicates the pseudo-second-order model is a better approach for the As(V) removal kinetics. Due to the second-order reaction kinetics, we can say that the ion exchange of nitrate with As(V) is the rate-determining step [30]. Therefore, we can conclude about the As(V) mechanism that the reaction rate of arsenate adsorption mainly depends on the anion exchange process between nitrate ions and As(V) and also that the rate of reaction depends upon the concentrations of NO_3^{-1} ions bound to the MgAl– NO_3 -LDH and As(V) ions in the model solution. It can also be concluded that the anion exchange process effectively takes place between the NO₃ and HAsO $_4^{-2}$, due to the relatively higher charge density of HAsO₄⁻² ions. The SEM and XRD analyses already confirmed that the basic LDH structure remains almost intact, which is aligned with the above description of the mechanism.

Figure 10 shows the as-prepared MgAl-LDH film thickness along with respective EDS analysis in at.% in-plane

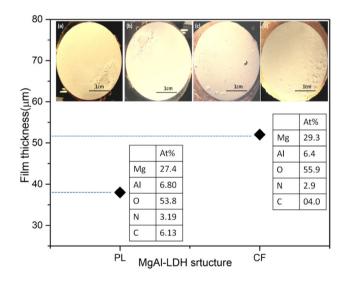


Fig. 10 MgAI-LDH thickness of the as-prepared films; *PL* platelet structure, *CF* cauliflower structure along with EDS atomic percentage, insight is the specimen surfaces after treatment with sodium arsenate for 24 h, platelet structure (a) 0.08 M, (b) 0.1 M, cauliflower-like structure, (c) 0.08 M, (d) 0.1 M

Kinetic model	LDH morphology	Q _e (mg/g)	R ²	<i>K</i> ₁	Slope
Pseudo-first-order	Platelet	99.86194	0.913	0.0057	0.0057
	Cauliflower	102.3293	0.915	0.0058	0.0058
	LDH morphology	$Q_2 (\mathrm{mg/g})$	R ²	K ₂	Slope
Pseudo-second-order	Platelet	113.6364	0.986	0.004188	0.0088
	Cauliflower	122.1001	0.982	0.001454	0.0082

SN Applied Sciences A SPRINGER NATURE journat scanning mode, while the thickness is around 38 ± 0.98 and $52 \pm 1.06 \,\mu$ m for platelet and cauliflower-like structure, respectively, measured by the cross-sectional analysis of the film. Figure 10 (insets) shows the MgAl-LDH surface, just for representation, and after treatment with 0.1 M sodium arsenate solution for 24 h, the MgAl-LDH film has sustained the initial structure and no dissolution was observed on contact with the arsenic solution.

6 Conclusion

From the obtained findings in this study, it can be concluded that surface morphologies of MgAI-LDH have an influential effect on the arsenic adsorption capacity, and here, cauliflower-like structure has shown slightly better adsorption capacity (238 mg/g) compared to the platelet structure (212 mg/g). The results also showed that the Langmuir model has shown the better fitting of the adsorption capacity, while the Lagergren pseudosecond-order equation better translates the kinetics of arsenic adsorption. From the results, it is clear that arsenic adsorption occurs due to (1) surface adsorption and (2) by the anion exchange of NO_3^{-1} ions with the As(V), but the anion exchange process of NO₃ and As(V) is the main arsenic adsorption process. The in situ growth MgAINO₃-LDH adsorption capacity of arsenate was found to be higher as compared to the MgAINO₃-LDH developed by the coprecipitation method (literature study), possibly due to the higher film thickness, high surface area, and layer-by-layer uniform structural geometry of in situ growth MgAl-LDH. This work could be useful also as a baseline for the possible utilization of such systems as a coating material for the removal of arsenic.

Compliance with ethical standards

Conflicts of interest The authors declare no conflict of interest.

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