



# Biomass allocation in five semi-arid afforestation species is driven mainly by ontogeny rather than resource availability

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## Abstract

• **Key message** The changes in the relative biomass allocation to roots in juvenile stands of fast-growing (*Leucaena leucocephala* Lam., *Moringa oleifera* Lam., and *Jatropha curcas* L.) and slow-growing (*Anacardium occidentale* L. and *Parkia biglobosa* Jacq.) afforestation species are driven mainly by ontogeny rather than resource availability. However, silvicultural management aiming at increasing availability of water and particularly nutrients enhances biomass production in all species.

• **Context** Understanding the patterns of biomass allocation among tree species in response to ontogeny and to variation in resource availability is key to the successful restoration of degraded land using forest plantations.

• **Aims** This study assessed the effects of resource availability and ontogeny on biomass accumulation and partitioning in five semi-arid afforestation species.

• **Methods** The aboveground and belowground biomass production of fast-growing *Leucaena leucocephala* Lam., *Moringa oleifera* Lam., and *Jatropha curcas* L. and slow-growing *Anacardium occidentale* L. and *Parkia biglobosa* Jacq. was monitored following the application of manure (1 kg plant<sup>-1</sup>) and/or supplemental irrigation (0.5 L per sapling daily) during the first two rainy seasons and the intervening dry season on degraded cropland in Northern Benin.

• **Results** Biomass accumulation in the fast-growing species was positively impacted by fertilization and irrigation during both rainy seasons. The slow-growing species responded positively to the silvicultural treatments during the dry and second rainy season. The application of fertilizer alone increased the biomass of *P. biglobosa* by up to 335% during the dry season. Fifteen months after planting, manure-treated *L. leucocephala* accumulated the most biomass (2.9 kg tree<sup>-1</sup>). The root fraction decreased with increasing tree size in all species. The comparison of root versus shoot allocation in trees of equal size indicated that the treatment-induced shifts in biomass partitioning were controlled by ontogeny, which explained 86–95% of the variation in root-shoot biomass relationships.

• **Conclusion** While ontogeny was the main driver of biomass partitioning, increased resource availability induced a larger production of biomass, overall leading to greater aboveground production in all species.

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**Contribution of the co-authors** All authors conceived and designed the experiment. FN and JBN collected the field data. FN performed statistical analyses and composed the manuscript. AK, JBN, and JPAL contributed to the development of the manuscript.

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## 1 Introduction

Conserving and restoring natural forests is essential in coping with the growing demand for timber and non-timber forest products and ensuring environmental sustainability in tropical regions (Steege et al. 2015). In sub-Saharan Africa (SSA), these efforts are increasingly constrained by the decline in the per capita availability of agricultural land (de Graaff et al. 2011), required to satisfy the food, feed, and fuel demands of the ever-increasing population. The resulting pressure on marginal land (Azuka et al. 2015) has dramatically reduced the duration of fallows, the traditional practice used to restore soil fertility (Nandwa 2001). The increasing variability

of rainfall in SSA (Sylla et al. 2016) has become an additional obstacle for efforts to improve cropland productivity and regenerate forest areas. In this regard, the afforestation of degraded cropland and intercropping with multi-purpose tree species offer the possibility of increasing on-farm tree cover, reversing land degradation, and contributing to rural livelihoods (Chamshama and Nduwayezu 2002; Garrity et al. 2010; Khamzina et al. 2012).

Though afforestation is a promising strategy, nutrient and water limitations on degraded land may hamper its success. Intensive silviculture, including fertilization, irrigation, pest control, and weed control, has been successful in overcoming environmental constraints and increasing forest production (e.g., Fening and Gershenson 2002; Mead 2013; Coyle et al. 2016). However, this is yet to be adopted in most dryland afforestation systems due to insufficient incentives for farmers to invest in ecological restoration (e.g., Djalilov et al. 2016) but is practiced for tree species of commercial importance, such as *Jatropha curcas* L. in biofuel production (Baumert et al. 2016). For instance, fertilization alone mitigated the negative effects of nutrient stress, resulting in enhanced stem growth (6–30%) in *J. curcas* and *Parkia biglobosa* Lam. saplings on nutrient-poor soils under semi-arid conditions (Noulèkoun et al. 2017) and greater aboveground biomass (AGB) accumulation (125–200%) in two 3-year-old *Populus deltoides* genotypes in a temperate climate (Coyle and Coleman 2005). Irrigation also improved tree productivity under scarce and erratic rainfall (Khamzina et al. 2008; Trichet et al. 2008; Noulèkoun et al. 2017). Early growth responses of trees to silvicultural treatments are reflected in the adjustment of biomass partitioning between roots and shoots. This relative allocation of biomass can be indicative of the physiological potential and long-term growth trajectory of tree species (e.g., Niklas 1994; McConnaughay and Coleman 1999; Reich 2002; Lamers et al. 2006). Critical physiological processes regulating resource acquisition and plant growth under nutrient and/or water stress occur belowground (Coyle and Coleman 2005). Hence, the root share can account for a substantial portion of total biomass (27–68%) in (sub-)tropical dry forests or plantations (Mokany et al. 2006).

Variations in belowground biomass (BGB) proportions during tree growth are controlled by both resource availability and development (i.e., ontogeny) (Coyle et al. 2008, 2016). Increased resource availability (e.g., nutrient and water) enhances AGB accumulation (which is usually of interest to farmers) as a result of allocation shifts from root to shoot (e.g., Albaugh et al. 1998; McConnaughay and Coleman 1999; Coleman et al. 2004; Coyle et al. 2016). Changes in the relative biomass allocation to roots in

response to resource availability are predicted to accord with the optimum partitioning theory (OPT), which suggests that plants allocate more biomass to the organ that acquires the most limiting resource (Brouwer 1963; Bloom et al. 1985). However, studies comparing plasticity in root/shoot ratios at common plant size or development stage (not at common plant age or time) have suggested that the shifts in biomass partitioning cited in support of the OPT may largely be induced by accelerated development (e.g., McConnaughay and Coleman 1999; Coleman et al. 1994, 2004). Therefore, it is important to separate the dynamics of development-induced (i.e., ontogenetic drift) changes in biomass allocation from those modulated by silvicultural treatments (i.e., “true” plasticity in allocation) when identifying the controlling factors (McConnaughay and Coleman 1999; Reich 2002; Coleman et al. 2004).

Multi-species plant studies, aiming at distinguishing between biomass allocation changes that result from ontogenetic drift and true adjustment, are scarce in tropical climates, particularly in the context of the afforestation of degraded drylands. This is partly because the required sequential sampling of AGB and BGB is challenging in open-field research (Robinson 2004). We assessed biomass allocation patterns in five afforestation species grown on degraded cropland under varying levels of nutrients and water supply during early stand development in Northern Benin, West Africa. Three of the tested species (*Leucaena leucocephala* Lam., *Moringa oleifera* Lam., and *J. curcas*) were previously classified as fast-growing and two (*Anacardium occidentale* L. and *P. biglobosa*) as slow-growing species, based on the response of root-level and shoot-level morphological traits to nutrient and water amendments (Noulèkoun et al. 2017). Yet, the previous studies did not address the dynamics of plant aboveground and belowground responses to silvicultural treatments with emphasis on biomass production and allocation. We hypothesized that (i) increased resource availability boosts biomass production significantly more in species characterized by faster ontogenetic development than in slow growers, (ii) biomass partitioning to belowground decreases with ontogeny, and (iii) when ontogeny is accounted for as a function of plant size, increasing resource availability triggers greater biomass allocation to aboveground versus belowground.

## 2 Materials and methods

### 2.1 Study site description

The study site, plant materials, and experimental design were previously described in detail by Noulèkoun et al. (2017). The research was carried out in Pouri village (N 10° 54' 8.4" and E 1° 4' 47.4") located on the periphery of the Dassari catchment in the department of Atacora, Northern Benin. The area is

characterized by a semi-arid climate, marked by one dry season (November to March) with temperatures rising as high as 45 °C and one rainy season (April to October) with temperatures falling to 14 °C. The average annual precipitation for the 15-month study period, spanning the 2014 and 2015 rainy (growing) seasons, was 795 mm. The vegetation is typical for a Sudan savanna ecological zone, with parklands, forest patches, and agricultural and fallow land, all of which are impacted by human activity (Saïdou et al. 2004; Chabi et al. 2016). The soils are compacted, limited in their depth by gravel and lateritic formations, and have low to moderate inherent fertility (Saïdou et al. 2004; Azuka et al. 2015). The soil at the research site was classified as an epipetric plinthosol with sandy-loamy texture in the surface layer and loamy soil dominated by clayey loam in deeper layers. Total carbon and NPK measured 0.5%, 0.1%, 1.2 mg kg<sup>-1</sup>, and 44.3 mg kg<sup>-1</sup>, respectively.

## 2.2 Woody species

The afforestation trial established at the degraded cropping site in July 2014 included five (semi-)deciduous tree species that differed in terms of their growth potential and tolerance to drought (Noulèkoun et al. 2017): the fast-growing *Moringa oleifera* Lam. (drumstick tree, Moringaceae), *Jatropha curcas* L. (physic nut, Euphorbiaceae), and *Leucaena leucocephala* Lam. (leucaena, Fabaceae) and the slow-growing *Anacardium occidentale* L. (cashew, Anacardiaceae) and *Parkia biglobosa* Jacq. (African locust bean, Fabaceae). Seeds were germinated in a local nursery and planted at the experimental site after 2.5–4 months.

## 2.3 Experimental design

Three experimental factors were considered to influence plant growth and biomass production and allocation: (i) the five species and two silvicultural treatments consisting of (ii) drip irrigation and (iii) fertilization by manuring. The species and irrigation variables were subject to a fully factorial design with three repetitions, while manuring was nested within plots. The design thus consisted of a total of 30 pure species plots, each 12 × 8 m<sup>2</sup> in size. Each plot contained 48 seedlings spaced at 2 × 1 m, resulting in an initial density of 5000 plants ha<sup>-1</sup>.

The irrigation and fertilization treatments resulted in four resource management options: no resource manipulation (control, C), irrigation (I), fertilization (F), and irrigation plus fertilization (IF). An adapted drip irrigation system (Fig. 1) was used during the dry season to supply 0.5 L of water per sapling daily to meet 30% of the evaporative demand (Allen et al. 1998). This amounted to a total of 72.5 mm of irrigation. In addition, saplings received 33 mm of rainfall during the 2014–2015 dry season. Manure was supplied in the amount of 1 kg per plant in both growing seasons, i.e., at 2.5 and

11 months after planting (MaP). The manure was composted cow and pig dung and contained on average 18.7% carbon (C), 1.2% nitrogen (N), and 0.8% phosphorus (P).

## 2.4 Biomass production

Plants were sampled at 5, 11, and 15 MaP, representing the first growing (rainy) season (0–5 MaP), the subsequent dry season (5–11 MaP), and the second growing season (11–15 MaP). At each sampling, two to six saplings were randomly selected from each experimental plot and harvested entirely. The plants were felled, and the AGB separated into leaves (including the petiole), stems with branches, and (if present) reproductive parts. Roots were manually excavated and sectioned into coarse ( $\phi > 2$  mm) and fine ( $\phi < 2$  mm) fractions. The roots were gently washed on a sieve and freed of soil. The fresh mass of each fraction was recorded in the field using a 7-kg portable scale (accuracy 0.1 g) and oven-dried at 75 °C until a constant weight. Woody samples were cut into small pieces before oven drying.

At 11 MaP, biomass was also sampled from a 10-year-old natural fallow, located in the vicinity of the experiment to compare biomass productivity between afforested (active intervention) and fallowed (natural succession) sites. An area of 0.5 ha was delineated in the fallow land where three sub-plots of 4 × 4 m each were installed along one of the diagonals for sampling. Grazing was not controlled. The AGB and BGB of the predominant vegetation, consisting of shrubs (e.g., *Combretum glutinosum*) and trees (e.g., *Terminalia macroptera*), were quantified for each species and sub-plot as described above.

## 2.5 Statistical analyses of treatments

A linear mixed-effect model was used to assess the effects of the abovementioned experimental factors on biomass production at 5, 11, and 15 MaP. The effect of irrigation was tested at 11 and 15 MaP. The statistical analyses consisted of several steps. First, the effects of species, treatment (F and I), and treatment interaction on total biomass production were tested. Second, differences observed between the species in biomass production with respect to the silvicultural treatments were analyzed. Species, fertilization, and irrigation were considered as fixed terms, while plot identity and fertilization (nested within plots) as random terms. The robust variance estimates and unstructured covariance options in STATA 14 (StataCorp 2015) were used to control for any deviation from the assumption of normality in the data. A least significance difference (LSD) test was used to compare means when significant treatment effects were found.

**Fig. 1** Drip system using local 1.5-L plastic bottles for the supplemental irrigation of saplings during the dry season (December 2014–May 2015) at the afforestation site in Northern Benin



## 2.6 Relative biomass allocation analysis

The effect of the treatments and ontogeny on biomass allocation was evaluated using clasmometric and allometric analyses (Poorter and Sack 2012). For the clasmometric analysis, the root mass fraction (RMF) was computed and plotted against the natural log-transformed estimate of the whole-plant dry biomass, which was used as a reference for the ontogenetic stage (e.g., Evans 1972; Poorter and Pothmann 1992; Poorter and Sack 2012). The allometric analysis was performed using an approach that allows for the differentiation between the effects of ontogeny and resource availability on biomass allocation (Hunt 1978; Coleman et al. 2004):

$$\ln Y = a + k \ln X \quad (1)$$

where  $Y$  is the total root (coarse and fine) dry mass,  $X$  is the total shoot (leaf and stem) dry mass,  $a$  is the y intercept, and the slope  $k$  is the allometric coefficient. The greater the value of  $k$ , the greater the biomass allocation to the roots relative to the shoots. If  $k$  is similar between the treatments, any shift in biomass allocation can be attributed to development (Hunt 1978). An analysis of covariance (ANCOVA) was used to reveal differences in  $k$  between treatments. The ANCOVA consisted of fitting the model in Eq. 1 to the treatment class variable. A significant interaction between the covariates (natural log-transformed stand-level estimates of root and shoot dry mass) and the treatment variable is considered indicative of a treatment-induced shift in biomass allocation. An ANCOVA was also used to compare the RMF versus tree size relationship between the treatments as revealed by clasmometry. In the case of significant interactions between the covariate and treatment factor, an LSD post hoc test was used to compare the slopes. In the case of equality of slopes between the treatments, a test of differences in intercepts was additionally performed. Following Niklas and Enquist (2002) and Poorter et al. (2015), the reproductive biomass fractions (fruits of *J. curcas* and pods of *M. oleifera*), which represented 2–5% of the total biomass, were omitted from the analyses.

To further assess the relative explanatory power of ontogeny and treatment effects on biomass allocation, stepwise multiple regression was conducted to characterize the effect of total dry mass, fertilization, and irrigation on root dry mass separately for each species. The natural log-transformed mass values and their residuals were judged to be normally distributed and homoscedastic based on a combination of histograms and normality tests performed in STATA 14. Data from the three harvests at the experimental site were pooled to account for changes in biomass allocation over the full size range (Poorter et al. 2012; Poorter and Sack 2012).

## 3 Results

### 3.1 Biomass production

#### 3.1.1 Effects of resource supply

The shoot, root, and total biomass were affected by fertilization and irrigation, but the effects differed between species and assessment periods (Table 1). At 5 MaP, total biomass increased due to fertilization (F) compared to the control (C) by 18% in *M. oleifera*, 23% in *J. curcas*, and 25% in *L. leucocephala*. No significant difference in biomass accumulation was observed between the C and F treatments for *A. occidentale* and *P. biglobosa* saplings, which were the smallest in size (Figs. 2a and 3a, Supplementary material, Appendix S1). The fertilizer-related biomass increase in fast-growing species was mirrored predominantly in AGB accumulation (Fig. 2, Supplementary material, Appendix S1). At the end of the dry season (11 MaP), only the slow-growing species responded to treatments (Table 1, Supplementary material, Appendix S1). Both *A. occidentale* and *P. biglobosa* saplings accumulated greater root biomass under IF and F compared to C at this point in time (Fig. 2b). The continued growth during the dry season substantially enhanced total biomass accumulation in *P. biglobosa*, which increased by 176% under IF and by 335% under F compared

**Table 1** Significance of species, fertilization, irrigation, and fertilization  $\times$  irrigation factors for aboveground and belowground biomass production after the first growing season (0–5 MaP), the dry season (5–11 MaP), and the second growing season (11–15 MaP) for plantations of *M. oleifera*, *J. curcas*, *L. leucocephala*, *A. occidentale*, and *P. biglobosa* on degraded cropland in Northern Benin. *P* values lower than  $\alpha = 0.05$  are indicated in italics

Factors	Shoot <sup>a</sup>	Root <sup>b</sup>	Total <sup>c</sup>
5 MaP			
Species (S)	< 0.001	< 0.001	< 0.001
Fertilization (F)	< 0.001	0.002	< 0.001
S $\times$ F	0.002	0.021	0.005
11 MaP			
Species (S)	< 0.001	< 0.001	< 0.001
Fertilization (F)	0.549	0.915	0.677
S $\times$ F	0.579	< 0.001	0.035
Irrigation (I)	0.918	0.716	0.838
S $\times$ I	0.403	0.109	0.354
F $\times$ I	0.873	0.019	0.583
S $\times$ F $\times$ I	0.082	< 0.001	< 0.001
15 MaP			
Species (S)	< 0.001	< 0.001	< 0.001
Fertilization (F)	< 0.001	< 0.001	< 0.001
S $\times$ F	< 0.001	< 0.001	< 0.001
Irrigation (I)	0.183	0.827	0.281
S $\times$ I	< 0.001	< 0.001	< 0.001
F $\times$ I	0.253	0.002	0.114
S $\times$ F $\times$ I	0.832	0.104	0.711

<sup>a</sup> Shoot = leaves + stem; reproductive parts were not included

<sup>b</sup> Root = coarse roots + fine roots

<sup>c</sup> Total = shoot + root

to C. In contrast, the total biomass of the fast-growing species declined between 5 and 11 MaP mainly due to a loss of shoot biomass (Fig. 2b). At the end of the second growing season (15 MaP), total biomass followed the order F > IF > I > C for the fast growers, although the differences between treatments were not statistically significant in all cases (Fig. 2c). For the slow growers, the pattern of treatment effects on total biomass followed the order of IF = F = I > C. The range of the increase in total biomass under F compared to C was substantially greater (8–335%) in the slow growers compared to that in the fast-growing species (19–68%). A significant effect of second-order interactions (e.g., species  $\times$  fertilization  $\times$  irrigation) was observed on root and total biomass at 11 MaP only (Table 1) but resulted from differential responses of fast-growing and slow-growing species to treatments (Fig. 2).

Fertilization consistently affected every biomass component at 5 and 15 MaP. Conversely, the overall effect of irrigation was marginal. It was much more evident at 15 MaP and varied greatly between species (Table 1, Fig. 2).

### 3.1.2 Species differences

At 5 MaP, the species ranking in terms of total biomass accumulation followed the order *M. oleifera* > *J. curcas* > *L. leucocephala* > *A. occidentale* > *P. biglobosa* (Fig. 3a). Shoot biomass ranking followed the same order, but the root biomass was similar for *A. occidentale* and *P. biglobosa*. At 11 MaP, total biomass did not significantly differ between *M. oleifera* and *L. leucocephala* and between *J. curcas* and *A. occidentale* (Fig. 3b). At 15 MaP, *L. leucocephala* accumulated significantly greater total biomass than the other four species while *A. occidentale* and particularly *P. biglobosa* accumulated the lowest total biomass (Fig. 3c). For instance, total biomass in *L. leucocephala* was 53% greater than in *M. oleifera*, 183% greater than in *A. occidentale*, and 659% greater than in *P. biglobosa*. Shoot and root biomass followed a similar pattern (Fig. 3c). The total biomass on the 10-year-old fallow land was estimated at 19.98 Mg ha<sup>-1</sup>.

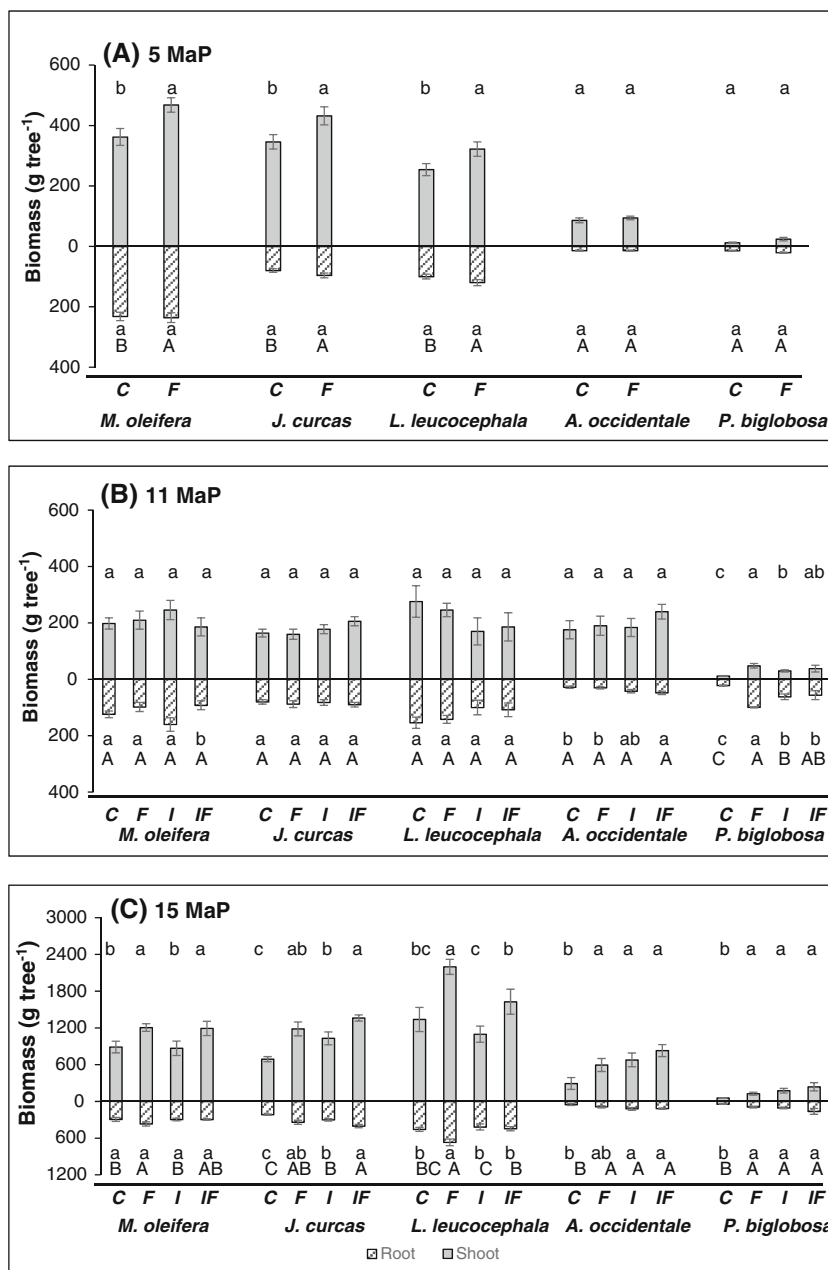
## 3.2 Biomass allocation

### 3.2.1 Effects of resource supply

For all species and treatments, the RMF fell (slope < 0, Fig. 4) with increasing tree size, reflecting ontogenetic drift, except in *A. occidentale*, which demonstrated no significant change ( $P = 0.68$ , Table 2, Fig. 4f). The decrease in RMF was influenced by the F and I treatments (Supplementary material, Appendix S2), but this exhibited great variation and was species specific (Fig. 4). For *M. oleifera*, the decline in the RMF was larger in saplings under C, I, and IF than in those under F, whereas a greater decrease occurred in *L. leucocephala* saplings under IF and I compared to C and F treatments (Table 2, Fig. 4a, c). However, treatment had no effect on the RMF in *J. curcas*, *A. occidentale*, or *P. biglobosa* ( $P > 0.05$ ; Table 2).

Allometric analysis revealed significant and strong linear relationships ( $R^2 > 0.8$ ) between root and shoot biomass (Table 3). For *J. curcas*, *L. leucocephala*, and *A. occidentale*, the variations in biomass allocation to belowground between treatments were all explained by ontogeny, as evidenced by a  $k$  that did not significantly differ between treatments (Table 3). The rate of increase in root biomass with increasing shoot biomass (i.e.,  $k$ ) did not differ ( $P > 0.05$ ) between treatments for *P. biglobosa*, but the intercept was significantly higher for saplings grown under C and I compared to F and IF ( $P = 0.038$ ; data not shown). Conversely, *M. oleifera* exhibited a significantly higher  $k$  under both F and IF compared to C and I treatments (Table 3), indicating that the treatments altered belowground allocation in this species. Due to the differing patterns of treatment-induced effects on the allometric relationship, the relative variation explained by ontogeny and by fertilization and irrigation was further distinguished. The ontogeny was always important in the allometric relationship between roots and shoots, explaining 86–

**Fig. 2** Aboveground (leaves and stem) and belowground (coarse and fine roots) biomass of five woody species at **a** 5 MaP, **b** 11 MaP, and **c** 15 MaP on degraded cropland in Northern Benin. Treatments included control (C), fertilization (F), irrigation (I), and irrigation + fertilization (IF). The “zero” line on the y-axis delineates the ground surface. Vertical bars are standard errors of the mean. Lowercase letters above the bars denote shoots and those below the bars roots. Capital letters relate to total biomass. For each species, means with different letters indicate the significant impact of the treatment according to a LSD post hoc test ( $\alpha = 0.05$ )



95% of the variation in all species ( $P < 0.001$ , Table 4). The amount of variation in the root versus shoot allocation explained by fertilization was 0.3–0.7% and that by irrigation 1.4–43%. Neither factor appeared significant in the models (Table 4).

### 3.2.2 Species differences

Of the species selected for this study, the RMF was the highest in *P. biglobosa*, ranging between 30 and 75% of the total biomass over the entire size range (Fig. 4f). The proportion of total biomass allocated to belowground was < 50% in *M. oleifera*, *J. curcas*, and *L. leucocephala* and < 30% in *A. occidentale*. The latter had the highest  $k$  value of the five

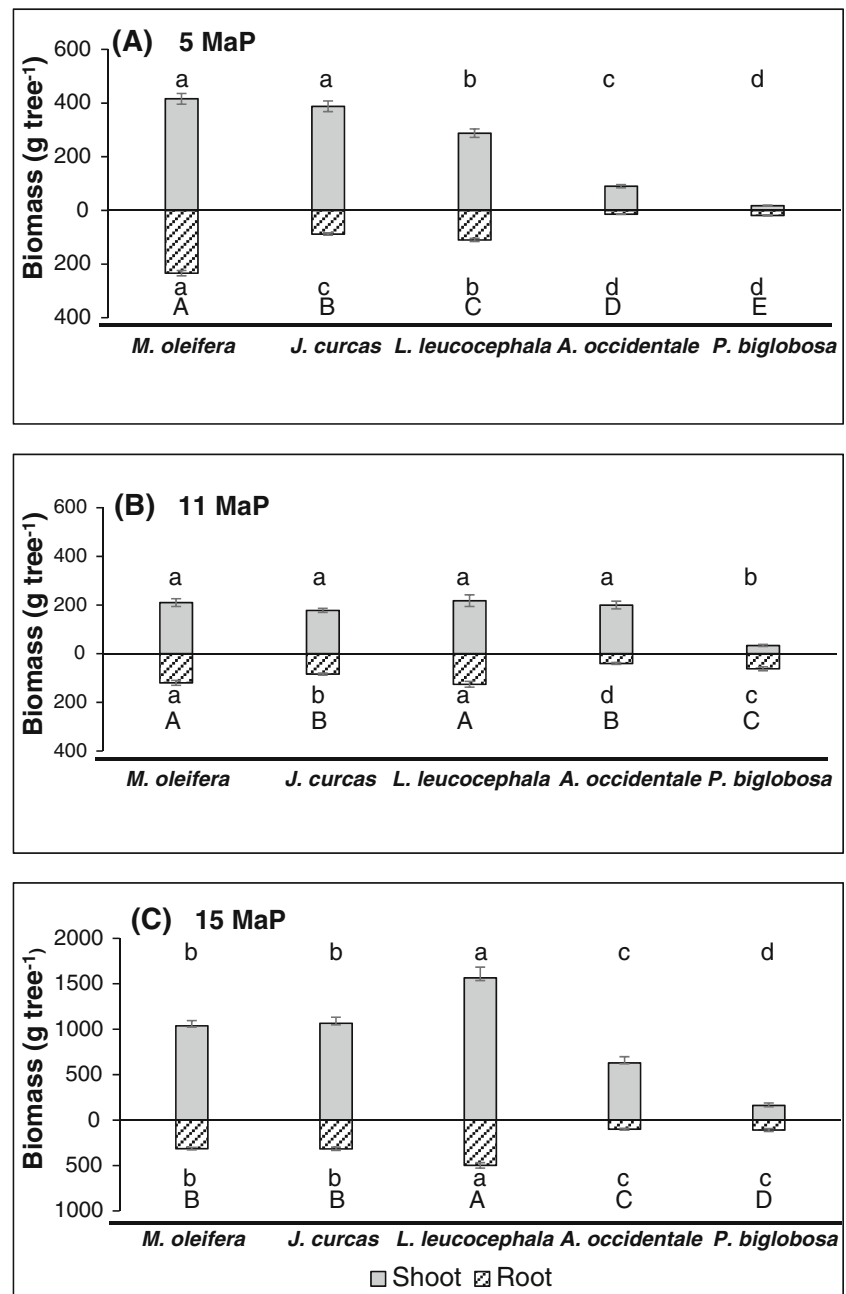
species, which was close to unity (a mean of 0.98 with a 95% confidence interval between 0.9 and 1.06, Table 3). The  $k$  values averaged  $0.66 \pm 0.09$ ,  $0.78 \pm 0.08$ ,  $0.78 \pm 0.07$ , and  $0.75 \pm 0.07$  for *M. oleifera*, *J. curcas*, *L. leucocephala*, and *P. biglobosa*, respectively (Table 3).

## 4 Discussion

### 4.1 Silvicultural treatment effects

The addition of manure and supplemental irrigation generally increased biomass production in the multi-species

**Fig. 3** Mean aboveground (leaves and stem) and belowground (coarse and fine roots) biomass across treatments for five woody species at **a** 5 MaP, **b** 11 MaP, and **c** 15 MaP on degraded cropland in Northern Benin. Treatments and symbols are as in Fig. 2

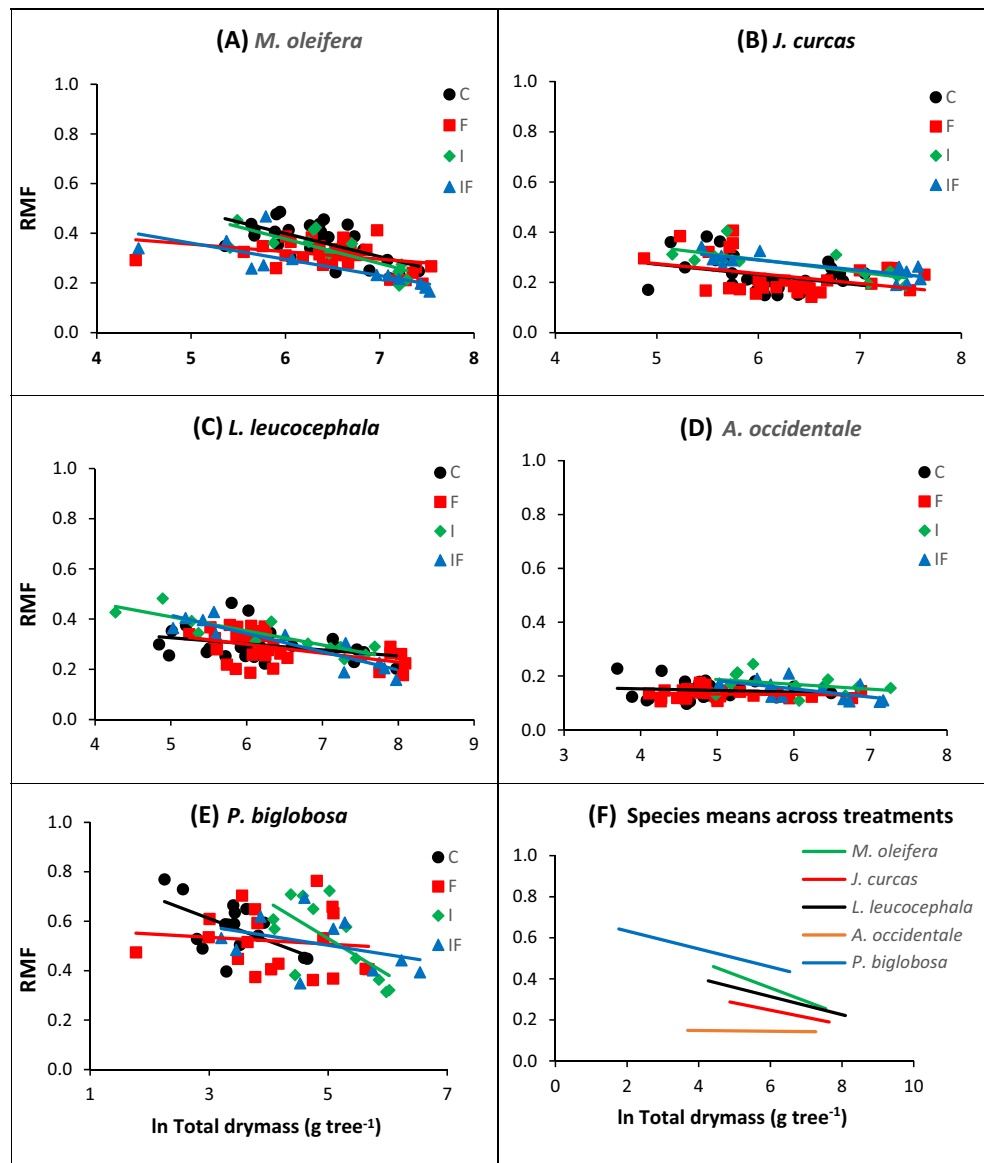


afforestation system. During the earliest stage of development (0–5 MaP) and in the second growing season (11–15 MaP), increased nutrient availability led to greater aboveground and total biomass accumulation but in *M. oleifera*, *J. curcas*, and *L. leucocephala* only. This confirms our hypothesis that increased resource availability has a more positive effect on the biomass production of fast-growing species than it does on slow-growing species. The higher sensitivity of the three fast-growing species to the experimental treatments is likely a result of their exploitative use of resources (Reich et al. 2003; Reich 2014; Noulèkoun et al. 2017). In contrast, both slow-growing species were responsive to silvicultural practices during the dry season (5–11 MaP), which is in line with

the hypothesis that increased resource availability also increases biomass production in slow growers. Fertilization and irrigation led to greater total biomass accumulation in *P. biglobosa* and to an increase in BGB in *A. occidentale* saplings receiving I and IF (Fig. 2b). Slow-growing species perform better when resources are scarce because of their conservative use of resources (Reich et al. 2003; Reich 2014; Noulèkoun et al. 2017), which explains the differences observed in the response to the treatments by the fast-growing and slow-growing species during the dry season.

Of the five species, *L. leucocephala* produced the greatest total biomass (2.1 kg tree<sup>-1</sup>) at 15 MaP on the degraded cropland, with the total biomass of manured *L. leucocephala*

**Fig. 4** Fractional allocation to roots (RMF) relative to total sapling dry mass for *M. oleifera* (a), *J. curcas* (b), *L. leucocephala* (c), *A. occidentale* (d), *P. biglobosa* (e), and all species combined (f) in response to the control (C), fertilization (F), irrigation (I), and fertilization + irrigation (IF). Total dry mass values were natural log-transformed. Data from three harvests (5, 11, and 15 MaP) are considered



saplings almost doubling that of *M. oleifera* and *J. curcas* (Figs. 2c and 3c). This demonstrates the strong potential of this species in afforestation schemes. The leguminous *L. leucocephala*, known as a “fertilizer tree,” has been

recommended for the restoration of impoverished lands (Garrity et al. 2010). In addition, we found that manuring boosted the biomass production of *L. leucocephala*, possibly due to the increased supply of P, which enhances N<sub>2</sub>

**Table 2** Slope comparison for the relationship between the fractional allocation to roots (RMF) and total dry mass for five woody species under the control (C), fertilization (F), irrigation (I), and fertilization + irrigation (IF) treatments

Treatment	<i>M. oleifera</i>	<i>J. curcas</i>	<i>L. leucocephala</i>	<i>A. occidentale</i>	<i>P. biglobosa</i>
Control (C)	-0.093b	-0.041a	-0.024b	-0.006a	-0.091a
Fertilization (F)	-0.030a	-0.039a	-0.036b	-0.002a	-0.014a
Irrigation (I)	-0.099b	-0.049a	-0.056ab	-0.018a	-0.146a
Fertilization + irrigation (IF)	-0.065ab	-0.041a	-0.073a	-0.031a	-0.038a
Mean across treatments	-0.066B	-0.035B	-0.044B	-0.002A	-0.044B

Within columns and for treatments, values followed by the same lowercase letter are not significantly different. Within a row, species means followed by the same uppercase letter are not significantly different. For all the cases of equal slopes between the treatments, the comparison of the intercepts showed no significant difference (data not shown). A significance level of  $P < 0.05$  was used for the LSD post hoc test



**Table 3** Allometric coefficients ( $k$ ) and coefficient of determination ( $R^2$ ) for the root and shoot relationship of five woody species under the control (C), fertilization (F), irrigation (I), and fertilization + irrigation (IF) treatments

Treatments	<i>M. oleifera</i>	<i>J. curcas</i>	<i>L. leucocephala</i>	<i>A. occidentale</i>	<i>P. biglobosa</i>
Control (C)	0.564b	0.721a	0.855a	0.965a	0.593a
Fertilization (F)	0.814a	0.755a	0.803a	0.982a	0.821a
Irrigation (I)	0.561b	0.747a	0.764a	0.866a	0.428a
Fertilization + irrigation (IF)	0.674ab	0.793a	0.667a	0.741a	0.799a
Mean across treatments	0.658C	0.785B	0.777B	0.981A	0.751AB
$R^2$	0.836	0.812	0.918	0.942	0.851

Within columns and for treatments, values followed by the same lowercase letter are not significantly different. Within a row, species means followed by the same uppercase letter are not significantly different. For *J. curcas*, *L. leucocephala*, and *A. occidentale*, the comparison of intercepts between the treatments showed no significant difference. Intercepts were significantly different between the treatments for *P. biglobosa* (data not shown). A significance level of  $P < 0.05$  was used for the LSD post hoc test

fixation in trees growing in P-poor agricultural soil (Djumaeva et al. 2013).

The limited response to irrigation during the dry season can be attributed to (i) a greater importance of nutrient over water stress under semi-arid conditions (Reubens et al. 2011; Noulèkoun et al. 2017), (ii) the relatively low amount of water supplied per plant during the dry season, and (iii) the deciduousness of the tree species as a drought-escaping mechanism inducing dormancy during dry periods (e.g., Monasterio and Sarmiento 1976; Reich and Borchert 1984; Poorter and Markesteijn 2008). In the fast-growing species, the deciduousness also caused a decrease in biomass accumulation between 5 and 11 MaP. However, the effect of supplemental irrigation may extend beyond the dry season period in which it was applied, as shown by the enhanced longitudinal and cambial growth of shoots and roots in same trial at 15 MaP (Noulèkoun et al. 2017), the extended photosynthetically active period in coppicing trees in South Africa (Moyo et al. 2015), and the overall rise in AGB production in a long-term afforestation trial in arid Uzbekistan (Khamzina et al. 2008). This post-irrigation

effect was particularly evident in the slow growers, which exhibited comparable total biomass production under the I, F, and IF treatments at 15 MaP (Fig. 2c).

#### 4.2 Biomass allocation as influenced by ontogeny and resource availability

Previous studies aiming to understand the effects of increased resource availability on biomass allocation in deciduous species (Achten et al. 2010) and conifers (Coyle and Coleman 2005) have reported conflicting results, presumably because the impact of ontogeny was not always considered (McCarthy and Enquist 2007). In our study, except for *A. occidentale*, the RMF dropped for all species and across all treatments over time (Fig. 3 and  $k < 1$ , Table 3), confirming that the partitioning of biomass to belowground was predominantly controlled by ontogeny (Table 4). Similar ontogenetically induced falls in the RMF have been reported for seedlings (Coleman et al. 1998) and 3–4-year-old plantations of loblolly

**Table 4** Multiple regression parameters according to tree species demonstrating the relative importance of ontogeny (total dry mass), fertilization, and irrigation for belowground biomass allocation. The dependent variable is root dry mass. A and B represent the intercept and slope of the robust linear regression

Species	Independent variable	A	B	Prob > F	Model $R^2$
<i>M. oleifera</i>	Total biomass	0.299	0.775	<0.001	0.861
	Fertilization	5.335	-0.089	0.458	0.007
	Irrigation	5.331	-0.143	0.336	0.014
<i>J. curcas</i>	Total biomass	-0.684	0.874	<0.001	0.855
	Irrigation	4.645	0.488	0.006	0.109
<i>L. leucocephala</i>	Total biomass	-0.278	0.849	<0.001	0.945
	Irrigation	5.027	0.242	0.242	0.019
<i>A. occidentale</i>	Total biomass	-1.903	0.991	<0.001	0.944
	Fertilization	3.347	0.097	0.656	0.003
	Irrigation	2.973	1.251	<0.001	0.427
<i>P. biglobosa</i>	Total biomass	-0.265	0.909	<0.001	0.951

Backward-stepwise selection was used to identify the most important independent variables. Multiple regression was then run to determine the relative variation explained by each. Data presented here are for the most important variables only

pine and cottonwood genotypes (Coyle and Coleman 2005; Coyle et al. 2008) in humid sub-tropical climates.

After accounting for ontogeny (by comparing  $k$ ), our results revealed either that there was no impact by I and F on biomass partitioning (e.g., in *J. curcas* and *P. biglobosa*) or that there was a shift toward greater root biomass (i.e., greater  $k$  values) in nutrient-rich conditions (e.g., in *M. oleifera*). This indicates that resource availability has no consistent effect on BGB allocation in fast-growing and slow-growing species during early growth stages. The lack of consistent shifts in biomass allocation to belowground due to nutrient and water amendments, also evidenced by variations in the decrease in the RMF between treatments for *M. oleifera* and *L. leucocephala* (Table 2), has been reported earlier for grasses (McConnaughay and Coleman 1999; Reich 2002) and 3-year-old woody plants (Coyle and Coleman 2005). The evidence combined thus suggests that patterns in optimal biomass partitioning in relation to resource variation after controlling for size are variable and—to a large extent—unpredictable. According to OPT, a greater proportion of roots, the organs responsible for resource acquisition, is expected under low-nutrient or low-moisture conditions (McConnaughay and Coleman 1999). This pattern was observed in *M. oleifera* (Fig. 4a), but the opposite occurred in N<sub>2</sub>-fixing *L. leucocephala* (Fig. 4c). Therefore, much of the variation in biomass partitioning explained by OPT is driven by differences in plant size and tree species rather than results from true plasticity in biomass allocation (McConnaughay and Coleman 1999; McCarthy and Enquist 2007).

The scaling slope ( $k$ ) of the allometric relationship between AGB and BGB was close to 1.0 for *A. occidentale* (Table 3). This finding is consistent with the allometric scaling theory (Enquist and Niklas 2002; Niklas and Enquist 2002), given that the relationship between root and shoot biomass is isometric over the considered size range (40–1400 g). In contrast, the scaling exponents for the other species were lower, ranging between 0.66 and 0.78. This discrepancy between observation and theory could be due to the variation in biomass partitioning between aboveground and belowground as a consequence of species-specific adaptations to water-limited and nutrient-limited conditions (Enquist and Niklas 2002). It is also possible that there were systematic errors associated with excavating the entire rooting systems, leading to an underestimation of the root biomass (Robinson 2004) and thus reducing the numerical values of the scaling exponents (Niklas 2004).

### 4.3 Implications for the afforestation of degraded croplands

Trade-offs between rapid growth, resource utilization, and site adaptability arise when selecting suitable species for forest plantations (Aubrey et al. 2012). Fast-growing species are considered competitive but are narrow site adapted and require

adequate resources to optimize productivity. Slow growers are more stress-tolerant and thus suitable to a broad range of environments (Aubrey et al. 2012). Of the five tested species, planting fast-growing *M. oleifera*, *J. curcas*, and *L. leucocephala* with an adequate nutrient supply has the highest production potential, as evidenced by their greater biomass production and strong positive response to fertilization (Fig. 2). This could increase return rates for investment in afforestation, but only for a limited expanse of plantation acreage due to the narrow site requirements of these species. In contrast, the slow-growing *A. occidentale* and *P. biglobosa* achieved reasonable biomass production and exhibited a more plastic response to treatments, as evidenced by the large range of treatment-induced increase in biomass. Considering the vast area of degraded croplands in SSA, afforestation using slow growers may produce a larger volume of biomass over the landscape as a whole. Overall then, a mixed cultivation of fast-growing and slow-growing species (Khamzina et al. 2006; Noulèkoun et al. 2017) seems to be most suitable for sustaining land cover.

The productivity of the 2-year-old afforestation site (5.8 Mg ha<sup>-1</sup> year<sup>-1</sup>) was about three times greater than that of the 10-year-old fallow (1.9 Mg ha<sup>-1</sup> year<sup>-1</sup>). This comparative advantage of plantations over natural succession is due to both the silvicultural management in plantations and the slow regeneration and unsustainable grazing of fallow land.

## 5 Conclusion

Belowground biomass represented a substantial share (up to 77%) of the total biomass of afforestation species in the early growth stages. However, with increasing tree size, a greater proportion of biomass was allocated to aboveground. Ontogeny rather than resource availability was the main driver of the observed shift in biomass partitioning. This implies that biomass allocation in young multi-species plantations can be captured by simple allometric coefficients, thereby improving predictions of growth models for young forest stands. On the other hand, increased resource availability through silvicultural treatment accelerated sapling biomass production, which overall led to the greater production of AGB. Fast-growing tree species produced more biomass, responding more vigorously to silvicultural treatments during the rainy seasons than did the slow growers. Significant treatment-induced increases in biomass and the ability to adjust production in resource-limited conditions were the key characteristics of slow-growing species. For these reasons, we suggest the use of both fast-growing and slow-growing tree species in the afforestation of degraded croplands.

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#### Compliance with ethical standards

**Data availability** This manuscript has data included as electronic supplementary material. The datasets generated and/or analyzed during the current study are available from the first or corresponding author on reasonable request.

**Conflict of interest** The authors declare that they have no conflict of interest.

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