



# Adsorption studies of methyl orange dye removal from aqueous solution using Anchote peel-based agricultural waste adsorbent

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## Abstract

A low-cost locally available agricultural waste-based Anchote peel adsorbent was studied to remove the methyl orange (MO) dye from wastewater which was created at laboratory conditions. The adsorbent was characterized by Fourier infrared and x-ray powder diffraction spectroscopic techniques. The adsorption capacity of the proposed adsorbent was investigated using the batch adsorption method. The best performance was achieved after careful optimization of experimental parameters such as adsorbent dose (0.4 g), initial dye concentration (70 mg/L), contact time (140 min), pH of the solution (8.0), and temperature (40 °C), respectively. The removal of MO from water using anchote peel adsorbent achieved a removal efficiency of 94.47% following the pseudo-second-order kinetic model, Freundlich isotherm, and chemisorption mechanism, respectively. The adsorbent showed heterogeneous surfaces and the adsorption of MO was thermodynamically spontaneous. Furthermore, the present results reveal that Anchote peel adsorbent is promising in future for the removal of organic dyes and other contaminants like toxic heavy metals from water and wastewater.

**Keywords** Adsorption · Agricultural waste · Anchote peel · Biosorbent · Methyl orange

## Introduction

Water is the most abundant and the scarcest substance required to sustain life in this World. The reason for this is only a very small portion of water (<3%) is accessible for human utilization, and also the available part is highly affected by natural and artificial pollution (Padowski et al. 2015; Wei et al. 2021). Water pollutants can be organic and inorganic pollutants, thermal, physical, and biological. Many industries such as textile, rubber, plastics, printing leather, cosmetics, paper, pharmaceutical, agricultural, food technology, hair coloring and photoelectrochemical cells widely use

dyes and discharge in the form of waste into water bodies (Tkaczyk et al. 2020). The contamination of water resources with dyes becomes a global concern because of their inertness and possible health effects. Among the different organic dyes anionic, cationic, and nonionic, azo dyes are a large class of synthetic dyes containing azo groups (–N=N–). The presence of aromatic rings and –N=N– groups in azo dyes structure make them highly toxic, carcinogenic, and teratogenic as well as harmful to the environment and biological organisms (Bai et al. 2020; Haque et al. 2021). From the large quantities of the azo dyes used in textile industries, up to 20% are eventually discharged to the environment as effluent thereby causing high toxicity and mutagenicity in aquatic life and humans (Mahmood et al. 2016).

Methyl orange (MO) is a common anionic azo dye that is harmful to the environment and biological systems, so it must be treated mildly before it can be discharged into the environment (Wu et al. 2021). MO is soluble in water, and is resistant to degradation; its removal from aqueous solutions by existing water treatment methods is difficult. Moreover, MO persists in the environment if not treated correctly and presents a hazard to living organisms. The leakage of dyes into wastewater has been identified as a factor affecting

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human health because of the carcinogenic and mutagenic effects of dyes (Alabbad 2020).

In this study, MO was selected as a target because of its wider applications in textile, printing, paper manufacturing, pharmaceutical, food industries, and chemical laboratories, as well as its negative effects on the environment, animals, and human beings (Alseddig et al. 2017). Thus, wastewater treatment was proposed using different methods to reduce health and environmental problems. There are various methods employed including photo-catalytic and electrochemical combined treatments, photo-catalytic degradation, sonochemical degradation, biodegradation, electrochemical degradation, adsorption process, chemical coagulation or flocculation, ozonation, cloud point extraction, oxidation, nano-filtration, chemical precipitation, ion-exchange, reverse osmosis and ultra-filtration (Bharathi and Ramesh 2013; Mittal and Mittal 2015). Contrary to other methods, adsorption is capable of removing the dyes from concentrated wastewater with low operating costs and high flexibility. Thus, adsorption is the most selective and effective technology for water pollution remediation to remove pollutants from complex matrices (Dai et al. 2018). The removal of different ionic dyes from aqueous media using adsorption techniques are reported by different scholars (Dai et al. 2018; Latif et al. 2019; Kadhom et al. 2020). In particular, the adsorption of dyes from water using agricultural wastes which are loose, porous in structure, and have amine, carbonyl, carboxyl, and reactive hydroxyl groups become attractive due to their dual advantages. These advantages are reducing the environmental burden and using it as a cost-effective and reusable adsorbent based on the principle of treating waste using waste materials (Huang 2017; Kadhom et al. 2020). In line with this, many research reports indicated that the removal of MO from wastewater through the application of an adsorption process using agricultural wastes attracted numerous attention across the World (Sigh et al. 2003; Mittal et al. 2007; Dakhil 2020). Subsequently, pomel peel (Zhang et al. 2020), eggshell (Belay and Hayelom 2014), corn leaf (Fadhil and Eisa 2019), chitosan (Conde et al. 2020), halloysite and chrysotile nanotubes (Wu et al. 2021) adsorbents were used to remove MO from water. These adsorbents were reported as having good performance for the adsorption of MO from water. Hence, the effort for exploiting other locally available and low-cost adsorbents is still demanding.

Anchote (*Coccinia abyssinica*) is an indigenous vegetable tuber crop cultivated for human consumption by Western and South-eastern Ethiopian communities of the Oromo people (Regasa et al. 2018). The skin or peel removed from Anchote tuber during cooking is discarded as waste without use for a long time. However, these agricultural waste biomaterials have different organic compounds with various functional groups to interact with other molecules. Therefore, in the

present study, Anchote peel waste biomaterials were applied as cost-effective, cheap, and effective adsorbents for the removal of MO from synthetic wastewater using the batch adsorption method. The adsorption efficiency, adsorption isotherm, adsorption kinetics, and adsorption thermodynamics of the Anchote peel adsorbent have been investigated for better performances for the removal of MO in the current study.

## Materials and methods

### Chemicals and apparatus

All chemicals and reagents used were analytical reagent grades. Analytical standard of methyl orange dye (MO) was obtained from (Loba Chemie and Qualigens, India). Analytical grade reagent hydrochloric acid (37%) (Merck, Germany) and sodium hydroxide (97%) (CDH, India) were used in the present investigation.

MO stock solution (1000 mg/L) was prepared with distilled water, and working solutions (10–100 mg/L) were prepared by diluting the stock solution before the batch adsorption experiment.

Ultraviolet–visible (UV–vis) spectrophotometer (Shimadzu DU-8800D, Japan) was used for measuring the concentration of MO dye before and after each batch of adsorption experiments.

Fourier Transform Infrared (FTIR) spectrophotometry (Shimadzu FTIR 8400S, Japan) and X-ray powder diffraction (XRD) (model Xpert MPD, Philips, Netherlands) spectrometry were used for the characterization of Anchote peel adsorbent materials. The chemical and crystal nature of the Anchote peel powder was performed using FTIR and XRD. FTIR spectra were recorded and used to study the available functional units and possible molecular interactions. XRD data were taken with  $\text{CuK}\alpha$  radiation ( $\lambda = 1.5418 \text{ \AA}$ ), on the powder diffractometer operated in the  $2\theta = 0.5\text{--}80^\circ$  range and step scan of  $\Delta 2\theta = 0.5^\circ$ .

### Biosorbent preparation and characterization

Anchote peel (*C. abyssinica*) was collected from the local fruit stalls in Nekemte City, Ethiopia washed thoroughly with distilled water to remove impurities, chopped into small pieces, sun-dried for four days, and finally oven-dried at  $105^\circ\text{C}$  until it became crispy. The dried Anchote peels were pulverized into the fine powder by a mechanical grinder and sieved through a 1 mm sieve. The obtained Anchote peel powder was used for experimental purposes. The fine powder of the adsorbent was kept in a plastic bag for the subsequent experiment and adsorbent characterization.

## Adsorption studies

### Adsorption of MO dye onto Anchote peel adsorbent

The batch adsorption experiments were used to determine the efficiency of the adsorbents for the removal of MO dye from an aqueous solution. The adsorption experiments were carried out by taking 25 mL of adsorbate solutions with varying initial concentrations (10–100 mg/L) in different conical flasks (100 mL). The fixed quantity of adsorbent was added to each flask and covered in an isothermal water bath shaker and shaken with a speed of 120 rpm at 298 K to achieve equilibrium condition. After filtration using Whatman filter paper grade 42, the residual concentration of MO was measured three times ( $n = 3$ ) by a UV–vis spectrophotometer at a pre-optimized wavelength of 462.10 nm. The experimental parameters affecting the batch adsorption of MO were optimized based on the design of the experiment using the Minitab method. The extent of dye adsorption can be expressed by three adsorption parameters, namely, adsorption capacity ( $Q_m$ ), adsorption capacity at any time ( $t$ ) ( $Q_t$ ), and removal efficiency (also known as percent removal, %R) given by the following Eq. (Douglas and Vermeulen 1981):

$$Q_m = \frac{V(C_o - C_e)}{W} \tag{1}$$

$$Q_t = \frac{V(C_o - C_e)}{W} \tag{2}$$

$$(\%) \text{ dye removal} = \frac{C_o - C_e}{C_o} \times 100 \tag{3}$$

where  $C_o$  and  $C_e$  are the initial and the equilibrium MO dye concentrations (mg/L),  $V$  is the volume of solution (mL) and  $W$  is the amount of adsorbent used (g).

### Adsorption isotherm study

The equilibrium relationship between the adsorbate dosage and the adsorbent uptake with time was investigated by using Freundlich and Langmuir isotherms.

Langmuir assumes that a monolayer is formed at the maximum adsorption that occurs on localized sites on the homogeneous surface to attain the Langmuir isotherm linear form given by the following Eq. (Cheung et al. 2001):

$$\frac{C_e}{q_e} = \frac{1}{K_1 q_m} + \frac{C_e}{q_m} \tag{4}$$

where  $C_e$  is the equilibrium dye concentration (mg/L),  $q_e$  (mg/g) is the amount of dye adsorbed at equilibrium,  $q_m$  (mg/g) is the amount of dye adsorbed at saturation and  $K_1$  (g/L) is Langmuir constant.

Freundlich isotherm uses Freundlich empirical to describe the distribution of a solute between solid and aqueous phases at equilibrium expressed as (Baup et al. 2000):

$$\ln q_e = \ln K_f + \frac{1}{n} \ln C_e \tag{5}$$

where  $K_f$  (g/L)<sup>1/n</sup> and  $n$  are Freundlich constants, which give a measure of both intensity and capacity of adsorption, respectively.

### Adsorption kinetic studies

Adsorption kinetics was used for determining the mechanism of dye adsorption onto the adsorbent material, the dye adsorption rate onto the adsorbent’s particle surface, and show the influence of different conditions on the speed of the adsorption process (Al-Harby et al. 2021). Therefore, pseudo-first-order kinetic model and pseudo-second-order kinetic model were studied to understand the kinetics of the adsorption of the MO dye onto Anchote peel adsorbent at 25 °C.

**Pseudo-first-order kinetic model** This model also known as Lagergren Model determines the relationship between the change in time and the adsorption capacity with an order of one as shown here (Lagergren 1898).

$$\text{Log}(Q_m - Q_t) = \text{log}Q_m - \frac{K_1}{2.303}t \tag{6}$$

where  $Q_e$  and  $Q_t$  are the adsorption capacity at equilibrium and time  $t$  (mg g<sup>-1</sup>), respectively,  $k_1$  is the pseudo-first-order rate constant (min<sup>-1</sup>) and  $t$  is the time (min). The values of  $Q_m$  and  $k_1$  can be determined from the intercept and the slope of the linear plot of  $\log(Q_m - Q_t)$  versus  $t$ .

**Pseudo-second-order kinetic model** The pseudo-second-order kinetic model is the model that shows the relationship between the adsorption capacity and concentration with second order. Pseudo-second-order Kinetic is expressed by (Ho and McKay 1998):

$$\frac{t}{Q_t} = \frac{1}{Q_m^2 K_2} + \left(\frac{t}{Q_m}\right)t \tag{7}$$

where  $K_2$  (gm/mg.min) is pseudo-second-order constant.

**Intraparticle diffusion model** Kinetic data obtained from the adsorption study were further analyzed using the intraparticle diffusion model (Weber and Morris 1963). This model

is used to determine the rate-controlling for the adsorption process and is expressed by:

$$q_t = (k_{\text{int}} t^{1/2}) + C \quad (8)$$

where  $q_t$  is the number of dye ions adsorbed at time  $t$  in mg/g,  $k_{\text{int}}$  is the intraparticle diffusion constant ( $\text{mg g}^{-1} \text{min}^{-1/2}$ ), and  $C$  is a constant ( $\text{mg g}^{-1}$ ) that is directly proportional to the boundary layer thickness. Both values of  $k_{\text{int}}$  and  $C$  can be obtained by calculations from the slope and intercept resulting from the linear curve of  $q_t$  versus  $t^{1/2}$ .

### Adsorption thermodynamic studies

The characteristic energy change associated with the adsorption of dyes on the surface of the adsorbent can be obtained from thermodynamic parameters. Thus, Gibbs free energy change,  $\Delta G^\circ$  ( $\text{kJ mol}^{-1}$ ), enthalpy change,  $\Delta H^\circ$  ( $\text{kJ mol}^{-1}$ ), and entropy change,  $\Delta S^\circ$  ( $\text{J mol}^{-1} \text{K}^{-1}$ ) were calculated for the adsorption of MO dye onto the Anchote peel adsorbent using four different temperatures (293, 303, 313, and 323 K) by using the following equations to identify the value at which optimum performance is observed.

From  $\Delta G = -RT \ln K_c$  and  $\Delta G = \Delta H^\circ - T\Delta S^\circ$ , then by rearrangement,

$$\ln K_c = \frac{\Delta S^\circ}{R} - \frac{\Delta H^\circ}{RT} \quad (9)$$

where  $R$  is the universal gas constant ( $8.314 \text{ J mol}^{-1} \text{K}^{-1}$ ),  $T$  is the absolute temperature (K) and  $K_c$  is the distribution

coefficient ( $QE/C_e$ ). By plotting  $\ln K_c$  versus  $1/T$ , the thermodynamic parameters can be acquired.

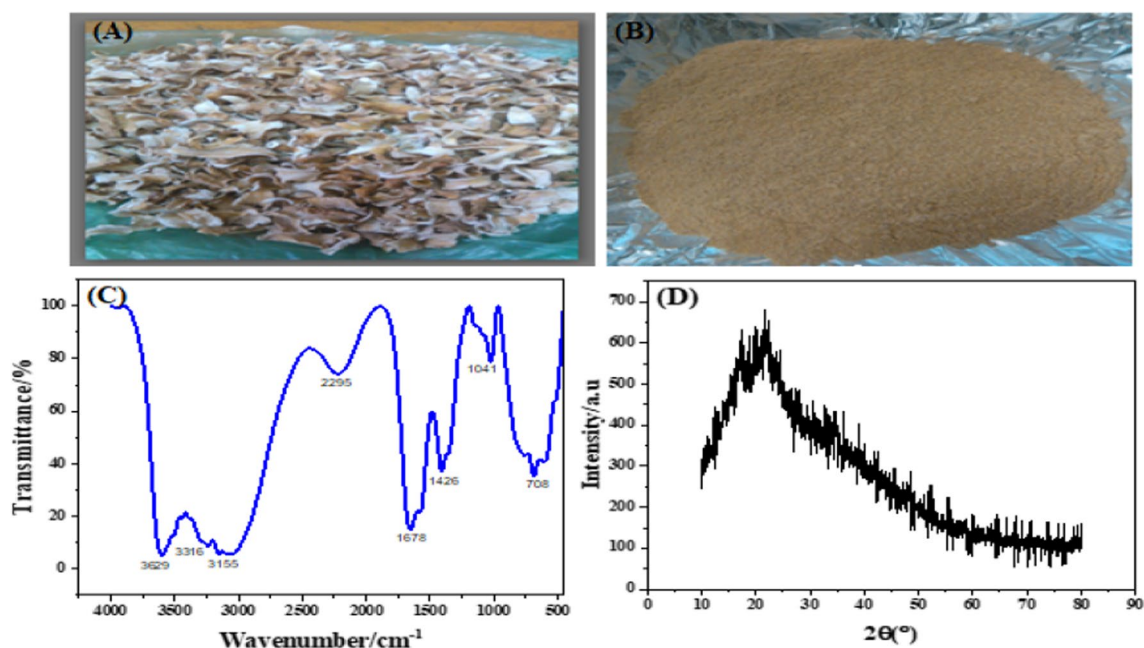
### Statistical analysis

In this study, all experiments were performed in triplicate, and the results are expressed as the mean  $\pm$  standard deviation. The computation of the isotherm, kinetic models, and thermodynamic results was performed using the Solver add-in (Microsoft Excel) to compute the parameters of the isotherm and kinetic models (Tran et al. 2017).

## Results and discussion

### Characterization of Anchote peel adsorbent

The characterization of adsorbent materials is very important to better understand the nature of the materials. Information obtained from the characteristic of adsorbents is necessary as their physical and chemical properties can influence the adsorption behavior. Thus, the characterization of Anchote peel was carried out by using FTIR and XRD (Fig. 1A and B). FTIR spectra were employed to ascertain the bond types, structures, and functional groups in the Anchote peel agricultural waste and determine the functional groups responsible for the adsorption of the dye (Fig. 1C). The bands observed between  $3800$  and  $2700 \text{ cm}^{-1}$  are due to  $-\text{C}-\text{H}$ ,  $=\text{C}-\text{H}$ , and hydroxyl groups ( $\text{O}-\text{H}$  groups) stretching (Samiey and Ashoori 2012). The presence of double



**Fig. 1** Dried Anchote peel **A**, Anchote peel powder **B**, and FTIR **C** and XRD **D** spectrums of Anchote peel powder



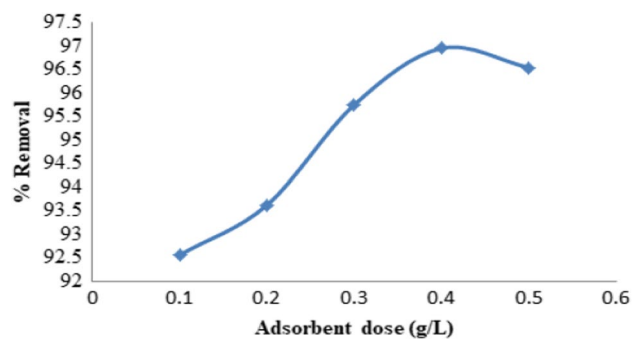
peaks with equivalent halves near  $3629\text{--}3315\text{ cm}^{-1}$  belongs to the primary amines ( $\text{NH}_2$ ) symmetric stretching that confirms the presence of amides. The absence of a single broad peak at about  $3400\text{--}3300\text{ cm}^{-1}$  and C–N stretching near  $1300\text{--}1000\text{ cm}^{-1}$  show the compound is amide (Benerjee and Chattopadhyaya 2017). The peak at  $2295\text{ cm}^{-1}$  is due to the  $\text{C}\equiv\text{N}$  stretching vibration while the two strong peaks at  $1678$  and  $1441\text{ cm}^{-1}$  are owing to the presence of C=O and N–H groups. The C=O from carboxylic acids, ketones, or esters can be associated with the band at  $1678\text{ cm}^{-1}$  and the presence of C=C that can be related to the lipid fraction of the Anchote peel was identified at  $1510\text{ cm}^{-1}$  (Pang et al. 2019). Furthermore, the peak at  $708\text{ cm}^{-1}$  corresponds to NH out-of-plane bending (Ahmad et al. 2009). Thus, the biomaterial mainly consists of long-chain amine compounds containing carbonyl functional units. The presence of both amine and carbonyl groups in the bionanomaterial makes it the ideal adsorbent for the adsorption of many compounds of interest.

The XRD spectrum of Anchote peel powder is depicted in Fig. 1D. It was perceived that the concentration values increased with increasing an angle up to a certain transition angle and the anomaly decreased further. The intensity of this adsorbent started from 10 degrees and increased up to 25 degrees, and then decreased gradually. A broad peak was observed at  $2\theta=20^\circ$ , corresponding to an interlayer spacing ( $d$ -spacing) of 1.75 nm indicating the noncrystalline nature of the adsorbent powder. From the broad spectrum, one can deduce that the Anchote peel material is highly amorphous and porous which is a characteristic of plant-based natural polymers (Lata 2017). Such kind of feature is important for adsorption since it increases the surface area to volume ratio of the material. The highest peak observed between 180 and 250 corresponds to the carbohydrate, particularly cellulose (Jiang et al. 2013; Ferrer et al. 2016) indicating the low crystallinity of the adsorbent.

## Adsorption studies

### Effect of adsorbent dose

Adsorbent dye removal efficiency in treating polluted wastewater mainly depends on the number of active adsorption sites available on the adsorbent. The amount of adsorbent used in adsorption is particularly important because it determines the adsorbate-adsorbent equilibrium in the system and predicts the treatment cost of adsorbent per unit of dye solution (Wanyonyi et al. 2014). The effect of the adsorbent dose for Anchote peel adsorbent was investigated using the adsorbent dose of different quantities in the range of  $0.1\text{--}0.5\text{ g/L}$  (Fig. 2). As can be seen from the results, the removal efficiency increased rapidly from 92.6 to 97.8% and then decreased due to saturation. The increase in removal



**Fig. 2** Effect of adsorbent dosage on MO dye removal using Anchote peel adsorbent at  $25^\circ\text{C}$

efficiency from 0.1 to 0.4 g of adsorbent might be due to the increase in total surface area and the availability of more binding sites for adsorption.

Thereafter, no significant change has been observed, indicating that the adsorption saturation is reached at the adsorbent dose greater or equal to 0.4 g/L. Thus, this concentration was chosen as the optimum dose for further studies. The effect of increasing the initial MO dye concentration on its adsorption showed positive effects, and the dye adsorption followed the Freundlich isotherm model.

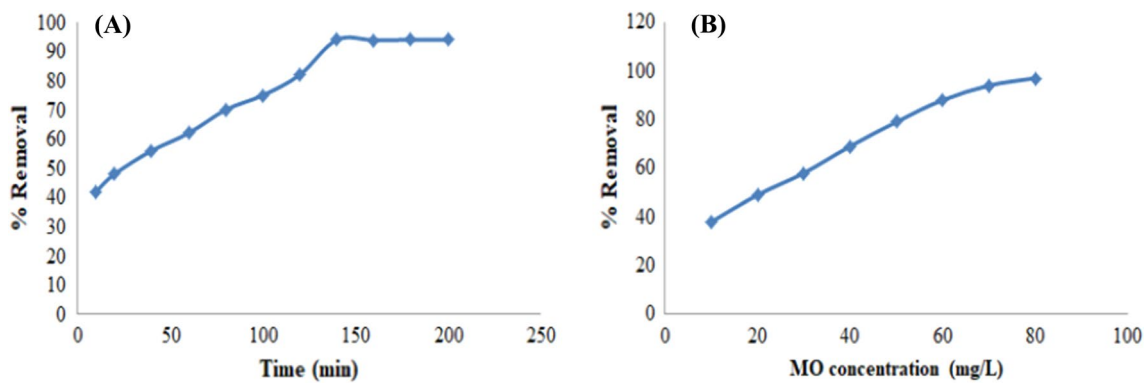
### Effect of contact time and initial MO concentration

The effects of contact time and initial dye concentration on the adsorption capacity were determined based on different times of agitation and different initial concentrations of MO as shown in Fig. 4. The obtained result shows the adsorption increases with contact time and finally reaches the equilibrium condition at 140 min (Fig. 3A). However, the increase is relatively higher during the initial 10 min.

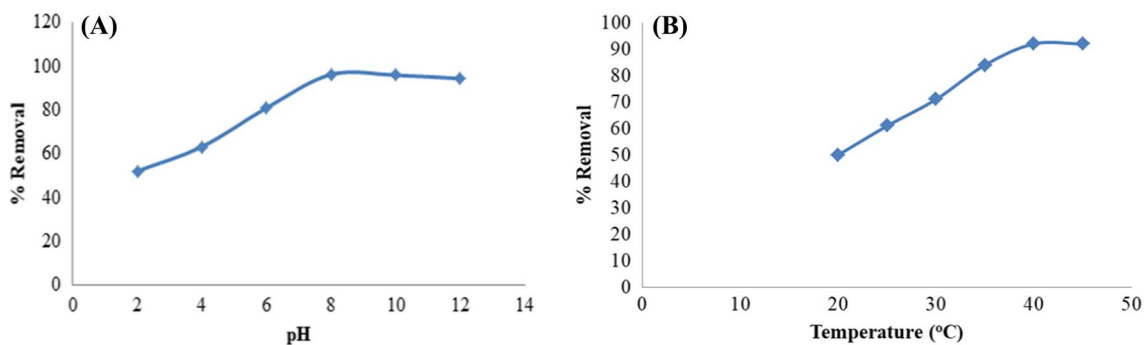
The rapid increase in adsorption during the initial stage may presumably be due to the availability of vacant active sites on the surface of the adsorbent. The slow increase at the later stage is due to the slow diffusion of dye into the adsorbent pores because the external sites are completely occupied. Similarly, the adsorption capacity increases with the increase in the initial dye concentration (Fig. 3B). It is also inferred that the adsorption capacity increases with increasing initial dye concentration until 70 mg/L and then continued with slight changes showing saturation of the adsorption.

### Effect of pH and temperature

The effect of pH on the removal efficiency of MO dye by Anchote peel adsorbent was evaluated as depicted in Fig. 4A. It has been observed that the removal efficiency increased from 50.0% to 96.23% on increasing the pH from



**Fig. 3** Effect of contact time and MO dye initial concentration on the adsorption processes



**Fig. 4** Effects of pH at 25 °C and temperature variation on the adsorption of MO onto Anchote peel adsorbent

2.0 to 8.0 for MO-Anchote peel adsorbent. Afterward, there is no major change in the removal efficiency with a further increase in pH values. MO is an anionic dye (carries the negative charge); in an acidic medium, the hydronium ions are predominant in the solution; therefore, the vacant sites of the adsorbent will be occupied by these cations, and the positive sites of the support will adsorb the MO molecules. The number of positively charged sites on the adsorbent decreased as the pH increased. Hence the adsorption of the dye molecules to the surface of the adsorbent increased as the pH value increased (Ahmad and Alrozi 2011; Akar et al. 2013). As the pH of the system increases, the number of negatively charged sites increases, decreasing the biomaterial's adsorption capacity (Kaur et al. 2013). At high pH, the negative charge of the adsorbent sites, therefore, does not promote the adsorption of the MO dye anions due to electrostatic repulsion. In addition, the low adsorption rate of MO at high pH is due to the presence of excess hydroxyl ions, which will compete with the anionic molecules of dye for the adsorption sites (Conde et al. 2020).

The temperature has a pronounced effect on the adsorption capacity of the adsorbents. The high temperature will destroy the covalent bond thus it reduces the holding capacity of the adsorbents to trap the dye molecules. The

adsorption capacity largely relies on the chemical interaction between the functional groups of the adsorbent surface and adsorbate (Jayarajan et al. 2011). The effect of temperature on the dye uptake of the adsorbent was studied by using different temperatures such as 15 °C, 20 °C, 25 °C, 30 °C, 35 °C, 40 °C, 45 °C, and 50 °C as indicated in Fig. 4B. The dye removal efficiency of 94.16% was achieved at the temperature of 40 °C and then continued with a slight increase with a further increase in temperature. The dye uptake increases with temperature because the adsorption of the MO dye increases with the increase in temperature the reason that the dye molecules can obtain sufficient energy to interact with the active site on the adsorbent.

### Isotherm study

Adsorption isotherms provide information on the capacity of the adsorbent and the equilibrium relationships between adsorbent and adsorbate at fixed temperatures (Oyelude et al. 2018). Consequently, to postulate the suitable isotherm, we evaluated the equilibrium relations of the batch adsorption of MO using Anchote peel as an adsorbent biomaterial.

The MO adsorption parameters were determined to investigate the adsorption isotherm model that best fits the

**Table 1** Isotherm parameters for the adsorption of MO onto Anchote peel adsorbent at 25 °C

Isotherm models	Parameters	Value
Langmuir	$Q_m$ (mg/g)	103.03
	$K_l$ (L/mg)	0.0282
	$R_l$	0.9720
	$R^2$	0.8683
Freundlich	$n$	1.2998
	$R_f$ (mg L <sup>-1/n</sup> L <sup>1/n/g</sup> )	8.0229
	$R^2$	0.9488

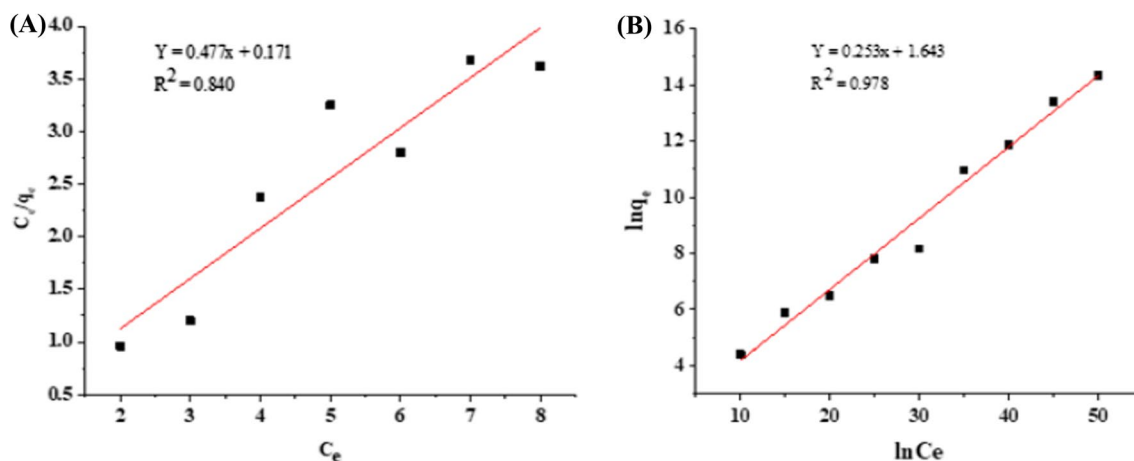
observed data of the adsorption process (Table 1). As shown in Fig. 5, the result confirmed that the Freundlich isotherm is the best reliable model with the studied process than the Langmuir isotherm model. Furthermore, the result reveals that the change in the behavior of the adsorption process occurred upon the addition of a specified MO amount to the heterogeneous surface of the proposed adsorbent which can be inferred from the higher  $R^2$  value for the Freundlich model. In the Freundlich isotherm, the factor  $1/n$ , was found to be greater than unity, suggesting that the adsorption of MO onto Anchote peel adsorbent is cooperative. Anchote peel adsorbent showed its surface heterogeneity and the exponential distribution of the active sites and their energies to interact with the MO for better performances (Al-Ghouti and Da'ana 2020). Thus, the presence of a heterogeneous surface with a multilayer adsorption mechanism enhanced the increase in adsorption with the increase in the MO concentration (Sewu et al. 2019).

**Kinetic studies**

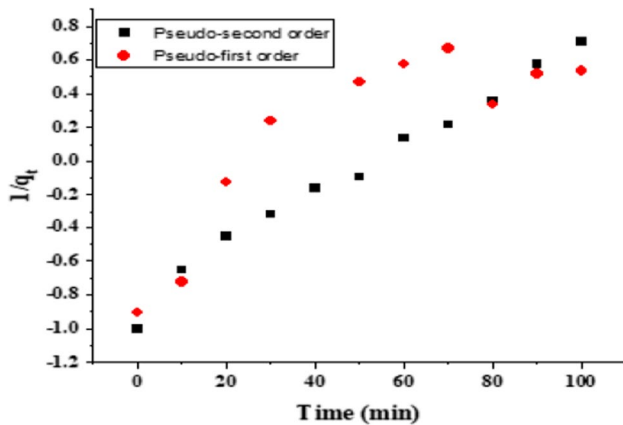
The adsorption kinetics of MO adsorption onto the Anchote peel adsorbent was studied using the pseudo-first-order and pseudo-second-order models. The kinetic parameters for both models were determined using Eqs. (6) and (7) as well as their respective plots (Table 2). The adsorption kinetics and the correlation coefficient ( $R^2$ ) determined to confirm that the adsorption data fit best with pseudo-second-order when compared to the pseudo-first-order kinetic model, as shown in Fig. 6 and Table 2. Furthermore, the  $R^2$  (0.996) indicated in Table 2 leads to infer that the pseudo-second-order kinetic model is the best one to represent the adsorption results of MO. This implies that the adsorbent has heterogeneous surfaces and thus the adsorption rate is limited by the chemisorption mechanism (chemical and valence forces by exchanging electrons between the adsorbent and the adsorbate) (Sterenzon et al. 2022). The result from this study is in agreement with the previous literature reports for the adsorption of MO on different adsorbents (Chen et al. 2011; Fadhil and Eisa 2019; Conde et al. 2020; Wu et al. 2021).

**Table 2** Kinetic parameters for the adsorption of MO onto Anchote peel adsorbent at room temperature (25 °C)

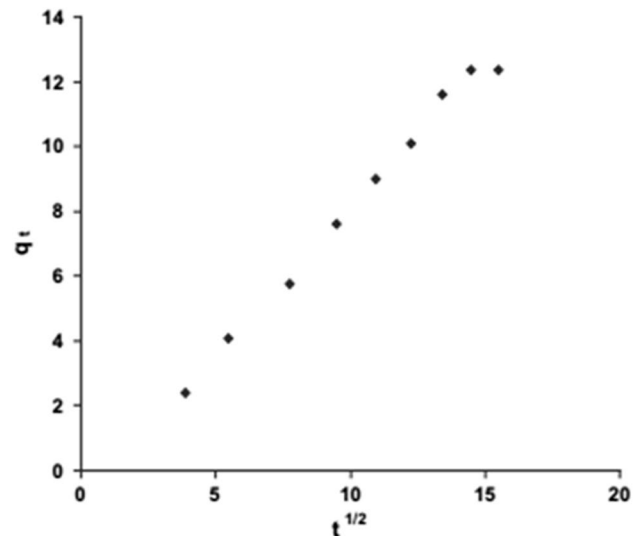
Kinetic models	Parameters		
	Reaction rate constant	$Q_e$	$R^2$
Pseudo-first-order	$K_1 = 0.0314$	0.1521	0.897
Pseudo-second-order	$K_2 = 1.3251$	0.4312	0.996



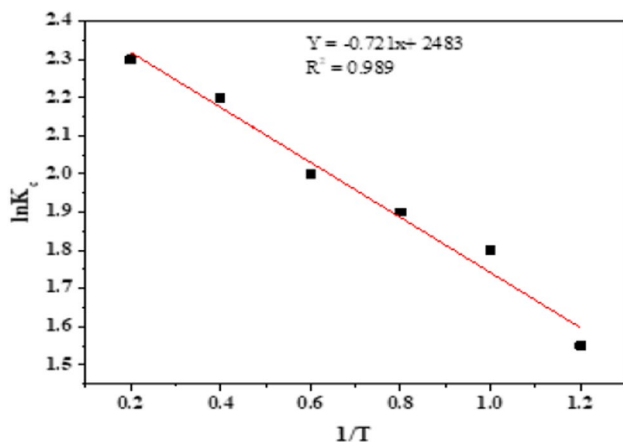
**Fig. 5** MO adsorption data fitted to Langmuir **A** and Freundlich **B** isotherms at 25 °C



**Fig. 6** Pseudo-first and pseudo-second-order plot for MO adsorption onto Anchote peel (70 mg/L, pH: 8, adsorbent dose: 0.4 g, contact time: 140 min, shaking speed 120 rpm, and temperature: 25 °C)



**Fig. 8** Intraparticle diffusion kinetic plot for removal of MO by Anchote peel powder



**Fig. 7** The thermodynamic plot of  $\ln K_c$  versus  $1/T$  for the adsorption of MO onto Anchote peel adsorbent

**Table 3** Determined thermodynamic parameters for the adsorption of MO onto Anchote peel adsorbent

Temperature (K)	$\Delta G^\circ$ (kJ/mol)	$\Delta H^\circ$ (kJ/mol)	$\Delta S^\circ$ (KJ/mol)
293	-9.05		
303	-8.53	-14.14	-27.9
313	-8.34		
323	-8.21		

### Thermodynamic studies

To understand the energy associated with the adsorption of MO onto the Anchote peel adsorbent, the thermodynamic plot is shown in Fig. 7, and the values obtained for  $\Delta G^\circ$ ,  $\Delta S^\circ$ , and  $\Delta H^\circ$  at 25 °C are -8.53 kJ/mol, -14.14 kJ/mol,

and -27.90 kJ/mol, respectively, as shown in Table 3. The increase in  $\Delta G^\circ$  from -9.05 to -8.21 kJ mol<sup>-1</sup> values with an increase in temperature indicated more efficient adsorption at higher temperatures due to the high mobility of more MO ions to the adsorbent surfaces (Adebayoa et al. 2019). The trend of free energy with temperature further indicates the spontaneous nature and feasibility of the adsorption process with an increase in temperature. Thus, the proposed adsorbent is thermodynamically suitable for removing MO from aqueous media with increased affinity at higher temperatures (Lawal et al. 2019).

The negative value of  $\Delta H^\circ$  (-14.14 kJ/mol) and  $\Delta S^\circ$  (-27.90 kJ<sup>1</sup> mol) confirmed the exothermic nature of the process and the increased predictability at the solid-solution interface during the adsorption process and a good affinity of the adsorbent for MO.

### Intraparticle diffusion model

This model is suggested to describe the adsorption of MO onto anchote peel better than pseudo-first-order and pseudo-second-order kinetics equations. It is an empirical model that describes the dye uptake mechanism and varies almost proportionally with  $t^{1/2}$  rather than with the contact time,  $t$ . When adsorbate in solution is mixed with the adsorbent, there occurs transport of the MO dye into the pores of particles from the solution through the interface between the solution and the adsorbent. The plot of  $q_t$  against  $t^{1/2}$  shown in Fig. 8 is linear passing through the origin for the known initial concentration of MO implying that the adsorption process followed the intraparticle diffusion model (Benerjee and Chattopadhyaya 2017). This implies that the system



is characterized by the high concentration of MO, vigorous mixing, and large particle size of Anchote peel powder. The plot shows that the quantity of MO adsorbed is proportional to the initial concentration of MO in an aqueous solution and the square root of time taken for adsorption (Weber and Morris 1963).

The mechanism of adsorption of MO on Anchote peel adsorbent is due to intraparticle diffusion. Briefly, the active functional groups in anchote biosorbent such as  $-OH$  and  $-C=O$  groups can form covalent or non-covalent (hydrogen bonding, ionic interaction,  $\pi-\pi$  stacking, etc.) while the  $-NH_2$  groups can interact with covalent (imine bonds), non-covalent interaction and semi-covalent (amide bonds) with the MO dye molecules (Fig. 9).

Various agricultural waste materials-based adsorbents have been investigated for the removal of MO dye from water. Therefore, the adsorptive capacity of the adsorbent investigated here was compared with those reported in the literature (Table 4). Anchote peel used in this study is a naturally available material with cost low adsorbent which exists in a natural environment as waste and can be used with small treatment and modification. The values of the adsorption capacity given in Table 4 show the proposed adsorbent is promising in removing the MO dye from an aqueous solution when compared with pomelo peel, eggshell, corn leaves, chitosan, halloysite, and chrysotile nanotubes.

### Conclusion

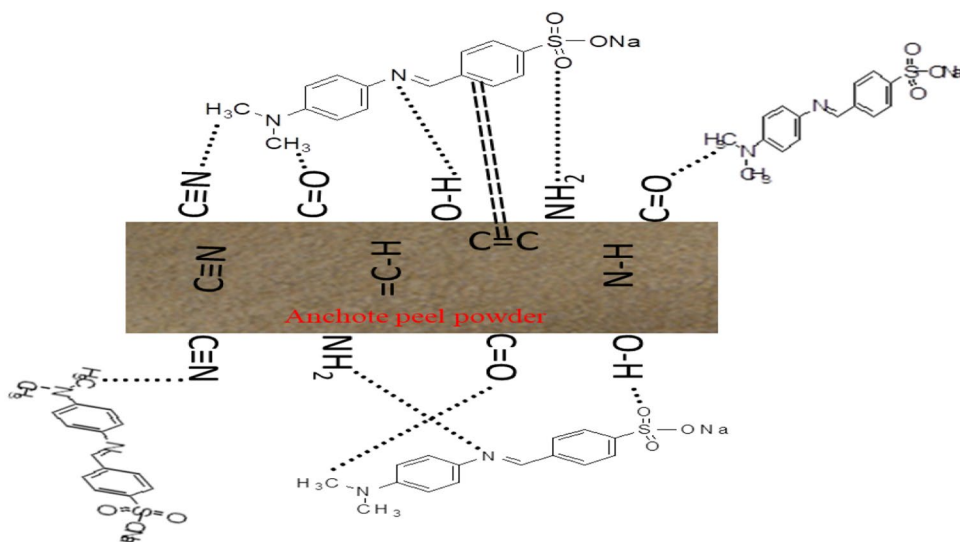
Sustainable wastewater treatment technology can be achieved by establishing innovative, cost-effective, and environmentally materials. Toxic organic dye removal from wastewater using adsorption technology and agricultural wastes is interesting in this regard. In the present

**Table 4** Comparison of adsorption capacity ( $Q_m$ ) of various adsorbents used to remove MO from water

Adsorbent	Source	$Q_m$ (mg g <sup>-1</sup> )	References
Pomelo peel	Plant	140.00	Zhang et al. (2020)
Eggshell	Animal	12.50	Belay and Hayelom (2014)
Corn leaf	Plant	13.85	Fadhil and Eisa (2019)
Chitosan	Animal	19.31	Conde et al. (2020)
Halloysite	Mineral	13.56	Wu et al. (2021)
Chrysotile	Mineral	31.46	Wu et al. (2021)
Anchote peel	Plant	303.03	Present work

work, Anchote peel adsorbents were found to be an effective low-cost adsorbent for the removal of MO dye from an aqueous solution. The maximum adsorption capacity was achieved at the optimized experimental conditions of adsorbent dose, initial dye concentration, contact time, pH, and temperature of the solution of 0.4 g, 70 mg L<sup>-1</sup>, 140 min, 8.0, and 40 °C, respectively. Anchote peel is indeed a viable and affordable adsorbent for the adsorptive removal of MO dye from water with an adsorption capacity of 103.03 mg/g, percentage removal of 94.47% with a relative standard deviation of 7.35% ( $n = 3$ ) according to the present result. The adsorption data best fits with the pseudo-second-order kinetic model, Freundlich isotherm, chemisorption mechanism, and thermodynamically spontaneous. The proposed biomaterials can be further applied for the water pollutants removal with a further physical and chemical treatment to enhance the performance of the adsorbent. Finally, the use of agricultural wastes such as Anchote peel as adsorbent is economical, practical, and more viable, sound in physicochemical nature, and thus its application to actual wastewater treatment will be the focus of our future work.

**Fig. 9** Schematic of adsorption mechanism of MO adsorption onto Anchote peel powder



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## Declarations

**Conflict of interest** The authors declare that they have no conflicts of interest to publish the work.

**Consent for publication** The authors declare that they have agreed to publish the work without any objection.

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