Applicability of gypsum in selective removal of anionic dye molecules from aqueous medium

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Abstract

This article describes how selectively anionic organic molecules could be removed from aqueous medium using naturally available gypsum (GS) adsorbent. Gypsum ($\text{CaSO}_4 \cdot 2\text{H}_2\text{O}$) shows strong interactions with anionic dye molecules while it does not show any affinity towards cationic dye molecules. We have shown the removal efficiency of gypsum taking chlorazole yellow (anionic dye) and methylene blue (cationic dye) as examples of adsorbates. The GS has been to be even more effective than activated carbon in attaining chlorazole yellow (CY) removal. Three well-known kinetic equations e.g. pseudo-first-order, pseudo-second order and intraparticle diffusion were exploited to interpret the experimental data. Results show best fitting with second order kinetic process with excellent regression coefficient ($r^2 = 0.99$) for the adsorption process. The equilibrium data were analyzed exploiting some adsorption isotherm models. It was apparent that the Freundlich isotherm model superbly fitted for CY dye adsorption process. And the maximum adsorption capacity, Q_m , was obtained as 12.85 mg g^{-1} at room temperature. The negative values of Gibb's free energy change (ΔG°) suggests that the CY dye molecule adsorption process is spontaneous in nature. Moreover, negative enthalpy change (ΔH°) indicates the exothermic nature of the adsorption process. The outcome could be exploited where anionic organic molecules are required to be separated, selectively.

Keywords Adsorption · Chlorazol yellow · Environmental · Kinetics · Equilibrium isotherm · Gypsum

Introduction

Nowadays, the increasing world population has created exponent of an alarming contamination of surface and ground water. Over the years, textile and dyeing industries are playing an important role on economic and industrial

revolutions around the world. Azo dyes, a class of colored organic pollutants, are used largely in textile, lather, paper and food industries and represent the main source of environmental contaminations. The dyeing process that is used in different industries produces a huge amount of wastewater, which are discharged directly into the aqueous mainstreams without any prior treatment. Thus, dye contamination of water causes a variety of environmental problems (Wang et al. 2005; Rao and Khan 2009; Magdy and Altaher 2018; Amin 2009; Malik 2004; Hameed 2009; Bhattacharya and Sharma 2005; Sivaraj et al. 2001; Kim et al. 2007; Tutem et al. 1998; Ho and McKay 1999; Panday et al. 1985; Brown 1987) as they are non-biodegradable and highly toxic to aquatic creatures and carcinogenic to humans.

Removal of dye molecules from wastewater is extremely difficult and is also a most thought-provoking matter in environmental network and industries. Over the years, several methods such as coagulation, sedimentation, filtration, oxidation, adsorption (Islam et al. 2019; Youcef et al. 2019; Acar et al. 2015), photo catalytic degradation (Gotoh et al. 2004; Miyah et al. 2018), biodegradation (Hasnat et al. 2005)

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and electrochemical degradation (Fan et al. 2008; Azmi et al. 1998), are reported for the removal of organic pollutants/dyes. Due to the complexity of the matrix, none of the methods can remove pollutants successfully (Hasnat et al. 2015). Among these processes, adsorption has been explored to be most popular to decimate organic contaminants from wastewater in aspects of basic expense, simplicity of design, ease of act and insensitivity of toxic substances (Rahman 2011; Hasnat et al. 2007; Uddin et al. 2009). Adsorption process is one of the popular techniques used in liquid-phase to remove organic pollutants from wastewater because in most cases it produces high-quality treated effluent. Adsorption is the adhesion of adsorbates (the molecules, atoms, ions or biomolecules being accumulated) on the surface of the adsorbent (solid, liquid or gas phase) and creates a thin film on the adsorbent surface. Activated carbon, also called activated charcoal, is of great utility in order to decimate dyes/pollutants from wastewater because of its excellent properties (Caner et al. 2009; Kumar et al. 2018; Kannan and Sundaram 2001; Giannakoudakis et al. 2016; Goswami and Phukan 2017; Andersson et al. 2011; Ahmaruzzaman and Gupta 2011; Klimmek et al. 2001; Hameed et al. 2017). But, due to its high cost, the widespread use of activated carbon is restricted. Therefore, inexpensive, highly effective, easily available and obvious economical adsorbents are still required. It is well known that Bangladesh's economy is highly dependent on the garments-based industries. Consequently, lots of industries have been installed in this country. Following these many non-conventional and inexpensive inorganic compounds were utilized as adsorbents in the world e.g. zeolite, bentonite and siliceous materials (Rahman 2011). This study aims to investigate the use of commercially available gypsum for treating synthetic wastewater. Gypsum is an inorganic substance which has a high affinity to metal ions that led scientist to investigate the potential use of this environmentally friendly compound to remove toxic metal ions from wastewater (Rauf et al. 2009; Gou and Lua 2003; Mall et al. 2007; Ayrancci and Duman 2005; Yan et al. 2014; Mittal et al. 2008; Deniz and Saygideger 2010; Yagub et al. 2014; Nissinen et al. 2013). Naturally, GS particles possess positive charges (Raii et al. 2012). Thus, GS particles should have strong interactions towards anionic dye molecules. Considering this point, we assumed that GS could be applied to remove anionic dye molecules efficiently. Hence, in this study, we have selected gypsum to remove Chlorazol Yellow, an anionic dye, from synthetic wastewater as an example. Chlorazol Yellow (CY) is a heterocyclic organic compound, with a molecular formula $C_{28}H_{19}N_5O_6Na_2S_4$, and its molecular weight is 695.72 g/mol.

For this purpose, at first, adsorbing capability of GS towards CY molecules was compared with that of well-known activated carbon. Next, detailed adsorbent related experiments were carried out under the impact of time of

contact, solution pH, amount of adsorbent, initial concentration of dye molecules and various anion effects. Different kinetic models e.g. pseudo-first order kinetic model, pseudo-second order kinetic model and intra-particle diffusion model, in addition to isotherm models given by Langmuir, Freundlich and Temkin were investigated. Finally, the adsorption was assessed by characterizing the adsorbent surface with Scanning Electron Microscopy (SEM).

Experimental

Materials and instruments

All the general chemicals were of analytical grade received from Merck, Germany and were used without further purifications. The main chemical, Chlorazol Yellow (CY) dye was supplied by Philip Harris Limited, Shenstone, England. The adsorbent, gypsum, was received from the responsible authority of Lafarge Surma Cement Ltd., Chatak, Sunamgonj, Sylhet (Bangladesh). It is to be noted that the sample was collected with permission letter (see supplementary file) given by the authority specifying appropriate permission for this study. The UV visible spectral changes were monitored using Avaspec, Spectrophotometer (Netherlands). The morphology of the adsorbed surface was analyzed employing a scanning electron microscope (SEM) (JSM-760F, Japan).

Batch adsorption experiment

The adsorption related experiments were monitored by means of UV–visible spectroscopy. A CY solution exhibits its absorption maxima $\lambda_{\max} = 390$ nm. To determine the concentration of CY molecules, absorbance was measured before and after adsorption at this absorption maxima. At first, the required amounts of CY and gypsum were taken in a conical flask such that the total volume was 25 ml. In this way a series of mixtures were prepared by varying their composition. The solutions as prepared were stirred mechanically using magnetic bars rotated at a speed of 200 rpm. The experiments were performed at room temperature (25 ± 1.0 °C). In order to monitor the concentration change, 2 ml solution was withdrawn periodically and after measurement the solution was returned to the reactor so that total volume of the reactor remained constant throughout the entire period of the experiments. When necessary, the pH of the reaction mixture was adjusted by adding dilute solutions of hydrochloric acid or sodium hydroxide.

The quantity of the dye removal (DR) in terms of percentage was evaluated by applying the following equation (Rahman et al. 2014; Akter et al. 2016):

$$\%DR = \frac{A_0 - A_t}{A_0} \times 100 \quad (1)$$

where A_0 and A_t represent the absorbance of CY at the beginning of the experiment and after a certain time period, respectively.

The amount of adsorption q_t (mg g^{-1}) at time t , was calculated by the following relation (Rahman et al. 2014; Akter et al. 2016):

$$q_t = \frac{\Delta C}{M} \times V \quad (2)$$

Here, ΔC (mg L^{-1}) refers to the change in CY concentration after a stipulated time period which was determined using Beer-Lambert law, V ($=0.025 \text{ L}$) is the volume of the solution, and ' M ' is the mass (g) of adsorbent gypsum.

The amount of CY removed at equilibrium due to adsorption q_e (mg.g^{-1}) was calculated by means of Eq. (3) (Rahman et al. 2014; Akter et al. 2016)

$$q_e = \frac{\Delta C_e}{M} \times V \quad (3)$$

ΔC_e (mg L^{-1}) being the change in CY concentration at equilibrium. Meanwhile, the standard error of the mean (S_a) was estimated using (Rahman et al. 2014)

$$S_a = \frac{SD}{\sqrt{N}} \quad (4)$$

SD and N being standard deviation and number of experiments repeated, respectively. The SD values of the observations were evaluated using the following statistical relationship (Rahman et al. 2014):

$$SD = \sqrt{\frac{1}{N} \sum_{i=1}^N (x_i - \bar{x})^2} \quad (5)$$

where, x_i ($x_1, x_2, x_3, \dots, x_N$) are the observed quantities of the sample's items, N is the number of observations, and $\bar{x}$ is the mean value of the observations.

The Chemical Oxygen Demand (COD) of the CY solution was determined before and after the adsorption process following the method prescribed by American Public Health Association (APHA) (Arami et al. 2006). For this purpose, the following equation was used for determination of %COD reduction (Akter et al. 2016)

$$\%COD \text{ reduction} = \frac{COD_{ini} - COD_{eqm}}{COD_{ini}} \times 100 \quad (6)$$

Results and discussion

Relative adsorption by activated carbon and gypsum

Figure 1a shows the spectral changes of CY dye after 20 min long adsorption experiment under identical experimental conditions. It is clearly seen from this figure that the concentration of CY molecules fallen more rapidly when gypsum was used as adsorbent instead of activated carbon (AC). Consequently, after 140 min long experiment (shown in Fig. 1b), gypsum attained $84 \pm 0.2\%$ removal of CY molecules from aqueous medium whereas at the same time, activated carbon removed only $58 \pm 0.2\%$ CY molecules. This observation suggests that gypsum materials have stronger interaction towards CY molecules than activated carbon. Thus, detailed adsorption related experiments were conducted using gypsum as adsorbent which are discussed in the following sections.

Contact time effect

Figure 2 shows the effect of contact time on CY ($[CY] = 70 \text{ mg L}^{-1}$) adsorption onto gypsum (7 g L^{-1}). It is observed that the CY molecules rapidly adsorbed on gypsum in the first 20 min and under the experimental condition about $71.6 \pm 0.2\%$ CY was removed due to a huge number of active sites available on the adsorbent surface. The extent of adsorption increased gradually until 100 min. contact

Fig. 1 Relative adsorption by gypsum (GS) and activated carbon (AC) in a 70 mg L^{-1} CY aqueous solution, adsorbent dose $= 7 \text{ g L}^{-1}$, rotational speed $= 200 \text{ rpm}$, temperature $= 25 \pm 1.0^\circ\text{C}$. Contact time $= 0$ to 20 min. (a), and 0 to 140 min. (b)

Fig. 2 Impact of contact time on CY dye removal efficiency by Gypsum in a 70 mg L^{-1} CY aqueous solution, Gypsum dose = 7 g L^{-1} , pH = 7.0, temperature = $25 \pm 1.0^\circ\text{C}$, rotational speed = 200 rpm

time. Thus, this time point was designated as the time of equilibrium in the present research where the quantity of CY removal was found to be $90.7 \pm 0.2\%$. However, above 100 min, the %DR was decreased. This observation may be interpreted by assuming that after saturation, the number of adsorption sites decreased which allowed less room for CY adsorption, which is reflected by the decrease of %DR removal.

Influence of gypsum dosage

Figure 3 shows the influence of gypsum dosage on the %DR. It was found that %DR increased with increase in gypsum dose up to 7 g L^{-1} at a fixed CY concentration of 70 mg L^{-1} . For instance, when the gypsum dose was 1 g L^{-1} the DR% was estimated to be $79.8 \pm 0.2\%$, which reached $91.7 \pm 0.2\%$ when the adsorbent dose was 7 g L^{-1} . It is worth mentioning that, as the gypsum dosage was increased active sites for CY adsorption also increased resulting in an increase of DR%. However, the %DR decreased as gypsum dose increased above 7 g L^{-1} . Note that at 10 g L^{-1} of gypsum dosage, the

Fig. 3 Influence of Gypsum dosage on CY dye removal efficiency by Gypsum in a 70 mg L^{-1} CY aqueous solution, pH = 7.0, contact time = 100 min, temperature = $25 \pm 1.0^\circ\text{C}$, rotational speed = 200 rpm

DR% was found to be $85.02 \pm 0.2\%$. Such a decrease of DR% suggests that above 7 g L^{-1} of adsorbent dosage, gypsum particles aggregated with CY and, consequently, the surface activity decreased which gradually hindered CY adsorption onto the gypsum adsorbent.

From COD experiment, we observed that the COD of CY dye decreases with time for the adsorption onto Gypsum. On the other hand, the %COD reduction increased as same as %DR increased by passing time. It was found that the initial value of COD for 70 mg L^{-1} CY was $412.2 \pm 0.2 \text{ mg L}^{-1}$ and after 100 min the lowest COD value was $67.8 \pm 0.2 \text{ mg L}^{-1}$. This phenomenon ascertains that the CY dye highly adsorbed onto Gypsum by passing time 100 min. Beyond that, the COD value of CY dye increased with time because the surface of adsorbent was blocked by CY molecules. Figure 4 shows the %COD reduction as time is increased. The similar phenomenon was investigated by other researchers (Arami et al. 2006).

Influence of pH

The adsorption of anionic CY dye was strongly affected by solution pH. Figure 5 represents the adsorption behavior of CY molecules onto Gypsum granules within the pH range of 4–9. It can be seen from the plot that the highest %DR was found to be 91.9 ± 0.2 at pH = 7. Below this pH value, the %DR was found to decrease. This may be due to the fact that at lower pH values, CY molecules are protonated as well as gypsum particles also, perhaps, gain positive charges. Thus, a coulombic repulsive force might predominate at lower pH values leaving no room for CY adsorption onto the gypsum surface. Meanwhile, above pH 7.0, CY molecules become deprotonated and gypsum particles would bear negative charges. Thus, the similar polarity again repelled the dye molecules refraining them from being adsorbed onto the gypsum surface. The outcome of pH observation

Fig. 4 Influence of time on %COD reduction for the adsorption of CY onto Gypsum in a 70 mg L^{-1} CY aqueous solution, Gypsum dose = 7 g L^{-1} , pH = 7.0, temperature = $25 \pm 1.0^\circ\text{C}$, rotational speed = 200 rpm

Fig. 5 Impact of solution pH on CY removal efficiency by Gypsum in a 70 mg.L⁻¹ CY aqueous solution, Gypsum dose=7 g L⁻¹, temperature=25 ± 1.0 °C, rotational speed=200 rpm. time of contact=100 min

is consistent with literature reported for different processes (Silva et al. 2004; Rahman et al. 2014).

Influence of CY concentration

Change of initial CY concentration also influenced the adsorption behavior. Figure 6 delineates how CY concentration influenced adsorption quantity onto gypsum surface. This trend confirms that the %DR decreased as the CY concentration was increased from 30 to 90 mg L⁻¹. It is seen that when CY concentration was 30 mg L⁻¹, the corresponding DR% was found to be 99.4 ± 0.2% for a dye concentration of 90 mg L⁻¹ the, DR% value decreased to 85.0 ± 0.2%. During this experiment, the adsorbent quantity was fixed, which means that the total adsorbing surface was also fixed. Consequently, as the CY concentration was increased, the surface to CY ratio decreased which resulted in a decrease of DR% with increased CY concentration.

Fig. 6 Influence of initial CY concentration on the CY dye molecule removal efficiency by Gypsum at pH=7, Gypsum dose=7 g L⁻¹, contact time=100 min, temperature=25 ± 1.0 °C, rotational speed=200 rpm

Influence of electrolytes

Wastewater may contain different salts apart from the organic pollutants. Thus, the influence of different electrolytes on CY dye adsorption onto Gypsum from water were investigated and the results are shown in Fig. 7. In the presence of different anions, mixtures of 70 mg L⁻¹ CY and 7 g L⁻¹ Gypsum were prepared separately. While, in the absence of these anions, the %DR was found to be 91.9 ± 0.2%, the %DR was found to decrease in the order F⁻ > CO₃²⁻ > SO₄²⁻ > NO₃⁻ > Cl⁻. The analysis unveiled that %DR was minimum in presence of fluoride ions. This observation concludes that F⁻ ions might have strong interactions with Gypsum surface which consequently decreased %DR. Meanwhile, Cl⁻ exhibited the least influence on CY removal because, perhaps, this anion had least interactions with Gypsum particles. Hence, in the presence of Cl⁻ ions, maximum %DR was observed. Similar phenomenon was investigated and noticed by other researchers (Abbas et al. 2012).

Adsorption kinetics

Adsorption kinetics unveils the efficiency of an adsorbent pertaining to a specific adsorption process. In the present case, the time-dependent CY adsorption process on gypsum was fitted with different kinetic models such as pseudo-first order kinetics, pseudo-second order kinetics, and intraparticle diffusion model (Ahmad et al. 2007; Sengil and Ozacar 2008; Akter et al. 2016).

Note that the pseudo-first-order kinetic model may be expressed as follows

$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (7)$$

Here, q_e and q_t define the amount (mg g⁻¹) of CY adsorption at equilibrium, and at time t (min), respectively, whereas k_1 (min⁻¹) refers the adsorption related rate constant relevant to

Fig. 7 Influence of anions on CY dye removal efficiency by Gypsum in a 70 mg L⁻¹ CY aqueous solution, Gypsum dose=7 g L⁻¹, pH=7.0, temperature=25 ± 1.0 °C, rotational speed=200 rpm. time of contact=100 min

pseudo-first-order adsorption. Figure 8a shows $-\ln [1-(q_t/q_e)]$ variation with time and the associated constants are reported in Table 1. It was found that the value of the regression coefficient (r^2) was equal to 0.89. Such a smaller value of r^2 value suggests that the pseudo-first order model is not well-fitted with respect to CY adsorption on to gypsum (Ahmad et al. 2007). However, in many cases the adsorption kinetics cannot be described fully by using the above equation. Thus, we fitted our data with respect to the equation of pseudo-second order kinetic model (Eq. (8)) at constant temperature (Miyah et al. 2018; Ahmad et al. 2007).

$$\frac{t}{q_t} = \frac{1}{q_e^2 K_2} + \frac{1}{q_e} t \quad (8)$$

where, the symbols represent their usual meanings as mentioned earlier and k_2 ($\text{g mg}^{-1} \text{min}^{-1}$) denotes the rate constant of a process that follows pseudo-second order adsorption kinetics. The fitting of the CY adsorption data with this model is shown in Fig. 8b. The data is tabulated in Table 1. In this case, the value of the regression coefficient (r^2) was found to be 0.99. Thus, it is obvious that CY adsorption followed pseudo-second order kinetics. This means that both CY and gypsum equally control the adsorption process.

On the other hand, Intra-particle diffusion model is often employed to evaluate the adsorption process across the porous matrix. Thus, in accordance with reference (Miyah et al. 2018), we also fitted our data with relevant Eq. (9).

$$q_t = k_i \sqrt{t} + f \quad (9)$$

Fig. 8 Kinetic Investigation-plots of pseudo-first-order kinetic (a), pseudo-second-order kinetic (b), and Intra-particle diffusion for adsorption (c) of CY dye adsorption onto Gypsum in a 70 mg L^{-1} CY aqueous solution, $\text{pH}=7.0$, gypsum dose = 7 g L^{-1} , time of contact = 100 min. , temperature = $25 \pm 1.0^\circ \text{C}$, rotational speed = 200 rpm

Table 1 Kinetic constants based on different kinetic models for the adsorption of CY dye onto Gypsum

Kinetic models	Rate constant	r^2
Pseudo-first-order	$K_1 = 3.28 \times 10^{-2} \text{ min}^{-1}$	0.89
Pseudo-second-order	$K_2 = 5.46 \times 10^{-2} \text{ g mg}^{-1} \text{ min}^{-1}$	0.99
Intra-particles diffusion	$K_i = 6.72 \times 10^{-1} \text{ mg g}^{-1} \text{ min}^{0.5}$	0.76

Gypsum dose = 7 g L^{-1} , $[\text{CY}] = 70 \text{ mg L}^{-1}$, $\text{pH} = 7.0$, time of contact = 100 min. , temperature = $25 \pm 1.0^\circ \text{C}$, rotational speed = 200 rpm

where, k_i is the rate constant related to intra-particle diffusion, and ' f ' indicates the thickness of boundary layer. Figure 8c represents a plot of q_t vs $t^{0.5}$ for CY adsorption onto Gypsum. It can be noted that the CY adsorption related data poorly fits with this model having a r^2 of 0.76. At the same time, the q_t vs $t^{0.5}$ plot did not pass through the origin, which entails some degree of boundary layer control and the intra particle CY diffusion did not control the rate limiting step. Hence it infers more than one step of CY adsorption process onto gypsum. It is noteworthy that instantaneous adsorption occurs in the first sharper portion (a), second stage is a slower gradual adsorption (b) then finally decreasing stage (c).

Adsorption isotherms

An adsorption isotherm comprises information concerning the capacity of an adsorbent, types of interactions, adsorption layers etc. To obtain such information, several

adsorption isotherm models are extensively studied depending on potential theory, thermodynamics and kinetics.

The Langmuir isotherm is often used to unveil the eventual monolayer formation by the adsorbate species onto an adsorbent under constant temperature condition (Ahmad et al. 2007). The linearized form of Langmuir isotherm can be expressed as per Eq. (10).

$$\frac{C_e}{q_e} = \frac{1}{K_L Q_m} + \frac{C_e}{Q_m} \quad (10)$$

Here, K_L ($L \cdot mg^{-1}$) denotes Langmuir adsorption constant, Q_m denotes optimum adsorption capacity in units of $mg \cdot g^{-1}$. Figure 9a shows the C_e/q_e vs C_e plot concerning adsorption of CY dye on gypsum adsorbent. The estimated values of the above-mentioned quantities are tabulated in Table 2. It was observed that the value of Q_m was evaluated as $12.85 \text{ mg } g^{-1}$. This value of Q_m indicates the maximum amount of CY that could be removed by one gram of gypsum adsorbent.

Meanwhile, the Freundlich adsorption isotherm provides information considering multilayer adsorption on a heterogeneous surface in terms of logarithmic dependency of surface energies with surface coverage. This isotherm is often expressed as follows:

$$\ln q_e = \ln K_f + \frac{1}{n} \ln C_e \quad (11)$$

Here, K_f ($L \cdot g^{-1}$) and ' n ' ($g \cdot L^{-1}$) designate the capability and intensity of the adsorption, respectively. The values of $\ln$

Table 2 Adsorption isotherms constants for different adsorption models for the adsorption of CY dye onto Gypsum

Isotherm models	Adsorption Isotherm Constants		r^2
Langmuir	$K_L = 0.11 \text{ L } mg^{-1}$	$Q_m = 12.85 \text{ mg } g^{-1}$	0.98
Freundlich	$K_F = 1.97 \text{ L } mg^{-1}$	$n = 1.99$	0.99
Temkin	$B_1 = 2.48$	$K_T = 1.67$	0.96

Gypsum dose = $7 \text{ g } L^{-1}$, [CY] = $70 \text{ mg } L^{-1}$, pH = 7.0, time of contact = 100 min., temperature = $25 \pm 1.0 \text{ } ^\circ\text{C}$, rotational speed = 200 rpm

q_e is plotted against the values of $\ln C_e$ and a linear graph is obtained (Fig. 9b). In Table 2, the constants evaluated from the graph are tabulated.

Finally, CY adsorption data were fitted with Temkin isotherm which provides information about adsorbent-adsorbate interaction on the basis of Heat of adsorption. It is considered that change in heat of adsorption of all molecules decreases linearly with the increase of adsorbate-adsorbent interactions (Mittal et al. 2008). The semi-logarithmic Temkin isotherm may be represented as follows:

$$q_e = B_1 \ln (K_T C_e) \quad (12)$$

In this equation, K_T denotes the equilibrium binding constant ($L \cdot mg^{-1}$) and B_1 corresponds to the heat of adsorption. The B_1 is defined as $B_1 = \frac{RT}{b}$ which contains the universal gas constant R ($J \cdot K^{-1} \cdot mol^{-1}$) and temperature T in kelvin. A plot of q_e vs $\ln C_e$ is shown in Fig. 9c and the constants

Fig. 9 Representation of adsorption isotherm models. Plot of C_e/q_e with respect to C_e : Langmuir model (a), plot of $\ln q_e$ against $\ln C_e$: Freundlich model (b), plot of q_e vs $\ln C_e$: Temkin model (c) of CY dye adsorption onto Gypsum in a $70 \text{ mg } L^{-1}$ CY aqueous solution, pH = 7.0, gypsum dose = $7 \text{ g } L^{-1}$, time of contact = 100 min., temperature = $25 \pm 1.0 \text{ } ^\circ\text{C}$, rotational speed = 200 rpm

are tabulated in Table 2. The Freundlich isotherm gave the best correlation with the experimental data ($r^2=0.99$). This observation indicates the heterogeneous nature of gypsum surface and the adsorption of CY proceeds by the formation of multilayer coverage at the outer surface of the gypsum.

Adsorption thermodynamics

The CY adsorption on gypsum has been identified as decreasing with the increase in temperature from 30 to 70 °C at pH 7.0 as shown in Fig. 10a. This observation is an indication of an involvement of exothermic process (Miyah et al. 2018). At the equilibrium, the Gibbs free energy change was estimated using Eq. (13) by evaluating equilibrium constant (K_L) as is discussed in previous section for Langmuir isotherm.

$$\Delta G^\circ = -RT \ln K_L \quad (13)$$

In thermodynamics, the Gibbs free energy change is often employed to correlate with the enthalpy change (ΔH°) and entropy change (ΔS°) as per the following equation which is known as Van't Hoff equation.

$$\ln K_L = \frac{\Delta S^\circ}{R} - \frac{\Delta H^\circ}{RT} \quad (14)$$

Plotting the values of $\ln K_L$ with respect to the reciprocal of temperatures, a linear graph is obtained (Fig. 10b). Then the thermodynamic quantities were evaluated from the slop and intercept and tabulated in Table 3.

Note that, irrespective of the temperature, ΔG° values were found to be negative indicating that the CY dye adsorption onto gypsum is a spontaneous process. Additionally, lower temperatures favored the adsorption process, as supported by the negative sign of ΔH° that was estimated as -6.4 J mol^{-1} which implied the exothermic nature of CY dye adsorption onto gypsum. At the same time, the obtained value of $-18.34 \text{ J mol}^{-1} \text{ K}^{-1}$ for entropy change shows that the adsorption caused a decreasing disorder. Similar

Table 3 Thermodynamic parameters for the adsorption of CY dye onto Gypsum

T/K	$\Delta G^\circ/\text{kJ mol}^{-1}$	$\Delta H^\circ/\text{J mol}^{-1}$	$\Delta S^\circ/\text{J mol}^{-1} \text{ K}^{-1}$
303	-0.8212	-6.4	-18.34
313	-0.5878		
323	-0.4404		
333	-0.2771		
343	-0.0188		

Gypsum dose = 7 g.L⁻¹, [CY] = 70 mg.L⁻¹, pH = 7.0, time of contact = 100 min., rotational speed = 200 rpm

phenomenon was reported by other researchers (Akter et al. 2016) (Figs. 11, 12).

Morphology

Figure 11 shows the SEM images of Gypsum before and after adsorption of Chlorazol Yellow dye on GS particles. It can be seen that the surface of the Gypsum particles after adsorption became covered with sponge-like material. The clear difference between the two images of Gypsum before and after CY adsorption confirms that CY molecules were firmly adsorbed onto the Gypsum surface.

Methylene blue adsorption onto gypsum

Adsorption of a cationic dye, methylene blue, onto gypsum was studied and found that gypsum had better affinity for methylene blue (MB) in acidic medium (pH = 6) rather than in neutral medium (pH = 7). The %DR was $24 \pm 2.0\%$ and $\leq 38 \pm 2.0\%$ in neutral and acidic medium, respectively (Figure 12a). Whereas the %DR of CY in the current study is found to be ~92% in neutral medium under identical experimental conditions (Fig. 5). Figure 12b is a presentation of time-dependent removal of methylene blue by gypsum which shows that %DR remains constant beyond 73 min after reaching a maximum %DR of 26.3%. On the contrary, the maximum %DR was more than 90% for CY removal study (in Fig. 2). It is therefore clear that gypsum particles

Fig. 10 Plot of %DR at different temperatures (a), and Van't Hoff's plot of $\ln K$ vs $1/T$ (b) of CY dye adsorption onto Gypsum in a 70 mg L⁻¹ CY aqueous solution, pH = 7.0, gypsum dose = 7 g.L⁻¹, time of contact = 100 min., rotational speed = 200 rpm

Fig. 11 SEM images of pure Gypsum (a) and CY adsorbed Gypsum (b)

Fig. 12 Effect of pH (a) and time (b) on the %DR of methylene blue by Gypsum in a 70 mg L^{-1} MB aqueous solution, pH = 7.0, gypsum dose = 7 g L^{-1} , time of contact = 90 min., temperature = $25 \pm 1.0^\circ \text{C}$, rotational speed = 200 rpm

Fig. 13 %DR of methylene blue and chlorazol yellow by gypsum in a 70 mg L^{-1} aqueous solution of either MB or CY, pH = 7, Gypsum dose = 7 g L^{-1} , rotational speed = 200 rpm, temperature = $25 \pm 1.0^\circ \text{C}$, time of contact = 90 min

have better affinity for cationic dyes, like CY, inferring a good selectivity.

For easy comparison, a bar diagram (Fig. 13) regarding MB and CY adsorption is given. It is noteworthy that both experiments were performed under identical experimental conditions.

Adsorption mechanism

As described earlier, we tried to carry out experiments pertaining to adsorption of a cationic dye (methylene blue) onto GS particles but very limited adsorption was noticed at neutral medium. This observation confirms that GS particles carry positive charges as is reported in reference (Rai et al. 2012). Thus, the positive charge on GS particles impeded the adsorption of cationic methylene blue molecules due to electrostatic repulsive forces. Conversely, it is seen from Fig. 3 that CY molecules are highly adsorbed at pH 7.0. At acidic pH, CY molecules become protonated, hence, poor attractive forces exist between the positively charged GS particles and the protonated CY molecules. Above pH 7.0, the amount of positive charges on GS particles perhaps declined, reducing the attractive forces between GS and CY particles. At pH 7.0, optimum positive and negative charges on GS and CY particles, respectively induced the strong attractions and by consequence effective adsorption occurs. This is schematically presented in Scheme 1.

Conclusion

Gypsum is a naturally abundant adsorbent and can be employed for adsorbing anionic dye molecules from aqueous medium selectively. In the present work, we have demonstrated how Chlorazol Yellow (CY), an anionic dye, could

Scheme 1 Adsorption of CY molecules on GS particles via electrostatic attractions between CY and GS particles.

be selectively removed from water. Moreover, we have found that GS particles have poor adsorption affinity for methylene blue (a cationic dye) in neutral medium. It was observed that CY molecules are effectively removed at neutral pH and that fluoride ions are highly competitive with CY adsorption. A gypsum dosage of 7 g/L is recommended for an efficient adsorption. The adsorption process follows pseudo-second order kinetics and could be best explained by Freundlich adsorption isotherm. The process is spontaneous and is less favorable at high temperatures. Based on kinetic, thermodynamic, and mechanistic investigations, it was suggested that gypsum can be effectively applied for the selective removal of anionic dyes (e.g. CY in the current study) in aqueous medium.

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Compliance with ethical standards

Conflict of interest The authors declare that they have no conflict of interest.

Ethical Approval All procedures followed were in accordance with the ethical standards.

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