

Do heavy-metal grassland species survive under a Scots pine canopy during early stages of secondary succession?

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Abstract The highly specialized flora of localities affected by former metal ore mining and metallurgy is endangered by succession or intentional afforestation all over Europe. Its last remnants therefore deserve our attention. We examined whether Scots pine encroaching on a heavy-metal grassland (Olkusz Zn-Pb mining area, S Poland) is outcompeting specialized herbaceous species, as has been observed elsewhere. Plant species composition and richness sampled at 124 plots were analysed in relation to pine stand parameters (canopy cover, stand age, stand basal area), abiotic environmental factors (e.g. soil properties) and spatial variables (e.g. plot coordinates). Plots were divided into three shading categories and compared in terms of vegetation and habitat parameters. Scots pine outcompeted several light-demanding species, leading to a decrease of total species richness and cover. Characteristic species of this grassland (*Biscutella laevigata*, *Silene vulgaris*) and some metal-tolerant plants were clearly insensitive to shading. For these early successional species, more important was the availability of microsites with shallow skeletal soil or bare subsoil. Tree stand parameters differently affected grassland vegetation: canopy cover

caused primarily a compositional shift in the community, while stand age was the principal agent of decline in species richness. Scots pine increased the soil concentrations of available Ca and Mg, and negatively affected soil development (organic matter and mineral particle accumulation), which might be beneficial to some shade-tolerant grassland species. Maintaining the studied grassland's present species richness and composition would require cutting woody plants less frequently than recommended for dry grasslands of non-metalliferous sites, and disturbing the soil surface.

Keywords Community structure · heavy-metal grassland · metalliferous soil · *Pinus sylvestris* · secondary succession · Variation partitioning

Introduction

Metalliferous sites have recently become high-priority conservation targets in Europe (Baker et al. 2010; Baumbach 2012). They are cultural heritage sites associated with earlier metal ore mining and metallurgy. They are also, and primarily, localities of a highly specialized flora formed by plant species, subspecies and ecotypes that under strong selection pressure, have developed morphological, anatomical or physiological adaptations enabling them to survive in extremely hostile habitats (Shaw 1990; Baker et al. 2010; Baumbach 2012). These plants, known as metallophytes, withstand the high content and consequently high toxicity of

Plant nomenclature Mirek et al. (2002)

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heavy metals, and also survive under nutrient and water scarcity.

For plant ecologists, some of the most interesting metalliferous sites are those left after zinc, lead and copper mining. In Europe, these are localities of heavy-metal grassland of the *Violetea calaminiariae* class (Ernst 1974). The most studied and most extensive patches of this kind of vegetation are located in a few countries: Belgium, Germany, the Netherlands and the United Kingdom (Baker et al. 2010). Similar plant communities occur in Poland as well; their range is limited to ore-bearing parts of the Silesian-Kraków Upland, especially the Olkusz Ore-Bearing Region (OOR). Endemic metallophytes and some characteristic species of heavy-metal grasslands (e.g. *Minuartia verna*) are not noted there, but other equally interesting metal-tolerant taxa can be found (Grodzińska and Szarek-Łukaszewska 2009).

In the last few decades, metalliferous habitats have been consistently eliminated in the OOR, as in other parts of Europe (Baker et al. 2010). Considered a burdensome legacy of the metals industry and presumed to pose an environmental hazard, they have been developed, converted to municipal landfills, or subjected to afforestation, mainly with Scots pine (*Pinus sylvestris*).

Efforts to preserve the heavy-metal grassland in the OOR have led to its designation as a valuable remnant of this type of vegetation; a fragment covering a non-reclaimed mine spoil heap is on the list of Poland's protected areas. The protection programme for the grassland has not provided for any specific form of ecological management to prevent natural succession. There is some evidence, including literature reports and archival photographs (Wóycicki 1913; Dobrzańska 1955), that the spoil heaps in the Olkusz environs were not colonized by woody plants for a number of decades. The reason for this was thought to be the impact of heavy metals, which can be more toxic to long-lived plants (shrubs and trees) than to short-lived ones (Dickinson et al. 1991). Observations in recent years have challenged this idea. The heavy metals in the soil have not been toxic enough to harm Scots pine, which has proved to be highly resistant to such pollution, probably thanks to ectomycorrhizal associations (Colpaert et al. 2011). As dust and SO₂ emissions from a nearby smelter were significantly reduced (Danek 2007), Scots pine began to rapidly colonize the majority of open areas, even heavily contaminated ones. Pine dispersal is now a relentless process which, without

human intervention, may cause the OOR heavy-metal flora to decline.

In this study, we sought to find out whether, and if so how, the Scots pine encroaching on a heavy-metal grassland is outcompeting its many herbaceous species, leading to an overall decrease in species richness, as has been shown for grasslands on non-metalliferous soils (e.g. Rejmének and Rosén 1992; Davies and Waite 1998; Willems and Bik 1998). Are there any grassland species able to persist under pine canopy? Are metal-tolerant taxa among them? Which pine stand parameters are most influential? Do pine trees exert an effect only by shading or also by altering soil conditions? To answer these questions, we recorded the species richness and composition of vascular plants of a heavy-metal grassland in a large number of plots varying in their level of shading by pine trees. In the plots, we also measured several environmental variables, including pine stand parameters, in order to capture all factors relevant to the distribution of herbaceous species. Since the plant community under study was expected to be spatially structured, we incorporated additional predictors: spatial variables derived from the spatial coordinates of the plots. Using variation partitioning, we assessed the relative contributions of three datasets (pine stand, abiotic environment, space) to the variation of species richness and composition.

Material and Methods

Study Site

The study site is a spoil heap of calcareous gangue more than a century old, remaining from Zn-Pb ore extraction and processing. It is located near the town of Bolesław in the OOR, S Poland (19°28'21" E, 50°17'31" N). According to Grodzińska and Szarek-Łukaszewska (2009), the shallow rendzina-like soil on the heap contains 0.32–0.65 % Pb and 3.3–7.8 % Zn. Since 1997 ca 5 ha of the heap has been under legal protection as an ecological area and a Natura 2000 site known as Pleszczotka górská. It is a stretch of species-rich heavy-metal grassland with abundant populations of rare plants, especially *Biscutella laevigata* and *Gentianella germanica*. Over the last three decades, this site has been colonized by Scots pine. The process accelerated in the late 1990s. In the year of the study,

there were still patches of well preserved grassland vegetation there, not shaded by trees.

Field Sampling

In 2008, 124 circular plots (1 m² each) were established on a grid at 8 m intervals in the study area (Fig. 1). The grid interval was chosen arbitrarily to provide a reasonable number of samples and to minimize autocorrelation resulting from the tree effect (a given tree could not shade more than one study plot). Two topographic variables were measured at each study plot: slope (with a SUUNTO PM-5/1520 clinometer) and aspect (with a compass).

During the vegetation season, the study plots were repeatedly surveyed to determine their vascular plant species composition. Plant species were identified in the field or sampled for identification in the laboratory. The abundance of particular species in the herb layer was estimated on a modified Braun-Blanquet scale (OTV scale) with nine degrees of cover (van der Maarel 2005). Supplementary data on the overall cover of lichens and bryophytes were also recorded.

After careful removal of the top soil organic layer (when present), four soil samples of the top mineral soil horizon were collected with a split corer (10 cm deep, 3.5 cm in diameter) from each study plot and bulked into one composite sample. Those samples were taken from spots very near but not inside the plots, so as not to disturb the plot vegetation. These spots were also used to determine the thickness of the A horizon (mean of four measurements).

Three parameters describing the impact of pine were recorded for the study area: canopy cover, stand basal area and age. Canopy cover was estimated directly for each study plot on a three-point scale: 0 – open (exposed to direct sunlight), 0.5 – partially shaded (less than half of a plot shaded by pine branches) and 1 – deeply shaded (more than half of a plot shaded by pine branches). The latter two parameters were measured for 141 squares demarcated on the same grid of plots (Fig. 1). This was done in 2010 as part of another study (unpubl.). Stand basal area was calculated from tree diameter measured at breast height (1.3 m). Trees less than 1.3 m tall were not included in that calculation. The age of the pine trees was estimated by counting whorls of horizontal branches and then averaged for a grid square.

Soil Analysis

The composite fresh soil samples were sieved through 2 mm mesh and air-dried at room temperature. The fine fraction (<2mm) was weighed and expressed as the percentage of soil sample total dry weight. Soil suspension pH was measured electrometrically using a pH electrode following extraction with H₂O (1:5 w:v). Total nitrogen was determined by the Kjeldahl method (ISO 1995). The soil was digested in H₂SO₄ with Kjeltabs (K₂SO₄+CuSO₄·5H₂O; Foss Tecator Digester Auto) followed by distillation in a Foss Tecator Kjeltec 2300 Analyser Unit. Available (Olsen) phosphorus was measured with an ion chromatograph (Dionex ICS-1100) following extraction of soil with



Fig. 1 a – Location of the study site, b – sampling grids applied in the study.

0.5 M NaHCO₃, pH 8.5 (Olsen and Sommers 1982). The content of available K, Ca, Mg and Zn was determined by flame atomic absorption spectrometry (AAS, Varian 220 FS) after extraction with deionized water (1:10 w:v, samples shaken for 1 h) for water-soluble or with 0.1 M BaCl₂ at pH 7 for exchangeable (ISO 1994). The availability of Pb (another important heavy metal in the OOR) was not analysed, since this element is largely immobilized in soils developed on calcareous mining spoils (Kapusta et al. 2011).

Data Handling

Based on the floristic data collected in the field, plant species richness (total number of herbaceous species) and cover were determined for each study plot. Cover was calculated by summing the percentage cover of individual species.

For each plot the slope-aspect index (SAI) was derived from topographic variables, calculated according to the formula of Nielsen et al. (2004): $SAI = \sin(\text{aspect} + 225) \times (\text{slope} / 46)$. The index represents the amount of direct solar radiation and warming of a plot in relation to a flat surface. The scale of this measure runs from -1 for the combination of highest slope (46° in this study) and northeast aspect to +1 for the combination of highest slope and southwest aspect. SAI was used for statistical analysis instead of the original topographic variables.

The centres of the squares in which pine trees were measured did not fit the coordinates of the study plots. Therefore, the values for pine stand basal area and age had to be interpolated. This was done by ordinary kriging (Isaaks and Srivastava 1990). The pine variables were detrended before kriging, and their spatial dependence was modelled with a spherical function (variogram). After kriging, the spatial trends were added back to interpolated surfaces, and new values of the variables were extracted at the nodes of the study plot grid. Spatial analysis was performed with Surfer 8 (Golden Software Inc., Golden, CO, USA).

Exploratory spatial analysis (spatial correlograms) showed that both the response variables and explanatory variables were more or less spatially structured. The presence of spatial autocorrelation in the data complicates statistical testing because the assumption of independence of error terms is violated (Legendre et al. 2002). One method of dealing with this problem is to incorporate additional variables representing a variety of

spatial patterns into the explanatory dataset. In this study, we used the raw coordinates of the plots (X, Y) and principal coordinates of neighbour matrices (PCNM), which are eigenvector-based spatial functions (Dray et al. 2006). The former represented linear trends in the data, and the latter modelled more complicated spatial structures – orthogonal (linearly independent) wavy structures covering all spatial scales encompassed by the sampling scheme (Borcard et al. 2011). A total of 88 PCNMs were generated (using R PCNM), 62 of which, characterized by positive spatial autocorrelation, were selected for subsequent analysis. Negative spatial autocorrelation was not taken into consideration. It is usually associated with small-scale structures produced by biotic processes such as species territoriality and competition (Dray et al. 2006), not to be observed in this study due to relatively large intervals between patches of vegetation surveyed.

Before the statistical analysis, response variables (plant species richness, cover) and explanatory variables (except for canopy cover and spatial variables) were transformed with a logarithmic or exponential function, according to the formulae given by Økland et al. (2001), to improve their normality or at least to remove skewness. Species abundances (expressed on the OTV scale) were Hellinger-transformed. This is recommended by Legendre and Gallagher (2001) to make the data, which normally contain many zero values (due to the presence of many rare species), suitable for Euclidean-based ordination methods such as redundancy analysis (RDA).

Statistical Analysis

In order to separate the effect of Scots pine encroachment on the heavy-metal grassland vegetation from the effects of other factors, including spatial dependence, we used the variation partitioning approach (Borcard et al. 1992). First, the explanatory variables were grouped into three sets: (1) pine stand parameters (P) comprising canopy cover, age and basal area; (2) abiotic environmental factors (E) comprising SAI and soil properties; and (3) spatial variables (S) comprising the X and Y coordinates and 62 PCNMs. These sets were separately subjected to forward selection (using R packfor) to identify variables that significantly explain the variation of plant species composition (in RDA) and species richness (in multiple regression). The forward selection procedure (with 999 random permutations) was based, as recommended by Blanchet et al. (2008), on the

double stopping criterion: selection stopped when a preselected significance level (here, $\alpha=0.05$) was reached or when adjusted R^2 (R_a^2) computed for the current model after adding another predictor exceeded global R_a^2 (i.e., for the model involving all explanatory variables).

The P, E and S sets of forward-selected variables were used to decompose the variation in species composition and richness into the following fractions: pure effects of P, E and S; their combined effects; and unexplained variation. This was done with the 'varpart' function (R, 'vegan' library). The calculations were based on R_a^2 because unadjusted R^2 values provide biased estimates of the explained variation (Peres-Neto et al. 2006).

To get a better picture of pine-canopy-caused heterogeneity, the three categories of grassland plots (open, partially shaded and deeply shaded) were compared by one-way ANOVA in terms of soil properties and quantitative parameters of vegetation.

All statistical analyses employed R version 2.13.1 (R Development Core Team 2011).

Results

Habitat Conditions

The habitat conditions prevailing at the study site are summarized in Table 1. More than half of the plots were on flat or almost flat terrain. Most of the plots sloping more than 5° had a northern exposure. Northern and southern exposures prevailed. Soil was poorly developed: the A horizon was about 3 cm thick on average and the sampled topsoil contained large shares of coarse particles (cf. Fig. 2). Soil pH was slightly alkaline and showed very little variation across the study site. Other soil properties varied more but their coefficient of variation (CV) did not exceed 50 % except for water-soluble Mg. The plots differed largely in terms of pine canopy cover: 48 plots were beyond the reach of shade from tree branches and the others were moderately (36 plots) or strongly (40 plots) shaded by canopy (Fig. 2). Pine stand age (determined for 8×8 m squares) averaged 12 ± 4 years (minimum 0, maximum 23; oldest trees established in 1979) and was fairly uniform throughout the study site (CV 23 %). By contrast, stand basal area, averaging $9.8 \pm 8.1 \text{ m}^2 \cdot \text{ha}^{-1}$ (minimum 0, maximum 52.1), showed relatively high variation (CV 83 %).

The interpolated values (used in statistical analysis) of both parameters are presented in Table 1.

Vegetation Characteristics

We found 56 species of herbaceous vascular plants in the survey. The number of species per plot ranged between 3 and 23 (mean 16 ± 3); very few plots were species-poor (< 8 species per plot; Fig. 2). The most frequent plants were *Pimpinella saxifraga* (119 plots), *Thymus pulegioides* (117 plots), *Festuca ovina* (116 plots), *Leontodon hispidus* subsp. *hastilis* (116 plots) and *Lotus corniculatus* (109 plots). The grassland was composed mainly of plants characteristic of dry grassland and steppe (on average 7 ± 2 *Festuco-Brometea* species per plot, with 15 ± 10 % total cover), hay meadow and pasture (on average 6 ± 2 *Molinio-Arrhenatheretea* species per plot, with 13 ± 10 % total cover), and temperate sand grassland and therophyte sward (on average 2 ± 1 *Koelerio glaucae-Corynepherea canescentis* species per plot, with 16 ± 17 % total cover). Species typical of other plant communities, including ruderal and forest, seldom occurred. Most plants observed in the study were long-lived: perennial (44) or at least biennial (7). Among these, hemicryptophytes dominated (36). As regards Grime C-S-R strategies (Grime et al. 2007), only C-S-R plants (35) and competitors (10) were numerous. There were 43 forbs, 9 grasses and 4 legumes. The grassland was loose, with total cover not reaching 100 % on any plot (mean 47 ± 20 %, minimum 2 %, maximum 98 %; Fig. 2). The species that contributed most to total cover were the above-listed most frequent plants, especially *Festuca ovina*, which dominated in several cases (covering more than 50 % of a plot). There were also seedlings of woody plants in the ground layer (not taken into account in the analysis of species richness and composition). The most frequent (present on 56 plots) and abundant was *Pinus sylvestris*, and seedlings of *Euonymus* sp., *Juniperus communis* and *Quercus robur* occurred sporadically as single plants. Cover of the bryophyte and lichen layer was low, averaging 10 ± 15 %, and 87 values did not exceed the mean.

Decomposition of Variation in the Plant Community Data

In RDA, all forward-selected variables (Table 2) accounted for 20.4 % of the variation in plant species

Table 1 Descriptive statistics for pine stand parameters, SAI and soil properties.

Variable	Unit	Minimum	Median	Mean±SD	Maximum
Stand age ^a	years	7.1	11.8	12.1±2.3	19.6
Stand basal area ^a	m ² ·ha ⁻¹	4.2	9.5	9.7±3.0	16.7
SAI		-0.71	0	-0.07±0.20	0.37
A horizon thickness	cm	0.8	3.1	3.2±1.3	7.2
Soil fine particles	%	23.8	58.9	59.1±11.8	85.8
pH		7.3	7.7	7.7±0.2	8.2
Total N	%	0.11	0.33	0.35±0.13	0.81
Olsen P	mg·kg ⁻¹	8.1	19.1	21.0±10.2	84.6
Water-soluble K	mg·kg ⁻¹	25	62	64±24	138
Water-soluble Ca	mg·kg ⁻¹	37	80	89±41	261
Water-soluble Mg	mg·kg ⁻¹	15	34	40±26	199
Water-soluble Zn	mg·kg ⁻¹	4.4	9.2	9.5±2.9	18.3
Exchangeable K	mg·kg ⁻¹	32	113	120±44	223
Exchangeable Ca	mg·kg ⁻¹	648	2216	2228±752	4661
Exchangeable Mg	mg·kg ⁻¹	108	395	419±176	1215
Exchangeable Zn	mg·kg ⁻¹	33	113	118±56	261

^a Variables obtained by interpolation (see text for explanation).

composition. Pine stand parameters captured 6.1 %, environmental factors captured 9.2 %, and spatial variables captured 13.9 % of the variation (Fig. 3). Both the overall and the pure effects of the three explanatory datasets were statistically significant. Pure effects of the pine and spatial components exceeded joint effects. For the environmental component, joint effects, especially those involving the spatial component, predominated.

Table 2 shows that canopy cover was the variable contributing the most to the variability in the species data. It was responsible, together with pine stand age, for the main shift in herbaceous community structure (RDA axis 1; Fig. 4): from dry grassland with light-demanding species, such as *Euphrasia stricta*, *Gypsophila fastigiata*, *Plantago lanceolata* or *Thymus pulegioides*, towards assemblages with more shade-tolerant species (e.g. *Viola rupestris*). Other important changes in the plant community occurred along the gradient of soil properties (RDA axis 2; Fig. 4). With decreasing amount of coarse particles in topsoil, increasing A horizon thickness and increasing content of some macronutrients, pioneer species (e.g. *Biscutella laevigata*, *Silene vulgaris*) were replaced by species typical of more fertile habitats (e.g. *Ranunculus acris*, *Plantago lanceolata*, *Carex hirta*, *Achillea millefolium*).

In partial RDA, in which the relevant spatial variables (those from Table 2) were used as covariables, only two variables – canopy cover and soil fine particle fraction – were forward-selected to explain the variability of species data. Since these variables also played the primary roles in the previous RDA (Fig. 4), the compositional response of the plant community revealed by the two analyses was largely the same.

In multiple regression, 34.2 % of the variation of plant species richness was explained by all forward-selected variables (Table 3). The overall effects of the three explanatory datasets were statistically significant (Fig. 3). Stand age (the only pine stand parameter in the analysis) captured 9.6 % of the variation, environmental factors captured 13.3 %, and spatial variables captured 28.4 %. Pure effects of the spatial and environmental components were relevant whereas the pure effect of the pine component was virtually absent. Joint effects were important when they involved the spatial component.

Differences Between Open and Shaded Plots

Simple correlation analysis showed that pine canopy cover was related to other parameters of pine stand, i.e., age ($r=0.35$, $P<0.001$) and basal area ($r=0.48$, $P<0.001$). Hence, the differences detected by one-way

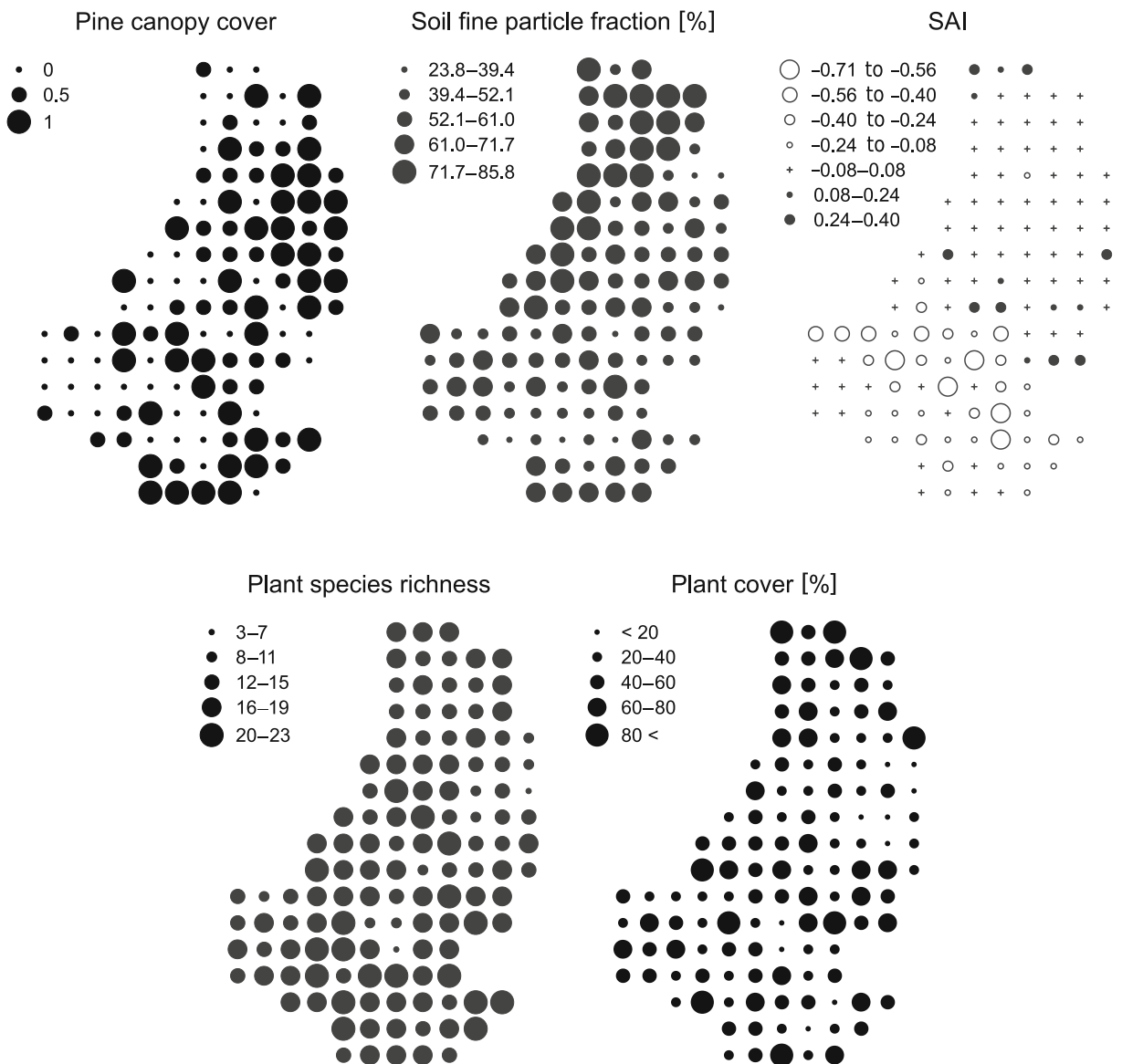


Fig. 2 Maps of selected habitat properties and herbaceous vegetation parameters.

ANOVA between open and shaded plots probably reflected both varying light conditions and duration of pine impact. In one-way ANOVA, A horizon thickness and the content of both water-soluble and extractable Ca and Mg were the only soil characteristics differing significantly between shaded and non-shaded plots (Fig. 5): the A horizon was thinner and availability of Ca and Mg was higher (data shown only for water-soluble Ca) under pine branches than in open plots. The herbaceous vegetation was significantly species-richer and denser in plots fully exposed to sunlight than in deeply shaded plots (Fig. 5). This pattern was

observed for most plant groups (representing different plant traits) examined separately. Opposite trends were also noted (e.g. for competitors or *Molinio-Arrhenatheretea* species), but they were not statistically significant in any case (data not shown).

Discussion

The concentration of heavy metals in soil is one of the strongest determinants of plant community structure (Shaw 1990). Its effect is evident in the grassland we

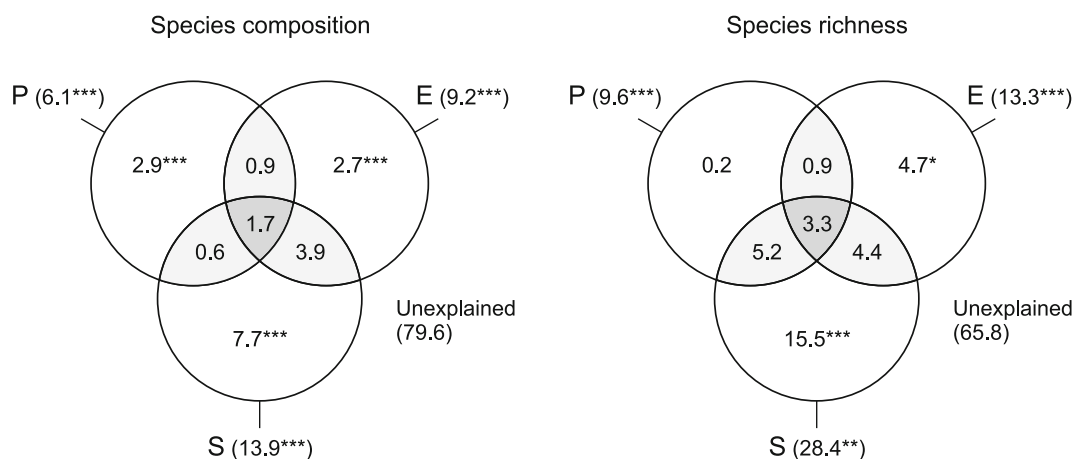
Table 2 Variables selected for RDA of species distribution data on the basis of forward selection.

Explanatory variable	RDA		
	R_a^2	F-ratio	P-value
<i>Pine stand</i>			
Canopy cover	0.042	6.4	0.001
Stand age	0.010	2.3	0.002
Stand basal area	0.009	2.2	0.003
<i>Abiotic environment</i>			
SAI	0.025	4.1	0.001
Exchangeable Ca	0.021	3.7	0.001
Soil fine particles	0.019	3.5	0.001
A horizon thickness	0.008	2.0	0.004
Water-soluble K	0.005	1.6	0.027
pH	0.005	1.6	0.028
Total N	0.005	1.6	0.030
Exchangeable Mg	0.004	1.6	0.036
<i>Space</i>			
Y	0.038	5.8	0.001
PCNM-4	0.025	4.3	0.001
X	0.014	2.9	0.001
PCNM-5	0.014	2.9	0.001
PCNM-17	0.012	2.6	0.001
PCNM-2	0.010	2.3	0.001
PCNM-7	0.006	1.8	0.007
PCNM-24	0.006	1.8	0.014
PCNM-8	0.005	1.7	0.026
PCNM-15	0.005	1.6	0.021
PCNM-34	0.004	1.6	0.017

studied; it is composed of plants known for their heavy-metal resistance, including *Biscutella laevigata*, *Silene vulgaris*, *Festuca ovina*, *Dianthus carthusianorum*, *Cardaminopsis arenosa* (syn.: *Arabidopsis arenosa*), *Armeria maritima* and *Viola tricolor* (Grodzińska and Szarek-Lukaszewska 2009).

Although metal contamination of soil affects the species richness and composition of this grassland, its proxy in our study (water-soluble and exchangeable Zn) presented low explanatory power, probably because of the factor's low spatial variability. The sampling area was small, located entirely on the northern slope of the heap, whose waste material (subsoil) was of the same geological origin and was largely homogenized. As a result, the samples did not differ much in available Zn concentration or in other soil properties. We did not determine the total content of heavy metals, but data on total Zn and Pb collected for this area by Grodzińska and Szarek-Lukaszewska (2009) suggest that it varies very slightly on the heap (*CV* ca 25 %).

The grassland community responded negatively to Scots pine encroachment, manifested in a significant reduction of the frequency and abundance of many light-demanding species of different vegetation classes, and in a decrease of total plant species richness and cover. In view of the paucity of detailed reports on early succession of heavy-metal grassland, it is difficult to say whether this result is typical. A few authors mention that these communities can be seriously threatened by spontaneous encroachment of woody vegetation or afforestation (Baker et al. 2010; Baumbach 2012), but they

**Fig. 3** Venn diagrams partitioning the variation of species composition and species richness between three sets of explanatory variables (adjusted R^2 in %): P – pine stand parameters, E – abiotic

environmental factors, and S – spatial variables. Significance levels: * – $P < 0.05$, ** – $P < 0.01$, *** – $P = 0.001$.

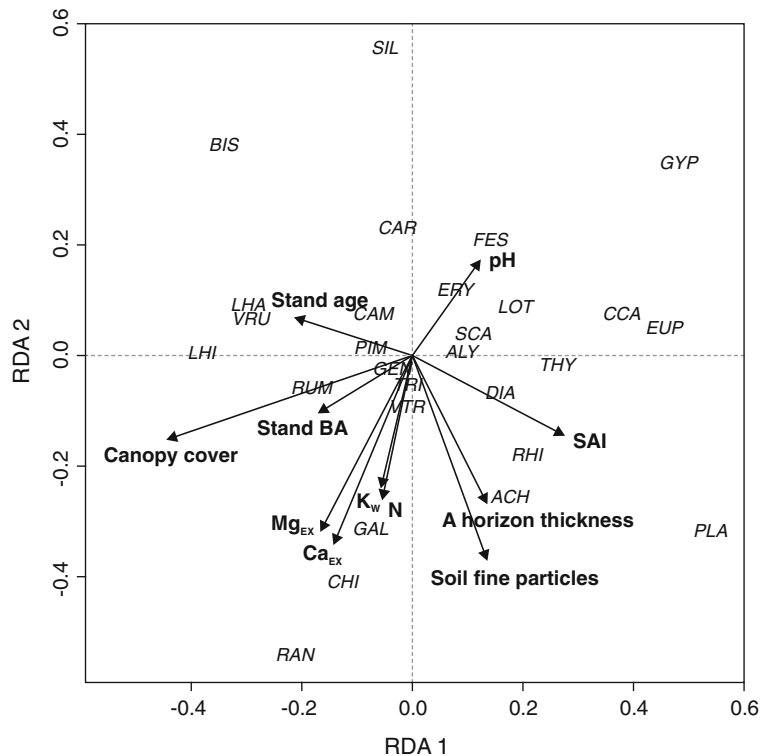


Fig. 4 RDA ordination of plant community data in relation to the best predictors obtained by forward selection (see Table 2). Axis 1 and 2 accounted for 6.3 and 5.2 % of the variation in the data, respectively. For clarity, only species with relevant scores are displayed, using the ‘orditorp’ function (R, vegan package). Species codes: ACH – *Achillea millefolium*, ALY – *Alyssum montanum*, BIS – *Biscutella laevigata*, CAM – *Campanula rotundifolia*, CAR – *Carlina vulgaris*, CCA – *Carex caryophylllea*, CHI – *Carex hirta*, DIA – *Dianthus carthusianorum*, ERY – *Erysimum odoratum*, EUP – *Euphrasia stricta*, FES – *Festuca*

ovina, GAL – *Galium album*, GEN – *Gentianella germanica*, GYP – *Gypsophila fastigiata*, LHA – *Leontodon hispidus* subsp. *hastilis*, LHI – *Leontodon hispidus* subsp. *hispidus*, LOT – *Lotus corniculatus*, PIM – *Pimpinella saxifraga*, PLA – *Plantago lanceolata*, RAN – *Ranunculus acris*, RHI – *Rhinanthus minor*, RUM – *Rumex acetosa*, SCA – *Scabiosa ochroleuca*, SIL – *Silene vulgaris*, THY – *Thymus pulegioides*, TRI – *Trifolium pratense*, VRU – *Viola rupestris*, VTR – *Viola tricolor*. Other abbreviations: BA – basal area, EX – exchangeable, w – water-soluble.

provide little information on the dynamics of herbaceous vegetation. Given our results and the scant published evidence, it can be expected that under favourable conditions (e.g. lack of disturbances, availability of seeds of woody plants, insufficient metal toxicity), heavy-metal grasslands will quickly succeed into woodlands, following the pattern described many times for a similar community type – dry grassland (e.g. Rejmének and Rosén 1992; Davies and Waite 1998; Willems and Bik 1998).

The most valuable components of the studied plant community are species that make the Olkusz heavy-metal grassland unique: *Biscutella laevigata*, a local metallophyte (Babst-Kostecka et al. 2014), and *Gentianella germanica*, whose population has apparently adapted to elevated soil loads of heavy metals (Grześ 2007). They co-occurred with metal-tolerant

populations/ecotypes of common dry grassland species such as *Silene vulgaris*, *Leontodon hispidus* s.l., *Festuca ovina* and *Potentilla arenaria*. All these plants are regarded as remarkable pioneers or at least early successional species. In fact, they grew mostly on very loosely vegetated plots with shallow, skeletal soil, or on bare mining waste rock (see also Wójcik et al. 2014). They are reminiscent of the well known metallophyte communities developed on bedrock material of natural (primary habitats) and anthropogenic (secondary habitats) metalliferous sites (Ernst 1974; Baker and Proctor 1990; Punz and Mucina 1997; Brown 2001; Becker and Brändel 2007).

The aforementioned species are considered light-demanding, often a feature of pioneer plants. In this study, however, they clearly did not respond to the insolation gradient. This was especially true of *Biscutella*

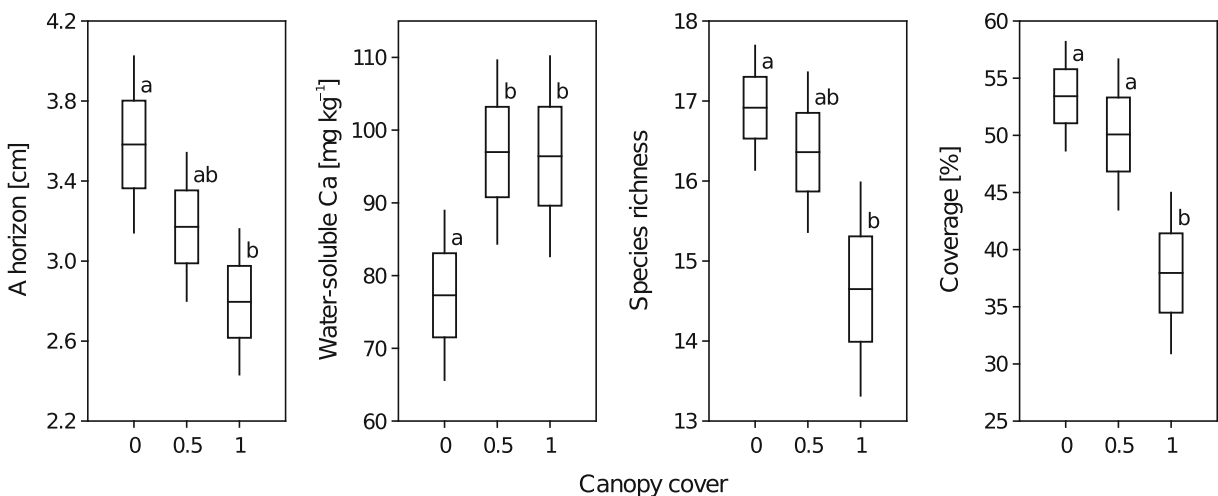
Table 3 Variables selected for multiple regression of species richness on the basis of forward selection.

Explanatory variable	Multiple regression		
	R_a^2	F-ratio	P-value
<i>Pine stand</i>			
Stand age	0.096	14.1	0.001
<i>Abiotic environment</i>			
Water-soluble K	0.036	6.0	0.012
pH	0.034	5.6	0.018
Water-soluble Mg	0.032	5.2	0.027
Water-soluble Zn	0.031	4.9	0.027
<i>Space</i>			
PCNM-27	0.075	11.6	0.003
Y	0.067	9.8	0.007
PCNM-12	0.050	8.5	0.006
PCNM-14	0.048	8.5	0.010
PCNM-4	0.026	5.3	0.020
PCNM-54	0.018	4.0	0.046

laevigata, *Silene vulgaris* and *Leontodon hispidus* s.l., which occurred abundantly both on open plots and on those shaded by pine canopy. Interestingly, this observation is reported from other parts of the OOR area. According to Grodzińska et al. (2010), many grassland species, including metallophytes and metal-tolerant species, can be found, though not in large numbers, in

relatively dense 20–40-year-old Scots pine monocultures planted for land rehabilitation directly on mining waste or on metal-polluted soil. The rare *Biscutella laevigata* and *Gentianella germanica* are among them. Suitable substrate and the presence of tree canopy gaps enable their survival in these forest habitats. Brown (2001), who surveyed some metalliferous sites in Germany, highlighted the occurrence of *Silene vulgaris* in open birch stands, suggesting that this species can tolerate moderate shade.

Scots pine can affect grassland vegetation not only by shading. Pine litter may suppress establishment of certain herb species through physical impediment and release of allelopathic substances (Navarro-Cano et al. 2010). As shown by Xiong and Nilsson (1999), the size of this effect depends on litter depth (or biomass). In this study, the accumulation of pine needles under shaded plots was small (<1 cm) and, given the findings by Xiong and Nilsson (1999), can be considered a negligible factor. Conifers, as well as some species of deciduous trees, tend to acidify the soil (Finzi et al. 1998; Thelin et al. 1998), which may lead to mobilization of heavy metals accumulated there (Wilson and Bell 1996). We did not observe the action of such a mechanism (no differences in pH and Zn availability between open and shaded plots), perhaps due to the recency of the pine invasion. Instead, we recorded increased concentrations of water-soluble and exchangeable Ca and Mg under pine canopy. This implicates pines in the provision of some nutrients to the topsoil, for example

**Fig. 5** Comparison of selected vegetation parameters and soil properties between the three canopy cover categories (0 – open plots, 0.5 – partially shaded plots, 1 – deeply shaded plots) using

one-way ANOVA. Boxes bearing different letters differ significantly according to Tukey's test ($P < 0.05$).

in the form of litter fall (Clarholm and Skjellberg 2013). Pine trees also seem to have influenced the topsoil structure. Shading caused a significant decrease of total plant cover, which averaged less than 40 % on deeply shaded plots. This probably resulted in lower accumulation of fine mineral particles and organic matter and/or greater susceptibility of the soil surface to erosion, and finally in a thinner A horizon. According to Hodačová and Prach (2003), who compared reclaimed and spontaneously revegetated spoil heaps after brown coal mining, a dense herb layer protected slopes better against rill erosion than dense tree stands with reduced herb cover. Of course, the negative relationship between soil depth and canopy cover might be the result of Scots pine's tendency to establish and grow in shallow soil rather than deep soil. However, our data did not confirm that. The correlation between the abundance of pine seedlings and A-horizon thickness was insignificant, suggesting that the former explanation is more likely. The adverse impact of pine trees on soil formation process is beneficial to pioneer plants (e.g. *Biscutella laevigata*, *Silene vulgaris*); this factor may compensate for unfavourable light conditions.

Variation partitioning showed that the grassland vegetation responded differently to different pine stand parameters. Species composition was determined primarily by canopy cover whereas species richness depended more on stand age. The explanation for this seems straightforward. The effect of pine canopy is apparent as a compositional shift in the community from the very outset of pine encroachment because shading clearly reduces the abundance of some shade-intolerant plants. So long as the pine trees are scattered, however, the reduction of the number of grassland species is negligible even under trees, since the still-extant open patches ensure a steady supply of grassland species seeds. As the stand ages, the open area contracts due to the increasing size and density of trees, and the grassland species lose their microhabitats for reproduction and start to withdraw (cf. Rejmének and Rosén 1992). If this process goes too far, it may be difficult to restore the community to its original state, as shown by many studies of semi-natural grasslands of non-metalliferous sites (Dzwonko and Loster 1998; Bakker and Berendse 1999; Bisteau and Mahy 2005; Dzwonko and Loster 2007; Piqueray et al. 2011). The difficulty is due mostly to the nature of the grassland soil seed bank, which is essentially transient (Bisteau and Mahy 2005), and to short-range dispersal of the majority of grassland species (Poschod et al. 1998).

The slope-aspect index (SAI) was among the important explanatory variables of plant species composition. It correlated with many species, including *Rumex acetosa*, *Leontodon hispidus* s.l., *Campanula rotundifolia*, *Biscutella laevigata* and *Dianthus carthusianorum* (negatively), and *Plantago lanceolata* (positively). Since this index is calculated from slope values, it can reflect the intensity of erosion and thus soil depth. In this study, the index did not correlate with A horizon thickness, nor with soil fine particle fraction. The effect of SAI may therefore be considered in terms of microclimate. As such, it seems to explain the distribution of *Biscutella laevigata*. This plant is a mountain species (Babst-Kostecka et al. 2014) which only exceptionally occurs in lowland localities. In the studied grassland, it preferred colder and wetter plots, that is, those with a northern exposure; the same was observed by Dobrzańska (1955). Possibly, these plots offered microclimatic conditions most similar to those at higher elevations. As regards other species, further studies are needed to elucidate the potential role of SAI in their distribution.

Most of the variation of plant species composition and richness was not explained. A random component, which tends to be dominant in field studies, obviously was largely responsible for that. It is also possible that some important spatially unstructured predictors were not taken into account in the models. Most of the explained variation was captured by spatial variables and spatial structures present in the pine and environment datasets. The primary role of spatial dependence in shaping the grassland vegetation was likely the effect of the fine spatial scale of the study. Topographical features (slope, aspect, elevation, etc.) sampled on such a scale usually show high spatial autocorrelation. If these features vary considerably over the study area, they will induce – by processes such as erosion, accumulation or runoff – spatial continuity of other environmental characteristics, and consequently of vegetation.

Maintaining the studied grassland's present species richness and composition would require complete removal of the woody vegetation, repeated as needed. For a similar community, a semi-natural calcareous grassland on non-metalliferous soil, Dzwonko and Loster (2007) recommended cutting down the woody vegetation every 5–6 years. To protect plants characteristic of metalliferous substrate, the intervals presumably can be longer because the flagship species, and some accompanying metal-tolerant plants, survive and apparently

reproduce under pine canopy. On the other hand, these species are sensitive to eutrophication and tend to withdraw from places where fine soil and organic matter completely cover the rocky substrate. Therefore, controlled disturbance of the soil surface, including removal of topsoil or mixing it with subsoil (Baker et al. 2010), should also be incorporated into plans for active management of the site. According to Lucassen et al. (2008), such management not only creates areas of bare soil or subsoil suitable for pioneers, but may also increase the availability of heavy metals, which benefits metallophytes.

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