Assessing Drivers of Forest Conservation in Simancuang Village Forest, West Sumatra

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Abstract

This study empirically assessed Social Forestry program implementation in Simancuang Village Forest or locally known as *Hutan Nagari* (HN) Simancuang in West Sumatra, Indonesia. We performed two analyses using primary and secondary data, namely propensity score matching to estimate the effects of the enactment of HN Simancuang in 2012 on forest cover loss and ordinal logistic regression (OLR) to predict the determinants of conservation awareness. The results of the forest cover analysis showed that forest cover loss in HN Simancuang between 2012 and 2019 was 0.038 percentage point lower than the adjacent protection forest. The relatively small impact was meaningful because although HN Simancuang is located much closer to settlements which increases the pressure on the forest, it could still maintain lower tree cover losses than the adjacent protection forest. This result indicated a certain degree of conservation awareness among HN Simancuang members, which prompted us to conduct a survey to 111 individuals from different households. To do this, we used the Ecosystem Services framework to conceptualise conservation awareness in HN Simancuang. Our OLR results showed that regulating and provisioning services of forests are the strong determinants of conservation awareness among the individuals in our sample. Our study indicates the need to implement social forestry program monitoring and evaluation, improve access to facilitation, and enhance agroforestry practice as the means to increase conservation awareness among forest-dwelling communities.

Keywords Village forest · Social forestry · Environmental services · Conservation awareness · Indonesia

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Introduction

Forests play a vital role in providing natural resources, supporting rural livelihoods, and supplying environmental services (de Groot 1992; Maryudi and Krott 2012; Elis et al. 2019). Yet, forests are subject to deforestation and degradation due to anthropogenic pressures (Spracklen et al. 2015). At the local level, deforestation and forest degradation threaten rural populations, whose livelihoods depend on forest-based resources (Sunderlin et al. 2005). To reduce such pressures, community-based forest management (CBFM) has been promoted globally to reduce forest destruction while simultaneously improving rural development (Persha et al. 2011; Resosudarmo et al. 2014).

Approximately one-third of global forests are managed under various CBFM schemes (FAO 2016). In Indonesia, the government is mainstreaming participatory forest management through the Social Forestry¹ (*Perhutanan Sosial*) policy, regulated primarily under the Minister of Environment and Forestry (MoEF) Regulation no. P9/2021 (see Supplementary Table 1). According to the regulation, Social Forestry consists of five schemes, namely community forests (*hutan kemasyarakatan*), village forests (*hutan desa/nagari*), community plantation forests (*hutan tanaman rakyat*), customary forest (*hutan adat*), and partnerships (*kemitraan*). As a common pool resource, the enactment of MoEF-managed forest estates (*kawasan hutan*) historically limits forest-dependent communities living inside or adjacent to the forests to extract and benefit from forest resources (Ostrom, Gardner and Walker 1994). With the growing view that involving local people to control forests proved to be effective in reducing both environmental degradation and poverty (Pokharel et al. 2007; Kanel and Acharta 2008; Ojha et al. 2009; Maryudi et al. 2012), Social Forestry was introduced to achieve this dual-goal by legally democratizing forest management to local communities.

Despite the Government's agenda to accelerate the progress of Social Forestry by allocating 12.7 million ha of forest area to local communities (Bappenas, 2014), the effectiveness and equity of Social Forestry programs at the local level is often questioned and contested (Agarwal 2001; De Royer et al. 2018). A systematic review of 86 studies on Social Forestry in Indonesia indicates that the success of Social Forestry in Indonesia is often primarily measured based on the direct economic contribution from forests to local livelihoods (Rakatama and Pandit 2020), for example, through sales of commodities and direct subsistence from forests (Lacuna-Richman 2012). However, other studies argue that the economic contribution of Social Forestry might not be sufficient to reduce the economic needs of the local people (Maryudi and Krott 2012; Galvin et al. 2018). Further, deforestation due to conversion to agricultural land might still occur along with the initial stage of economic growth among rural communities involved in CBFM (Tucker and Richards 1983; Richards and Tucker 1988; Agrawal et al. 2013), relegating environmental and forest degradation as a potential side effect of CBFM (Agrawal 2003).

¹ CBFM and Social Forestry can be used interchangeably and for consistency, we will use Social Forestry to better reflect the actual name of the CBFM program in Indonesia.

We argue that direct economic improvement remains the main criterion determining successful Social Forestry in Indonesia. However, other benefits provided by forests, such as ecological improvement and social inclusion, should also be considered when assessing the achievement of Social Forestry. Unlike the other Social Forestry schemes, village forests are mainly aimed at developing the village and improving the welfare of its inhabitants through sustainable forest management (Riyanto 2008). The benefits gained by the locals depend on the status of the forest estate where the village forest is located (either in a protection forest or a limited production forest) and on the forest's biophysical characteristics (Bock 2012). For example, village forests within a protection forest may be used only for the production and collection of non-timber forest products.

A success story from the Lubuk Beringin village forest in Indonesia shows that local communities could even gain monetary benefits from conservation efforts directly through potential Payment of Environmental Services via REDD+ (Feintrenie and Martini in Colfer and Pfund 2011; Bock 2012; Sari 2013). These findings are aligned with Birch et al.'s (2014) findings that the impact of village forests is largely influenced by socio-economic factors closely linked with the community's dependency on forests.

Our study site, Hutan Nagari Alam Pauh Duo Jorong Simancuang (HN Simancuang) in West Sumatra, represents an interesting case where the community members appear not to be economically and directly dependent on forests (e.g. through agroforestry practices). Yet, HN Simancuang, which was among the first village forests established in Indonesia, has also won the Wana Lestari Award from MoEF in 2016 (Muhammad 2019) in recognition of efforts conducted by HN Simancuang community members to conserve their forests (see MoEF regulation no. P.43/2016).

In this research, we aim to understand the factors driving community members to conserve forests. Using HN Simancuang as a case study might help us better understand the dynamics of Social Forestry practices on the ground and identify some driving factors of successful Social Forestry in Indonesia. First, we estimated the impact of village forestry in forest conservation by comparing tree cover loss trends between 2000 and 2019 in HN Simancuang with an adjacent protection forest. To ensure that the difference in tree cover loss between the two forest areas is due village forestry permit, we employed propensity score matching method to control the covariates that may confound deforestation (Santika et al. 2017; Carlson et al. 2018; Putraditama et al. 2019). Second, we assessed the driving factors behind the community's conservation awareness using an ordinal logistic regression (OLR) model based on data from a household survey we conducted in 2018. Together, these spatial and statistical analyses provide insights on the drivers that motivated local communities in forest conservation and how such drivers affected the forests after the designation of HN Simancuang.

Conceptual Framework

Conservation Awareness

To study the community's motivation in achieving conservation goals, such as minimising deforestation, we started from the concept of conservation awareness.

Despite the wide range of literature discussing conservation awareness, there is no consensus on its definition. Some scholars see conservation as strict nature preservation for the sake of nature itself without considering its benefits for people (see Soule 2013), whereas others define conservation more broadly as the intended actions which contribute to establishing or maintaining good relations with nature (Leader-Williams et al. 2010; Sandbrook 2015). The latter definition provides flexibility as it does not exclude human elements in nature protection. Further, Kaiser, Midden and Cervinka (2008) explain conservation behaviour or awareness as actions driven by conservation goals, i.e. protecting nature from any harm.

In the HN Simancuang's context, conservation awareness among the community has been present since the initial settlement and paddy fields were established in 1974 (Fauzi et al. 2019). According to the HN Simancuang management board (LPHN),² there was an agreement among the dwellers that logging the forest would be prohibited as it serves as the village's primary water catchment. This rule was formalised and strengthened when the HN Simancuang permit was granted in 2012 (Fauzi et al. 2019). Under the terms of the permit, the community is not only prohibited from timber logging, but also encouraged to plant high canopy tree species to improve forest cover (Fauzi et al. 2019). These indicate at least two major forms of conservation awareness with regards to forest protection among the locals of HN Simancuang: (1) logging prohibition and (2) high canopy tree planting. Based on these aspects, we collected site-level empirical findings to better understand the conservation motives in HN Simancuang.

Ecosystem Services

Ecosystem services (ES) is an evolving concept that explains the active and passive utilisation of ecosystems to produce or support human wellbeing (Fisher et al. 2009). The Millennium Ecosystem Assessment (2005) was among the first to connect ES with human well-being by designating four ES categories: provisioning, regulating, cultural, and supporting services. The Millennium Ecosystem Assessment's ES classifications have been applied in multiple studies and environmental valuations (Joshi and Negi 2011; Lahjie et al. 2019). In 2010, The Economics of Ecosystems and Biodiversity attempted to emphasise and integrate biodiversity in ES and proposed four ES classifications outlined in Table 1.

Although ES framework has been criticised because the benefits are often limited to the direct ones that are commensurable with economic values, the ES framework provides a clear and relevant logic for policy and decision making concerning ecosystem management (Ellis et al. 2019). Therefore, the ES classifications outlined in Table 1 are used in identifying the predictors of conservation awareness among our

² LPHN or *Lembaga Pengelola Hutan Nagari*, often used interchangeably with LPHD or *Lembaga Pengelola Hutan Desa* or *lembaga desa*, is a village-level management committee established by local communities to manage their Social Forestry areas. The establishment of LPHN is a prerequisite for applying Social Forestry areas to MoEF (see MoEF Regulation no. P.9/2021, similar to the preceding MoEF regulation no. P.83/2016). We use LPHN throughout the paper.

Table 1 Ecosystem Services classifications according to The Economics of Ecosystems and Biodiversity (The Economics of Ecosystems and Biodiversity 2010)

Ecosystem Services	Definition
Provisioning	The material outputs of the ecosystems, such as water, foods and raw materials
Regulating	The role of ecosystems to regulate the surrounding environment, such as local climate and natural hazards prevention
Habitat or Supporting	Providing living spaces for the animals and plants as well as maintaining biodiversity
Cultural	The non-material benefits that humans gain from their interaction with the ecosystem, such as recreational and spiritual purposes

respondents. This study draws heavily on provisioning and regulating services given that these services fit the local context and are perceived by the HN Simancuang's community.

Research Methods

Study Area

We conducted a field study in Kenagarian (village) Alam Pauh Duo, more specifically in the sub-villages (*Jorong*) of Simancuang and Karang Hitam Simancuang, which make up HN Simancuang (Fig. 1).

Both sub-villages were once intact forests, converted into paddy fields and settlement areas in 1974 (Fauzi et al. 2019). In 1992, the area became part of the Alam Pauh Duo village administration. Rice production has since become the main source of income for most households. Given that their livelihood is mainly supported by rice production, only a few people have cultivated non-timber forest products inside HN Simancuang, such as coffee, rubber, and fruits.

In 2010, the LPHN was established, participatory mapping was conducted, and the Social Forestry permit application was submitted under the facilitation of an NGO. In 2011, a temporary permit was granted, and in 2012 the full permit was issued for 35 years. The permit does not allow local communities to expand their agricultural land into the forest due to HN Simancuang's protection forest status. As stated in the LPHN's work plan submitted to the government, the primary objective of HN Simancuang's management is to secure water sources through forest protection.

As of 2018, 210 households and 720 individuals live in both sub-villages, covering 155 ha of paddy fields and approximately 25 ha of residential area. As is the custom in West Sumatra, local communities living around HN Simancuang follow the Minangkabau matrilineal culture. As such, women play a bigger role in the tenurial arrangement, whereby land is owned by the women while the men are responsible for managing it (Fauzi et al. 2019). This arrangement implies that the change in land

Fig. 1 Research location in Alam Pauh Duo, Solok Selatan District, West Sumatra Province. (Data source: Indonesian primary forests by Margono et al. 2014, and village forest by MoEF 2017)

ownership is dependent on the woman's decision. Under the Social Forestry context, forest lands are owned by the state, preventing land-use changes from occurring.

Data Collection

Primary Data Collection

Using a standardised questionnaire (see Supplementary Table 4), we conducted a household survey in HN Simancuang in April 2018 with the assistance of trained, college-educated local enumerators who spoke the local language. Upon data input, we investigated the quality of the data and corrected any mistakes in data input with the enumerators. The entire dataset was then processed with Stata. We interviewed a total of 111 individuals from different households who represented 15% of the total population and 53% of the households in both sub-villages. More specifically, 67 respondents lived in Jorong Simancuang, and 44 lived in Jorong Karang Hitam Simancuang.

The survey was conducted using a non-probability, convenience sampling method since we could not obtain households list in both sub-villages due to data privacy reasons. Convenience sampling has been commonly used where randomisation is not possible due to the large population size and limited available resources (Saumure and Given 2008; Etikan, Musa, and Alkassim 2017). Further, since our study design is non-experimental, using convenience sampling is appropriate. However, such a sampling could suffer from biases, which might make extrapolation to the population difficult (Hedt and Pagano 2011; Clapham and Nicholson 2014; Brodaty et al. 2014).

To reduce the biases, we ensured that the respondents were evenly distributed geographically regardless of road accessibility. The criteria used in selecting our samples include (1) a minimum age of 18; (2) an understanding of households' income and assets (normally head of household or the spouse); and (3) a resident in either of the two sub-villages. We first identified the distribution of populations in both sub-villages and randomly visited the houses within both sub-villages that were readily available (see Henderson et al. 2009).

In addition to the survey, we conducted ten semi-structured interviews and a focus group discussion with the provincial forestry agency, local NGOs, head of village, heads of sub-villages, and several members of HN Simancuang. The purpose of the qualitative data gathering was to gain in-depth information on HN Simancuang history, community management and on-the-ground Social Forestry implementation. Prior to the fieldwork, a scoping study was conducted with an aim to gather preliminary insights for designing the survey and developing questions for the focus group discussion and semi-structured interviews.

Each respondent signed a consent form agreeing to voluntary participation, restricted use of data for research purposes, and personal data protection, among others. Respondents could refuse to answer any questions should they feel necessary. All these procedures complied with the academic research ethical standards (Flick 2014) and approved by the internal review board of the World Resources Institute (WRI) Indonesia.

Secondary Data Collection

We primarily used open-access observational data from various sources to represent variables in the propensity score matching for deforestation analysis (Table 2), which will be further explained in the next section. To ensure the robustness of the analysis, the variables used in this study were determined according to the local context in Alam Pauh Duo and literature.

Data Analysis

Forest Cover Change Analysis

We examined the impact of the designation of HN Simancuang on the forest cover by spatially calculating forest loss before and after the permit was granted in 2012. We employed a quasi experimental design, namely propensity score matching using the *MatchIt* package in R by Ho et al. (2007), to predict the probability of treatment assignment given the covariates listed in Table 2 (Stuart and Rubin 2008; Austin 2011). We performed propensity score matching with 0.25 caliper, replacement and common support. Similar approaches have been used by several researchers focusing on estimating the impact of a certain policy or land use designation on the rate of deforestation (Gaveau et al. 2013; Santika et al. 2017; Carlson et al. 2018; Putraditama et al. 2019). To compare the distribution of treated and untreated cells, we

Table 2 Secondary data collected for forest cover change analysis

Data	Description	Type	Data source
Social Forestry area	Legal boundary of HN Simancuang	Categorical (HN and non-HN)	MoEF, 2017
Forest cover loss	Annual forest cover loss (2000–2019)	Continuous (hectare)	Global Forest Change data (Hansen et al. 2013 and Margono et al. 2014)
Settlement	Cost distance* to settlements	Continuous (meter)	Badan Informasi Geospasial [Geospatial Information Agency]
River	Cost distance* to river	Continuous (meter)	Badan Informasi Geospasial
Road	Cost distance* to road networks	Continuous (meter)	Badan Informasi Geospasial
Slope	Slope	Continuous (percent rise)	NASA SRTM with a 1-arc second or 30 m spatial resolution
Elevation	Elevation	Continuous (meter above sea level/masl)	NASA SRTM with a 1-arc second or 30 m spatial resolution

*In this analysis we employed path distance analysis to determine the accumulated cost of moving from the observed cells to different objects (settlement locations, river and road network) by taking into account the surface cost aspects including topographical elements

Fig. 2 Sampling areas, samples located in **a** HN Simancuang are marked in blue and **b** protection forest are marked in red. (Color figure online)

employed the Wilcoxon rank-sum test to compare both groups (Wilcoxon 1992). In our analysis, the treated cells are located within the HN Simancuang and the control cells are within the adjacent protection forest. We used Alam Pauh Duo, sub-district administrative boundary as our spatial extent.

Prior to estimating propensity scores, we analysed the existence of spatial autocorrelation of tree cover loss between 2012–2019 within the locus using the *gstat* package in R (Pebesma 2004). We found a semivariance asymptote (range) of 222 m, which indicates that the cells closer than this range are affected by the cells nearby, and are therefore not independent. From this estimation, we created independent fishnet grids with an area of 270×270 m or 7.3 ha each cell (see Fig. 2).

Forest cover loss data were treated as the outcome variable. Forest cover loss was produced by overlaying tree cover loss data by Hansen et al. (2013) with Indonesian primary forest boundary (Margono et al. 2014). The spatial resolution of the data is 30 m, which means each pixel represents a 30×30 m area with a value of 1 or “loss” if happened between 2012–2019 (after permit) and 0 if otherwise. Tree cover loss between 2000–2011 was set as the baseline following the temporary permit issuance. Meanwhile, the *control factors* were the biophysical drivers of deforestation, including the proximity to settlements, roads, and rivers, as well as slope and elevation (Laurance et al. 2002; Etter et al. 2006; Rasolofoson et al. 2015; Santika et al. 2017; Putraditama et al. 2019).

Ordinal Logistic Regression

We employed ordinal logistic regression (OLR) models to predict the factors that could potentially increase conservation awareness. An ordinal outcome variable is a categorical variable with a clear category-level ordering where the

outcome value is higher than the previous one and is an extension of logistic regression where the response is linearly related to the independent variables (Williams 2016). The results of OLR were presented in odds ratio where we estimated the odds of having a one-point increase in the conservation awareness score. We used Stata to manage and analyse the dataset particularly using the ‘*ologit*’ command (Liu 2015).

There are two assumptions in OLR: no multicollinearity and proportional odds or parallel line assumptions. Multicollinearity can be tested using Variance Inflation Factors (VIF) after running an ordinary least square model. A VIF value of > 10 indicates a harmful collinearity (Kennedy 1992; Hair et al. 1995). Besides, OLR should not violate proportional odds assumption to ensure that the estimated effects are the same across all categories. To do so, we ran a Brant Test after each OLR model where a significant p-value indicates a violation of this assumption (Brant 1990).

We used two conservation awareness-related actions in this study as the outcome variables which were also emphasised in the Social Forestry’s regulation, namely no logging and tree planting. As HN Simancuang is situated within the protection forest, MoEF Regulation No. P.9/2021 mandated that logging is prohibited, and tree planting and restoration activities are encouraged. The respondents were then asked their level of agreement with these two statements. On no logging, we asked, “*I cannot cut the trees inside the HN Simancuang area*”. Meanwhile on tree planting, we asked, “*I feel the need to plant high canopy trees inside of HN Simancuang*”. The responses for these two statements were captured using a five-point Likert scale where one represented the least agreement and five represented the most agreement. A higher agreement, thus, reflects a higher conservation awareness.

Due to the small sample size and to ensure the proportionality of the outcome variables, we recoded the scale from five to three by collapsing the adjacent scores. Such a method is quite common to deal with scalar or ordinal responses as it does not change the slope (Peterson and Harrell 1990; Greenland 1994). In this case, those answering 1–2 (very disagree and disagree) were recoded as 1 (disagree), those answering 3 (neither agree nor disagree) were recoded as 2 (neither agree nor disagree), and those answering 4–5 were recoded as 3 (agree). In this type of question, respondents might be reluctant to pick the most extreme values (1 and 5) which explained the significantly low n in these two scores (see Supplementary Fig. 1).

We then attempted to predict the factors influencing conservation awareness. Similar to the Moutouama et al. (2019), the HN Simancuang community has relatively high dependencies on forest for its provisioning services (water sources and non-timber forest products) and regulating services (disaster prevention). To construct the dependency on forest resources, we used annual household income, agricultural land size, water access, flood and landslide experience as well as landslide risk perception. The models were controlled by intrinsic factors, i.e. understanding on Social Forestry and environmental protection motivation, and demographic factors.

Table 3 Summary of the characteristics of control variables

Control variable		Protection forests (n = 611 grid cells)	HN Simancuang (n = 55 grid cells)
Distance to Settlement (m)	Min	625.5	525.4
	Mean	4001.0	1295.9
	Max	9132.5	2006.6
Distance to Road (m)	Min	54.0	895.7
	Mean	780.9	1676.5
	Max	2978.7	2739.4
Distance to River (m)	Min	69.5	609.7
	Mean	1785.7	1144.0
	Max	4629.1	1655.2
Slope (%)	Min	12.4	14.7
	Mean	38.7	50.3
	Max	94.0	98.6
Elevation (meter above sea level/ masl)	Min	545.3	796.7
	Mean	936.8	991.7
	Max	1331.4	1248.5

Results

Forest Cover Change

Characteristics of HN Simancuang and the Adjacent Protection Forests in Alam Pauh Duo

Situated in Bukit Barisan, a mountain range in Sumatra, the area surrounding HN Simancuang is relatively hilly. We summarised the biophysical characteristics of HN Simancuang and the adjacent protection forest in Table 3. Our analysis shows that HN Simancuang is generally higher (mean = 991 masl), than the adjacent protection forests in Alam Pauh Duo (mean = 936 masl). Despite the small elevation difference, on average, the slope in HN Simancuang (mean = 50.3%) is steeper than the protected forests (mean = 38.7%). The morphological characteristics influence the spatial distribution of settlement in Alam Pauh Duo. Typically distribution of settlements in mountainous areas are clustered (Zhang et al. 2014). We found a similar pattern in Alam Pauh Duo. Despite its remoteness, the HN Simancuang area is more accessible (e.g. closer to road networks) in comparison to the protected forests. To some extent, this is expected as the village forest scheme is intended for the local communities under a village institution to manage the forest area within their village boundaries.

Further, the average distance to rivers is smaller in HN Simancuang (mean = 1143.9 m) than in the protected forests (mean = 1785.7 m), an indication

Table 4 Control variables before and after matching analysis, after matching the number of units was reduced for both HN Simancuang (treatment) and the adjacent protection forest (control) to 30 and 27 grid cells respectively

Variable	Before/ After Matching	Mean		Differences in means HN & Protection forests	P-value
		HN Simancuang	Protection forests		
Settlement	Before	1295.9	4001.0	− 2705.1	0.000
	After	1472.0	1668.7	− 196.7	0.235
Road	Before	1676.6	780.9	895.5	0.000
	After	1564.1	1917.0	− 352.9	0.316
River	Before	1144.0	1785.7	− 641.7	0.000
	After	1234.7	1321.3	− 86.6	0.488
Slope	Before	50.3	38.6	11.7	0.000
	After	44.0	46.6	− 2.6	0.345
Elevation	Before	991.7	936.8	54.9	0.002
	After	1015.6	1024.0	− 8.4	0.631

that water source protection was the aim of the local community when applying for the HN permit.

Propensity Score Matching Analysis of Forest Cover Loss

We attempted to statistically estimate the effect of HN Simancuang's permit issuance on the forest cover change that happened between 2012 and 2019. propensity score matching allows us to predict the probability of control cells being treated, thus allowing us to estimate the counterfactual group by accounting the covariates. The cells located entirely inside HN Simancuang areas were assigned as the treatment, while cells in the adjacent protection forest were assigned as controls. Through matching, we could reduce the selection bias, making the treatment and the control groups comparable. This could be observed from the differences in means of all control variables where the corresponding p-values changed from significant at 1% level prior to the matching to be insignificant after the matching (Table 4).

We observed that tree cover loss that occurred in the HN Simancuang area between 2012 and 2019 was, on average, 0.038 percentage points lower than the loss occurring in the protection forests (Table 5). The Wilcoxon rank-sum test result indicates that the difference was statistically significant at 1% level (p -value=0.003). Therefore, we can infer that the HN Simancuang designation has a lower forest cover loss compared to the protection forests as a result of Social Forestry implementation with an average treatment effect-on-the-treated of −0.038 percentage points.

In parallel to the propensity score matching analysis, we analysed the forest cover change in HN Simancuang (650 ha) between 2001 and 2019 using tree cover loss data by Hansen et al. (2013). We found that approximately 2.7% of the forest cover in the HN Simancuang has been lost throughout 2001–2019 (17.7 ha). We observed that the use of HN area was concentrated close to the settlement area for agroforestry, which may explain the relatively low deforestation rate within HN

Table 5 The percentage of forest cover loss after the designation of HN Simancuang (2012–2019) before and after matching

Comparison		HN Simancuang (treated)	Protection forests (control)	Mean Difference	P-value
Forest cover loss 2012–2019 (%)	Mean unmatched	0.020	0.015	0.005	0.622
	Mean matched	0.005	0.043	−0.038	0.004

area. Considering that the community plays a crucial role in protecting the forest, understanding the driving factors that influence the local community's conservation awareness is central to reducing deforestation.

Forest Conservation Awareness

Descriptive Statistics

In contrast to most of the Social Forestry schemes which emphasise on the direct use of forests for timber forest products and non-timber forest product extraction, the establishment of HN Simancuang was aimed at securing the area from potential degradation that could affect water supply. The spatial data analysis shows that the main river which spans across Jorong Simancuang and Jorong Karang Hitam Simancuang starts in the HN Simancuang area, making it the primary water catchment area. As a result, HN Simancuang's role is central in maintaining water supply for the locals' consumption and agricultural production.

Based on our survey results, the respondents have relatively high environmental awareness in which tree planting awareness is relatively better distributed across respondents compared to no logging (Fig. 3). Even so, these trends still imply good environmental awareness among most of the respondents with a mean scores of 2.3 and 2.5 in a three-point Likert scale for no logging and tree planting awareness respectively. Table 6 provides greater details on these two outcome variables along with all covariates used in our models. Most of the respondents worked as a farmer or came from a farming household. They have lived in the area for, on average, 25 years, with only 37.8% of them having completed 9-year compulsory education (middle school).

With a mean annual income of IDR18.4 million (USD1,273), respondents in HN Simancuang on average earned IDR1.5 million (USD105.9) per month or around 29% lower than the provincial minimum wage. However, it is important to note that 75.7% of the households surveyed owned rice fields where on average 1.3 tons of rice were harvested in 2017, and 41.3% of which was consumed by the household. Agriculture contributed to 80.3% of the respondents' income. Thus, even though the average income is below the provincial minimum wage, most of them are self-sufficient when it comes to staple food.

Fig. 3 Conservation awareness score distribution (n = 111)

Table 6 Descriptive statistics of key variables

Variable	Mean	S.D	Min	Max
<i>Outcome Variable</i>				
No Logging	2.3	0.9	1	3
Tree Planting	2.5	0.6	1	3
<i>Provisioning Services</i>				
Total Annual Income (in 2018 USD rate)	1,272.5	2,221.3	4.1	13,052
Agricultural Land Size (in Hectare)	1.4	1.4	0	11
Good Water Access	82.0%	0.4		Dummy
<i>Regulating Services</i>				
Landslide Experience	22.5%	0.4		Dummy
Flood Experience	49.6%	0.5		Dummy
High Landslide Risk Perception	46.9%	0.5		Dummy
<i>Intrinsic Factors</i>				
Good Social Forestry Understanding	37.8%	0.5		Dummy
Having Environmental Motivation	87.4%	0.3		Dummy
<i>Demographic Information</i>				
Farmers	69.4%	0.5		Dummy
Household Size (# of People)	4.7	2.0	1	9
Length of Residency (Years)	25.4	12.6	1	70
Completed Middle School	37.8%	0.5		Dummy
Sample Size	111			

We also noted several environmental services that were perceived or identified by the HN Simancuang community. These services include, among others, provisioning and regulating services. The respondents managed on average 1.4 ha of agricultural

land located inside and/or outside of HN Simancuang, mostly for water-intensive rice fields. Approximately 82% of the respondents expressed that they had good water access. As for regulating services, respondents reported that 22.5% of them experienced landslides while almost half experienced floods. These incidents have increased their risk perception on landslides, where approximately 47% expressed their concerns about landslides.

The Drivers of Communities' Conservation Awareness

When we asked the respondents about their motivation to protect the forest, more than 85% expressed environment-related reasons, including disaster prevention and water conservation. We then ran OLR models to see the relationships between one's conservation awareness and HN Simancuang's ES (Table 7).

We estimated the odds of having one-point increase in conservation awareness score, i.e. no logging and tree planting, using the odds ratio for each predictor variable. Generally, the OLR results for both conservation awareness indicate that regulating services of HN Simancuang are better predictors for no logging, while provisioning services explain the tree planting awareness better. For no logging awareness, there seems to be an association between logging and natural hazards perceived by the respondents, especially when they have experienced such an incident in their lifetime. On the other hand, tree planting awareness seems to be associated with 'benefits' from such activity where the provisioning services of HN Simancuang became a good predictor of this conservation awareness.

No Logging

There are six predictors of no logging awareness in HN Simancuang which were significant at the 10% level, namely water access, flood incident, landslide risk perception, farmer, household size, and education. Within the provisioning category, only good water access appeared to be significant where the odds of having a higher no logging score increased by, on average, 2.9 times for those perceiving to have good water access. On the other hand, the odds of having a higher no logging score increased by, on average, 3.9 and 3.5 times for those who have experienced floods in the past and those who have high-risk perception of landslides, respectively. These results indicate an association between tree removal with natural hazards, which has also been observed in the literature (Bradshaw et al. 2007; Malek et al. 2015).

We found a negative relationship between household size and no logging awareness, whereby the odds of having a higher no logging score decreased by, on average, 19.4% for every additional household member. During our interviews with LPHN and survey respondents, it appeared that logging for non-commercial purposes, such as building houses, was still acceptable in areas outside of HN Simancuang. This justification might have undermined the no logging awareness. A similar trend is observed among those who completed at least middle school education where the odds of having a higher no logging score would decrease

Table 7 Ordinal logistic regression results on conservation awareness

	No logging Odds Ratio	Tree planting Odds Ratio
<i>Provisioning Services</i>		
Log Total Income	1.178 (0.219)	1.809*** (0.408)
Agricultural Land Size	1.057 (0.242)	1.608* (0.427)
Good Water Access	2.870* (1.627)	3.286* (2.027)
<i>Regulating Services</i>		
Landslide Experience	0.827 (0.491)	2.947 (2.048)
Flood Experience	3.910** (2.101)	0.341* (0.206)
High Landslide Risk Perception	3.480** (1.848)	4.156** (2.503)
<i>Intrinsic Factors</i>		
Good Social Forestry Understanding	0.767 (0.213)	2.708*** (0.858)
Having Environmental Motivation	2.077 (1.417)	3.757* (2.850)
<i>Demographic Factors</i>		
Farmer	2.675* (1.542)	8.510*** (5.793)
Household Size	0.806* (0.095)	1.030 (0.136)
Length of Residency	1.021 (0.024)	1.014 (0.027)
Completed Middle School	0.424* (0.221)	0.914 (0.511)
Observations	111	111
Hosmer–Lemeshow P-Value	0.143	0.531
McFadden Pseudo R ²	0.179	0.359
Log Likelihood	− 85.02	− 63.23
LR Chi ²	37.11***	70.71***

Robust standard errors in parentheses. Odds ratios are significant at $p < 0.01$ ***, $p < 0.05$ **, or $p < 0.1$ * level

by, on average, 57.6%. One explanation for this is that people with higher education levels have more options to work in sectors other than agriculture, such as being an entrepreneur or government official. Even so, farmers were 2.7 times more likely to have higher no logging awareness than those who worked in other sectors.

Tree Planting

We identified seven significant predictors of tree planting awareness in HN Simancuang, which are significant at the 10% level, namely income, agricultural land size, water access, landslide incident, Social Forestry understanding, environmental motivation, and occupation. For every one percent increase in annual income, the odds of having a higher tree planting score would, on average, increase by 1.8 times. Similarly, agricultural land size appeared to be a predictor of tree planting awareness whereby for every one hectare increase in one's agricultural land size, the odds of having a higher tree planting score increased by, on average, 1.6 times. These trends were likely associated with the fact that, on average, 83% of respondents' income came from agriculture, with farmers being 8.5 times more likely to have a higher tree planting score.

Moreover, those who perceived to have good water access were 3.3 times more likely to have a higher tree planting score. The benefit of HN Simancuang as the water catchment area, both for surface and groundwater, has been reflected in the ease of water access and the income which was derived mostly from water-intensive agricultural lands, such as rice fields. Similar to no logging awareness, landslide risk perception seems to be associated with tree removal, which increased the odds of respondents' awareness of tree planting by 4.2 times. Interestingly, respondents who have experienced floods were 66% less likely to have a higher tree planting score. Although trees or forests do not always work best for flood mitigation (Marapara et al. 2021), there seems to be a lack of understanding among the locals that under the right circumstances, trees and forests do reduce hydrological risks (Keesstra et al. 2018).

Further, we found that those having a good understanding of the Social Forestry program were 2.7 times more likely to have a higher tree planting score. In this case, local facilitators became the forefront of field implementation of the Social Forestry program, including in the HN scheme. Besides, they also introduced the needs of planting high canopy trees. Agroforestry trees were normally chosen to allow non-timber forest product's utilisation by the locals. Those having an environmental stewardship appeared to be 3.8 times more likely to have a higher tree planting score as opposed to those who did not.

Goodness of Fit and Robustness Checks

Both models appeared to have Chi^2 values that were significant at 1% level, indicating that the regression coefficients are not equal to zero. The overall Hosmer–Lemeshow test also indicated an overall good fit of the models with a p -value of 0.14 and 0.53 for no logging and tree planting respectively (Fagerland and Hosmer 2017). We also tested two key assumptions in OLR, namely multicollinearity and proportional odds assumption. The results showed a mean VIF value of 1.4 which indicated that there was no evidence of multicollinearity in our models. Meanwhile, proportional odds assumption was not violated given the $p > \text{chi}^2$ values for both models were not significant, i.e., 0.37 for no logging and 0.54 for tree planting. As such,

the estimated odds resulted from our models are proportional across all categories. Details of these tests are available in the Supplementary Tables 2–3.

Discussion

The tree cover analysis, and ordinal logistic regression that we have run showed three important points to discuss related to (1) deforestation reduction, (2) facilitation access, and (3) agroforestry enhancement.

Reducing Deforestation at the Local Level Through Consistent Monitoring and Evaluation

Improving socio-economic and forest conservation outcomes have become the expectations of Social Forestry. Yet, at the local level, such outcomes are difficult to see without clear evidence. Although both areas show low deforestation rates, the results on forest cover change indicates a lower trend in deforestation within HN Simancuang compared to the adjacent protection forest (see Table 5). Given that HN Simancuang is generally more accessible, thereby having greater socio-economic pressures than the adjacent areas, it is clear that the issuance of HN Simancuang permit has shown a positive contribution in avoiding deforestation. Considering the low deforestation rate within HN Simancuang, the conservation effort can be further improved.

Monitoring and evaluation is essential in controlling deforestation (Boucher et al. 2013; Nepstad et al. 2019). However, such an approach remains lacking at the local level (Burkard, 2009). This view is congruent with a study by Satria (2016) that the local authority in Solok Selatan District experienced difficulties in conducting monitoring and evaluation due to the lack of capacity and human resources to carry out field monitoring. Although there are several monitoring and evaluation frameworks from various NGOs that provide guidance for examining the implications of Social Forestry (see Rahayu et al. 2016; Sudarsono 2016; Fauzi et al. 2019), field monitoring is the most practical approach to utilising the frameworks. Besides, field monitoring is needed since not all local communities have the capacity to report to the authority. In fact, according to the regulation of the Directorate General of Social Forestry and Partnership no. 2/2017, monitoring should be carried out annually while evaluation should be done every five years. Thus, the monitoring and evaluation capacities of both local communities and authorities should be strengthened to ensure that the monitoring and evaluation is well-implemented as it has a huge potential to provide insights on how the local communities could better manage the forests.

Improving Access to Facilitation

While the community is expected to achieve Social Forestry objectives, access to facilitation, support, and training to help the community achieve the goals should

be made available. Our analysis shows that improving the locals' understanding of the Social Forestry program will likely increase their conservation awareness, more specifically on how tree planting could help mitigate flood risks (see Table 7). As such, strengthening environmental education through facilitation that is tailored to promote better awareness on no logging and tree planting may produce positive forest conservation outcomes. However, unequal access to facilitation and support for rural communities remains a challenge in the agricultural and forestry sectors in Indonesia (Chandra 2020). The government could allocate resources to improve access to facilitation and guide facilitators (e.g. local agricultural extension services) to provide the knowledge needed by the local communities to achieve Social Forestry goals.

Forest patrol became one of the LPHN's planned activities which played a role in safeguarding the forest from disturbance, especially illegal logging and mining. Yet, this program has not been consistently implemented due to the lack of financial resources of the LPHN. It appeared that the LPHN and the HN Simancuang community-at-large became the sole actors when it comes to protecting the forest, and its inconsistency might have made forest protection efforts less optimal. In fact, more actors should be involved in the efforts to protect the forest, including the forest management unit or known locally as KPH, the village government, and the HN Simancuang community. Such a co-management approach has been implemented in Bangladesh for CBFM and it was proven to be quite efficient in deterring illegal loggers and improving forest cover from avoided deforestation (Nath et al. 2016).

A similar approach could be developed in the HN Simancuang and other Social Forestry areas in general. This co-management approach could improve conservation awareness among the community members who performed the patrol function (Nath et al. 2016). In doing so, proper budget allocation from the government to the KPH, village fund allocations from the village government, and forest provision taxes should be leveraged in addition to continuous facilitation from the KPH to the LPHN. A balance between conservation, government interests, and community empowerment should be pursued to ensure forest protection efforts to be efficient (Menzies 2007).

Enhancing Agroforestry Practice and Market Access

As indicated in the OLR results, income had a significant influence on tree planting awareness, but not on no logging. A body of literature has also emphasised on how CBFMs like the Social Forestry program in Indonesia have been able to reduce deforestation (Putraditama et al. 2019), whereby dependency on forest-based income from non-timber sources has encouraged forest protection as forest is considered as their livelihood asset (see Scoones 1998, 2015). To increase this conservation awareness, the implementation of agroforestry practice can be enhanced in the HN Simancuang. The trees that will be planted should be high canopy with potential non-timber forest product to diversify farmers' sources of income. A similar strategy has been done in Beringin Jaya Community Forest in Lampung, who received financial support from the MoEF's General Affairs Bureau to plant nutmeg trees in

the forest area (Fauzi et al. 2019). Yet, since not all HN Simancuang members have agricultural plots inside HN areas, ensuring the inclusivity of agroforestry enhancement might be challenging. As such, there should be a clear mechanism that enables non-forest dwellers to participate, such as through the establishment of cooperatives that could facilitate non-timber forest product marketing with or without government support (see Nawir, Murniati, and Rumboko 2007).

However, evidence suggests that agriculture or forest income is associated with deforestation, for example, in the Amazon where timber extraction and land conversion activities are still rampant (Ojeda et al. 2020; Araujo et al. 2019). This could be a loophole in the current Social Forestry implementation in HN Simancuang, where forest patrol mechanism has not been implemented efficiently, if at all. Therefore, the enhancement of agroforestry and other non-timber utilisation of forest resources should be accompanied by a forest monitoring mechanism through the co-management approach that we have suggested.

Conclusions

This article aims to understand the impact of Social Forestry policy implementation in HN Simancuang, West Sumatra, Indonesia on curbing deforestation using forest cover loss as a proxy and how we could tap into community's conservation awareness to optimise forest protection efforts in Social Forestry areas. Using propensity score matching, we estimated that forest cover loss in HN Simancuang between 2012 and 2019 was, on average, 0.038 percentage point less as compared to Alam Pauh Duo protection forest. Although both areas generally have low forest cover loss due to low accessibility, HN Simancuang is located closer to settlements which increases deforestation pressures due to human activities. Yet, our findings showed that HN Simancuang still had a lower deforestation rate than its adjacent areas, which were generally more inaccessible (less human threats and pressures).

In the case of HN Simancuang, we were interested in identifying factors that influence conservation awareness about no logging and tree planting. Our OLR results showed that no-logging awareness was mostly explained by the regulating services of HN Simancuang. Meanwhile, tree planting awareness was mostly predicted by both provisioning and regulating services where HN Simancuang became a major water catchment area for the river and groundwater required by agricultural fields and direct consumption. By understanding the factors that could predict conservation awareness, policies to enhance forest protection and curb deforestation in Social Forestry areas could be better targeted. Since no logging and tree planting are among the common rules in all Social Forestry schemes in protected forests, this approach does not only apply to HN Simancuang or other village forests, but also other Social Forestry schemes situated within protected forests.

Our findings suggest that the implementation of Social Forestry can be improved by maintaining consistent monitoring and evaluation, improving access to facilitation, and promoting the enhancement of agroforestry practice and market access at the local level. Given that the Social Forestry program allows forest-dwelling communities to protect the forest while managing it for livelihood support, we suggest

that forest monitoring and evaluation should be conducted regularly to avoid a counterproductive outcome resulting from unsustainable agricultural practices. On top of these, Social Forestry could benefit from the implementation of co-management approach, especially for forest patrol and agroforestry enhancement.

Although the results of our study can point the policymakers to the right intervention to improve community's conservation awareness, we did not measure how HN Simancuang's conservation awareness was translated into actions. To better grasp the impact of community's conservation awareness, scholars and policymakers should explore the possibility of considering tree cover gain in addition to tree cover loss to get a better proxy of assessing how no logging and tree planting awareness are realised. This domain of study requires further research as there is a limited body of scholarship discussing this issue empirically.

In addition, it is important to note that the OLR estimates in this study are not meant for extrapolation in other contexts given the specific nature of the information and ecosystem services used in the study design. However, the methodological approaches that we employed in this study, i.e. propensity score matching for spatial analysis and OLR for conservation awareness, can be applied in different contexts, especially for measuring the effectiveness of community-based forest management.

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