Exogenous application of plant hormones in the field alters aboveground plant–insect responses and belowground nutrient availability, but does not lead to differences in plant–soil feedbacks

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Abstract

Plant–soil feedbacks of plants that are exposed to herbivory have been shown to differ from those of plants that are not exposed to herbivores. Likely, this process is mediated by jasmonic acid (JA) and salicylic acid (SA) defense pathways, which are induced by aboveground herbivory. Furthermore, exogenous application of these phytohormones to plants alters belowground communities, but whether this changes plant–soil feedbacks in natural systems is unknown. We applied exogenous sprays of JA and SA individually and in combination to field plots in a restored grassland. Control plots were sprayed with demineralized water. After three repeated application rounds, we transplanted seedlings of the plant–soil feedback model plant *Jacobaea vulgaris* as phytometer plants to test the effects of potential phytohormone-mediated changes in the soil, on plant performance during the response phase. We further measured how exogenous application of phytohormones altered plant-related ecosystem characteristics (plot-level); soil chemistry, plot productivity, insect communities and predation. Biomass of the phytometer plants only co-varied with plot productivity, but was not influenced by phytohormone applications. However, we did observe compound-specific effects of SA application on insect communities, most notably on parasitoid attraction, and of JA application on soil nitrogen levels. Although we did not find effects on plant–soil feedbacks, the effects of exogenous application of phytohormones did alter other ecosystem-level processes related to soil nutrient cycling, which may lead to legacy effects in the longer term. Furthermore, exogenous application of phytohormones led to altered attraction of specific insect groups.

Keywords Herbivory · Salicylic acid · Jasmonic acid · Insect communities · Field experiment · Plant–soil feedback

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Introduction

Insect herbivores are important components of ecological processes in plant-dominated ecosystems. They comprise some of the most abundant animal species on the planet, in terms of species diversity and abundance (Price et al. 2011). Therefore, it is not surprising that virtually all plant species host a few or even many herbivorous insect species. Insect herbivores feed on plant tissues, such as roots, leaves, stems, and inflorescences. As a result, insect herbivory can affect the performance of individual plants in plant communities, and the composition of species within those communities (Wilson 1987). For instance, herbivory by insects may reduce rates of succession (Brown 1984, 1985) and plant community composition (Brown and Gange 1992; Crawley 1996; Carson and Root 2000). Some insect species even cause regional outbreaks that can wipe out or set back some

plant species temporarily (Ludwig et al. 1978; Elkinton and Liebhold 1990). In many grassland ecosystems, the effects of insect herbivory on productivity or diversity of plant communities are limited although the impact of insects on individual plants in those communities is considerable (Hairston et al. 1960; Crawley 1989; Bagchi et al. 2014).

When insects feed on a plant, they damage plant tissues locally and release components from their saliva (Ali and Agrawal 2012) that induce the activation of defense mechanisms in the plant (Karban and Kuc 1999; Kessler and Baldwin 2002). Chewing herbivores tend to induce the production of jasmonic acid (JA), whereas phloem-feeding insects and biotrophic fungal pathogens tend to induce the production of salicylic acid (SA) (Thaler et al. 2012). The production of these phytohormones activates a systemic cascade of physiological responses in the host plant that can result in the production of secondary metabolites or the release of volatiles that attract natural enemies of the herbivores, such as predatory arthropods or parasitoids (e.g., Gols et al. 1999; Kessler and Baldwin 2001; Ozawa et al. 2004, 2008; Van Dam and Oomen 2008; Menzel et al. 2014a, b).

Plants deliver many resources to the soil by exuding organic compounds from the roots, and shedding litter into the soil (Steinauer et al. 2016; De Long et al. 2019). These plant-derived inputs influence the community of soil organisms around the roots of the plant (Philippot et al. 2013; Van der Putten et al. 2013; Pineda et al. 2017; Heinen et al. 2018a). For instance, most plants form specific mutualistic relationships with mycorrhizal fungi and interact with plant growth-promoting rhizobacteria in the soil (Gehring and Bennett 2009; Pineda et al. 2010). The soil is also home to many antagonists of plants, such as plant-parasitic nematodes, root feeding arthropods and soil pathogens (Johnson et al. 2012; Johnson and Rasmann 2015). Both JA and SA are important in mediating interactions with belowground organisms. Jasmonic acid plays a role in defense against root feeding insects, nematodes and necrotrophic pathogens (Glazebrook 2005; Thaler et al. 2012), whereas SA defends plants against biotrophic pathogens (Pieterse and Dicke 2007; Pieterse et al. 2009). Manipulation of these pathways can influence patterns of root exudation into the soil (Berendsen et al. 2012) and it has been shown that manipulation of JA and SA pathways can alter the composition of the rhizosphere community (Carvalhais et al. 2013; Lebeis et al. 2015). Plant-mediated changes in soil microbiota influence the performance of plants that grow later in the soil (plant–soil feedback, Kulmatiski et al. 2008; Van der Putten et al. 2013, 2016). Hence, manipulation of JA and SA pathways could also alter plant–soil feedbacks (conceptually represented in Fig. 1). Although foliar application of JA and SA or their chemical mimics is commonly practiced to study plant defenses (e.g., El-Wakeil et al. 2010; Bojórquez Péreznieta et al. 2013; de Freitas et al. 2019), we

are not aware studies that examine the effects of exogenous application of plant hormones on plant-mediated soil legacy effects and how they could alter plant–soil feedbacks in natural plant communities.

In three rounds, we applied exogenous sprays of solutions of SA, JA or a mixture of the two, to grassland plots that contained naturally growing plant communities. Control plots received only demineralized water. Subsequently, we recorded the effects on attraction of insects and on levels of predation on insects in the plots. Further, we assessed whether JA and SA would influence soil processes. We measured soil nutrients and organic matter, and conducted a phytometer plant assay, using *Jacobaea vulgaris*. This species is known to be highly sensitive to changes in soil microbial communities and plant–soil feedbacks (Wang et al. 2019; Van de Voorde et al. 2012; Kostenko et al. 2012). We hypothesized that (i) exogenous application of JA and SA individually to the plant community, through plant-mediated changes in the soil, will have a positive effect on the growth of phytometer plant *J. vulgaris*, compared to their growth in control plots that did not receive JA or SA, and that (ii) exogenous application of JA and SA together, would have additive positive effects on growth of the phytometer plant *J. vulgaris*, compared to their growth in control plots that did not receive JA or SA. Moreover, we hypothesized that (iii) application of JA and SA would suppress the attraction of chewing and phloem-feeding insects, respectively, and (iv) that application would result in attraction of natural enemies to the treated plots.

Materials and methods

Experimental field

The experiment was conducted in an experimental field based in a restoration grassland, ‘De Mossel’ (Ede, The Netherlands), managed by the organization Natuurmonumenten. The field was used for crop production until 1996, after which agricultural practices were ceased and the vegetation was left to develop naturally. Soils in this area are characterized as sandy loam (94% sand, 4% silt, 2% clay, ~5% organic matter, 5.2 pH, 2.5 mg kg⁻¹ N, 4.0 mg kg⁻¹ P, 16.5 mg kg⁻¹ K) (Jeffery et al. 2017). Daily temperatures in the area average 16.7 °C in summer and 1.7 °C in winter. Monthly precipitation averages range from 48 to 76 mm (www.climate-data.org).

The experimental area was fenced in 2016, after the top layer (~5 cm) of soil was excavated to remove existing vegetation. Then the area was left to be colonized by plants from the seedbank and the surroundings. In the following 2 years, the vegetation reached full cover and in spring 2018 at the onset of the experiment, abundant species in

Fig. 1 **a** Conceptual diagram of the hypothesized effects of simulated herbivory, via exogenous application of phytohormones, on soil biota and the hypothesized plant–soil feedback responses in phytometer plants. **b** The randomized allocation of phytohormonal treatments to the selected plots. White plots indicate controls; black plots indicate salicylic acid treatment; red plots indicate jasmonic acid treatment;

blue plots indicate combined jasmonic and salicylic acid treatments. **c** Conceptual overview of a typical plot, where phytometer plants (light circles) were planted equidistant in a triangle shape (minimum 10 cm distance from the plot border) and blue sticky trap (square) were placed in the plot middle. **d** A typical experimental plot, covered by vegetation at the onset of the experiment in May 2018

the communities were the grasses *Agrostis capillaris*, *Anthoxanthum odoratum*, *Aria caryophylla*, *Holcus lanatus*, the forbs *Achillea millefolium*, *Hypericum perforatum*, *Jacobaea vulgaris* and the legumes *Lotus corniculatus* and *Trifolium repens*.

Experimental design

In the experimental area 56 experimental plots were laid out of 60×60 cm each with eight rows of seven plots and 60 cm paths between plots, which ensured that the plots were not directly adjacent and soils of individual plots would be less likely to affect each other. Paths in between the plots were not mowed throughout the experiment. The reasons for this are as follows. First, we wanted to avoid mechanical induction of phytohormonal pathways via mowing that might influence the soils in the paths bordering the plots. Second, the natural vegetation on the paths acted as a natural ‘barrier’ between plots, ensuring that the plots were only different from surrounding vegetation in terms of phytohormonal treatment. Lastly, because the plots were separated only by 60 cm paths, and many flying insects are good dispersers, this leads to the risk that

the aboveground insect collection in the different plots is not fully independent. The barrier posed by the unmown natural vegetation in the plots prevented that insects would directly disperse into adjacent plots, hence limiting the effects of non-independence.

The plots were then randomly assigned to one of four treatments, ($n = 14$ for each); (i) exogenous application of JA solution, (ii) exogenous application of SA solution, (iii) exogenous application of a solution of JA + SA, and (iv) control, with exogenous application of demineralized water (see Fig. 1b). The Netherlands was hit by a drought that lasted from June to September 2018, as well as several heat waves during which temperatures of over 30 °C were recorded regularly (The Royal Netherlands Meteorological Institute (KNMI); www.knmi.nl). During this time, the plots were watered three times a week. The watering events markedly lowered the effects of drought stress in the plots, where plants were green and generally visually healthy, compared to the local vegetation, where plants were in very poor shape (scenescing or dead) well before our sampling events took place. However, the effects of extreme temperatures could not be countered and may have been a potential confounding factor in our results.

Exogenous application of plant hormones

Jasmonic acid (($\pm$)-1 α ,2 β -3-Oxo-2-(*cis*-2-pentenyl)cyclopentaneacetic acid) was obtained in liquid form and SA (2-Hydroxybenzoic acid) in powder form from Sigma Aldrich/Merck (Merck KGaA, Darmstadt, Germany). Both plant hormones were dissolved to 1 mM solution in demineralized water by shaking regularly until dissolved. For each JA or SA treatment plot, a 50 mL tube was filled with 25 mL of the respective solution, mixed with 25 mL demineralized water, giving a solution of 0.5 mM for each. This concentration has been used in previous studies and is known to induce defense responses in plants, but does not reach levels toxic to plants (Thaler et al. 2001; Lebeis et al. 2015). For the additive JA + SA treatments, 25 mL 1 mM JA solution was mixed with 25 mL 1 mM SA solution, so that the resulting solution was 0.5 mM for both plant hormones and the application volume was equal. The control tubes were filled with 50 mL demineralized water. Solutions were applied to the plots using four multi-purpose spray bottles (Gardena/Husqvarna, Ulm, Germany), one per treatment. The sprays resulted in the vegetation being covered by a fine mist, causing dew-like droplets on the plants, and prevented dripping of the applied solutions to the soil (Thaler et al. 2001), which is important, as we were only interested in plant-mediated effects of the hormone treatments. The treatments were applied at three consecutive times on May 4, May 14 and May 29, which were dry, sunny days. The repeated application was done to mimic recurring herbivory for a prolonged period of time, as is likely to occur in nature.

Phytometer plant assay and plot productivity

In order to assess how the plant community responded to the JA and SA treatments and created soil legacies that might lead to plant–soil feedbacks, seedlings of a phytometer plant, *J. vulgaris*, were transplanted into each plot. Seeds of *J. vulgaris* that were used in this study were collected from ‘De Mossel’ area in the summer of 2014. Seeds were separated from their pappus, sterilized using 2.0% hypochlorite solution and rinsed with demineralized water, in order to minimize the seed-associated microbiome. The seeds were germinated on autoclaved glass beads in a climate cabinet (light regime 16:8 h, L:D, day temperature, 21 °C, night temperature 16 °C). In order to maximize survival, the seedlings were transplanted in seedling trays in sterilized soil (γ -irradiation by Synergy Health, Ede, The Netherlands), and left to grow for 3 weeks until they had 4–6 true leaves. Sterile substrate was used to prevent responses of the phytometer plants to potential pre-existing biotic soil legacies present in unsterilized substrate. Three seedlings were then transplanted in each plot, 2 weeks after the last application of plant hormones to the plots, on June 13. A plastic label

was pinned next to each seedling, so that the transplanted seedlings could be relocated. The plants were carefully uprooted and harvested 5 months after planting, on November 7. Roots were washed and root and shoot biomass was dried at 70 °C for at least 72 h, before the dry weights were measured.

After the harvest of the phytometer plants, the biomass of the vegetation of the entire plot was also clipped at soil level and dried. We refer to the total biomass of the plots in the manuscript as plot productivity, and this was used as an independent random factor in the statistical analyses as plot productivity is known to affect both plant–plant interactions (e.g., light, nutrient, space competition) and plant–insect interactions (e.g., visual cues, foraging behavior).

Insect collection in experimental plots

In order to assess how hormone application in the experimental plots affected the resident insect community in each plot, one blue sticky trap of 10 $\times$ 10 cm (Koppert Biological Systems, Berkel en Rodenrijs, The Netherlands) was placed, sticky side up, in the middle of each plot for 24 h (see Fig. 1c). Using a stereomicroscope, the insects collected on the traps were identified to order and specified to suborder in Hemiptera. In the Hymenoptera, the distinction was made between parasitic wasps (predominantly from the superfamily Ichneumonoidea in this study) and ants (Formicidae). Insects were also categorized by their feeding guild; herbivorous insects (i.e., chewing and phloem-feeding insects), natural enemies (i.e., predators and parasitoids), and pollinators (predominantly Diptera).

Predation assay using model caterpillars

In order to assess how hormone application in the experimental plots affected the specific attraction of predators to the experimental plots, a model caterpillar predation assay was performed, modified from two recent studies (Low et al. 2014; Meyer et al. 2015). Specifically, green molding clay (JuniorKnet, Feuchtman GmbH Spielwarenfabrik, Burgbernheim, Germany) was used to create caterpillars of 6 mm diameter and 2 cm length. Caterpillars were pinned on a 2 $\times$ 4 cm piece of cardboard to facilitate placement and relocation in the field (Meyer et al. 2015). Four model caterpillars were placed in each plot (56 $\times$ 4 = 224 model caterpillars) on 18 July and left in the plots for 48 h. After collection, each caterpillar was assessed under a stereomicroscope in order to record presence and type of bite marks on the caterpillar surface. The bite marks were classified to coarse groups (i.e., arthropod, mammal, bird, reptile; Low et al. 2014). As there were only very few (< 5 per group) bite marks caused by mammals, birds or reptiles, for the

data analysis the caterpillars were scored as attacked or not attacked.

Soil characteristics

Two weeks after application of the last round of plant hormones, four soil samples were taken from each plot (1.3 cm diameter, 10 cm depth) and homogenized for soil chemical analysis. Soil samples were air-dried at 40 °C after which the soil was sieved through a 2 mm sieve to remove large stones and root fragments. Three grams of the air-dried soil was transferred to a 50-mL tube and 30 mL of 0.01 M CaCl_2 was added. This mixture was shaken for 2 h on a mechanical shaker with linear movement at 250 rpm. The samples were then centrifuged for 5 min at 3000 rpm and 15 mL of the supernatant was filtered through a Whatman Puradisc Aqua 30 syringe filter with cellulose acetate membrane. To measure soil extractable micronutrients (Fe, K, Mg, P, S, Zn), 12.87 mL of the filtrate was transferred to a 15-mL Falcon tube and 130 μL HNO_3 was added. The sample was mixed using a vortex and analyzed by inductively coupled plasma-optical emission spectrometer (ICP-OES, Thermo Scientific iCAP 6500 Duo Instrument with axial and radial view and CID detector microwave digestion system). The remaining part of the filtrate was transferred to a 15-mL Falcon tube to measure soil pH, $\text{NO}_2 + \text{NO}_3$ and NH_4 . After taking a pH measurement (inoLab pH 7310), the soil extracts were analyzed on a QuAatro Autoanalyzer (Seal analytical, Mequon, Wisconsin, USA).

Soil organic matter levels were determined for each plot by first drying ~5 g (fresh weight) of field soil at 105 °C for at least 24 h. The dried soils were weighed and then burnt overnight at 550 °C, after which they were weighed again. The organic matter was then calculated as the percentage of weight loss relative to the dry weight.

Data analysis

The effects of ‘JA treatment’ (yes/no) and ‘SA treatment’ (yes/no) on plot-level productivity were tested using a two-way ANOVA.

The effects of ‘JA treatment’ (yes/no) and ‘SA treatment’ (yes/no) on root and shoot biomass of the phytometer plant *J. vulgaris* were tested using general linear mixed models. Plot number was included as a random factor to account for multiple phytometer plants that were in the same plot. As growth of *J. vulgaris* is known to be influenced by competition with other plants, we also included plot productivity as an independent random factor in our statistical model.

The effects of ‘JA treatment’ (yes/no) and ‘SA treatment’ (yes/no) on arthropod numbers were analyzed using generalized linear mixed models using a Poisson distribution. As arthropod presence may vary with plot productivity, we

included plot productivity as an independent random factor in our statistical model. Arthropod numbers were only analyzed for the insect orders Coleoptera, Diptera, the parasitic members of the order Hymenoptera, the Hemipteran suborders Auchenorrhyncha, Sternorrhyncha and Heteroptera, the order Thysanoptera, and the arachnid order Aranea. The number of individuals in other arthropod orders were too low for meaningful statistical analysis, but they were included in the group analyses. Arthropods were grouped by functional group (i.e., total arthropods, herbivores, ‘natural enemies’ (predators and parasitoids) and pollinators) and the groups were analyzed using the same model.

The effects of ‘JA treatment’ (yes/no) and ‘SA treatment’ (yes/no) on the proportion of model caterpillars attacked by predators were analyzed using generalized linear mixed models using a binomial distribution. As predator presence may vary with plot productivity (and resulting openness of the canopy), we included plot productivity as an independent random factor in our statistical model.

The effects of ‘JA treatment’ (yes/no) and ‘SA treatment’ (yes/no) on soil chemistry (Fe, K, Mg, P, S, Zn, $\text{NO}_2 + \text{NO}_3$ and NH_4) and soil organic matter levels were tested using general linear mixed models. As soil chemistry may vary with plot productivity, we included plot productivity as an independent random factor in our statistical model.

Model assumptions were checked using QQ- and residual plots, and data were transformed to meet assumptions when necessary (and indicated as such in the respective statistical tables).

All analyses were performed in R Studio version 1.1.419 (RStudio, Inc., Boston, USA) using R version 3.3.1 (R Development Core team 2019). General linear mixed models were performed using the ‘nlme’ package (Pinheiro et al. 2018) and generalized linear mixed models were performed using the ‘lme4’ package (Bates et al. 2015). P values from the latter were obtained using the `car::Anova` command from the ‘car’ package (Fox and Weissberg 2019).

Results

Plot productivity and phytometer plant biomass

Overall, productivity (aboveground biomass) per plot in plots treated with JA and SA did not differ significantly from controls (Table S1A). Moreover, treatment with JA and SA did not alter shoot or root biomass of phytometer plants that were transplanted in the plots after the treatments had ended (Supplementary Table S1A, Fig. 2a, b). However, there was a positive relationship between plot productivity and phytometer root biomass ($F_{1,54} = 6.0$, $R^2 = 0.10$, $p = 0.018$; Fig. 2c), while phytometer shoot biomass was not related to plot productivity ($F_{1,54} = 2.3$, $R^2 = 0.041$, $p = 0.133$; Fig. 2d).

Fig. 2 Effects of simulated herbivory by exogenous phytohormone (JA and SA) application, mediated via soil, on mean ($\pm$ SE) **a** shoot and **b** root biomass of the focal plant, *Jacobaea vulgaris* per plot, and

relationships of **c** *J. vulgaris* shoots and **d** roots with plot-level productivity. The dashed line represents the mean value in the control treatments

Arthropod community responses

Our arthropod collection yielded predominantly flies, aphids, planthoppers, and parasitic wasps. Various spiders,

grasshoppers, and ants were also collected on the sticky traps during this study (Table 1), but numbers were too low to allow for meaningful statistical analysis and these

Table 1 The effects of phytohormone treatments on average numbers per plot of insect orders and per functional group that were collected on sticky traps in the field. Presented are mean numbers, followed by their standard errors (between parentheses)

	Control	JA	SA	JA + SA
Collembola	0 (0.00)	0.07 (0.07)	0 (0.00)	0.07 (0.07)
Thysanoptera	0.79 (0.50)	2.64 (2.04)	2.14 (0.72)	1.71 (1.05)
Hemiptera—Heteroptera	0.36 (0.17)	0.07 (0.07)	0.36 (0.17)	0.14 (0.10)
Hemiptera—Auchenorrhyncha	4.43 (0.55)	3.14 (0.64)	3.93 (1.04)	4.21 (0.48)
Hemiptera—Sternorrhyncha	0.14 (0.10)	0.21 (0.15)	0.29 (0.13)	0.64 (0.20)
Orthoptera	0.07 (0.07)	0.14 (0.10)	0.07 (0.07)	0.07 (0.07)
Coleoptera	0.64 (0.23)	0.64 (0.25)	0.93 (0.29)	0.64 (0.27)
Hymenoptera—Formicidae	0.36 (0.23)	0.14 (0.10)	0 (0.00)	0.14 (0.10)
Hymenoptera—Parasitica	0.71 (0.19)	0.79 (0.19)	1.36 (0.31)	1.57 (0.42)
Diptera	12.57 (1.10)	13.07 (1.96)	11.21 (1.36)	12.64 (1.86)
Arachnida	0.07 (0.07)	0.14 (0.10)	0.07 (0.07)	0.21 (0.11)
Total arthropods	20.14 (1.31)	21.14 (2.66)	20.43 (3.06)	22.14 (2.51)
Total herbivores	6.43 (0.72)	6.93 (2.01)	7.71 (1.87)	7.50 (1.43)
Total natural enemies	1.14 (0.29)	1.07 (0.16)	1.43 (0.34)	1.93 (0.49)
Total pollinators	13.29 (1.16)	13.93 (1.93)	12.64 (1.48)	14.29 (2.06)

Output of statistical models is presented in Supplementary Table S1B

numbers were thus only included in the analyses of total arthropod numbers.

Phytohormone treatments had little effect on numbers of most of the collected orders and most orders were low in abundance (Table 1; Supplementary Table S1B). However, JA application had a marginally significant negative effect on the number of Heteroptera collected on sticky traps, compared to plots that did not receive JA (Supplementary Table S1B). SA application had a marginally significant positive effect on the number of aphids (Hemiptera: Sternorrhyncha) collected on the sticky traps, compared to plots that did not receive SA (Fig. 3a, Supplementary Table S1B). Similarly, SA application had a positive effect on numbers of parasitic Hymenoptera collected on the sticky traps, compared to plots that did not receive SA (Fig. 3b, Supplementary Table S1B).

JA or SA application had no significant effects on arthropods when they were grouped into functional groups, although SA application had a marginally significant positive effect on numbers of natural enemies compared to plots that did not receive SA (Table 1, Supplementary Table S1).

The total number of arthropods was negatively related to plot productivity ($F_{1,54}=7.8$, $R^2=0.13$, $p=0.007$; Fig. 3c),

and the insect order Diptera, the most abundant order collected in the assay, followed the same pattern ($F_{1,54}=9.0$, $R^2=0.14$, $p=0.004$; Fig. 3d).

Predation of model clay caterpillars

In total, 45% of the caterpillars that were placed in the field, were attacked by predators. Only very few bite marks were caused by birds, reptiles or mammals, but examples of bird and mammal attack marks were observed (Supplementary Fig. S1). Although attack was common, JA or SA application had no effect on the levels of predation in the field (Supplementary Table S1C) and levels of predation were also not related to plot productivity.

Soil characteristics

Phytohormone treatments had little effect on most of the measured micronutrients, or soil organic matter (Table 2, Supplementary Table S1D). However, JA application led to significantly lower levels of nitrate and nitrite in the soil, compared to plots that did not receive JA (Fig. 4a, Table 2).

Fig. 3 Effects of simulated herbivory by exogenous phytohormone (JA and SA) application on mean ($\pm$ SE) number of **a** visiting Hemiptera, and **b** visiting parasitic Hymenoptera, and relationships between

c total insect visitation and **d** visiting Diptera and plot-level productivity. The dashed line represents the mean value in the control treatments

Table 2 The effects of phytohormone treatments on soil micronutrients, soil pH and soil organic matter. Presented are mean numbers, followed by their standard errors (between parentheses)

	Control	JA	SA	JA + SA
NO ₂ + NO ₃ (mg/kg)	0.93 (0.13)	0.82 (0.13)	1.12 (0.17)	0.68 (0.08)
NH ₄ (mg/kg)	2.20 (0.33)	2.33 (0.36)	1.98 (0.28)	1.68 (0.24)
Fe (mg/kg)	0.19 (0.01)	0.20 (0.01)	0.20 (0.01)	0.20 (0.01)
K (mg/kg)	17.16 (1.48)	19.51 (2.57)	18.90 (1.50)	16.66 (1.70)
Mg (mg/kg)	25.97 (0.84)	24.33 (1.13)	25.35 (0.83)	24.76 (0.99)
P (mg/kg)	4.99 (0.17)	4.72 (0.12)	4.84 (0.14)	4.85 (0.14)
S (mg/kg)	1.58 (0.10)	1.57 (0.06)	1.56 (0.08)	1.56 (0.08)
Zn (mg/kg)	2.15 (0.07)	2.01 (0.08)	2.08 (0.10)	2.04 (0.07)
Soil pH	5.12 (0.01)	5.14 (0.02)	5.13 (0.02)	5.12 (0.03)
Soil organic matter (%)	3.92 (0.12)	3.86 (0.04)	3.77 (0.07)	3.81 (0.06)

Output of statistical models is presented in Supplementary Table S1D

Iron and potassium levels both showed marginally significant positive relationships with plot productivity (Fig. 4b, c).

Discussion

In this study we examined how exogenous application of the plant phytohormones JA and SA to grassland community plots, would affect plant–soil feedbacks and ecosystem characteristics under field conditions. We did not find any evidence that exogenous application of phytohormones affects plant–soil feedbacks in this study. However, we found that exogenous application of phytohormones to plant communities can affect the aboveground arthropod community that is associated with the plant community, and can also alter levels of soil nutrients.

We found that application of SA, altered the insect community associated with the plots. Application of SA led to an increased number of aphids on the sticky traps in the field plots, although this effect was only marginally significant. This is contrary to what we predicted, as SA-induced plants are expected to be better defended against inducers of the SA pathway, such as aphids. One explanation may be that SA caused aphids to leave their colonized plants in search of better host plants, as aphids have been shown to settle less quickly on SA treated plants (Cao et al. 2014). Reports of SA effects on aphid performance have previously been shown to range from no effects (Moran and Thompson 2001), to a reduction in their numbers (e.g., Schweiger et al. 2014). Other studies have observed that SA induction affected numbers of chewing herbivores, but the direction is not consistent (Bi et al. 1997; Stout et al. 1999; Felton et al. 1999). We also observed positive effects of SA application on numbers of parasitoids in the field plots. This may indicate that SA application induced plant communities to produce herbivore-induced plant volatiles; SA plays an important role in the production of herbivore-induced plant volatiles (e.g., Ozawa et al. 2000; Van Poecke and Dicke 2002) and

can prime plants for the production of volatiles involved in enhanced attraction and efficacy of aphid parasitoids (Rostás and Turlings 2008; Karatolos and Hatcher 2009). We also observed a marginally significant positive effect of SA application on the total numbers of natural enemies, which was driven by the effects of SA on the presence of parasitoids. Other natural enemies, such as spiders or carabid beetles were low in abundance or absent. The latter is in line with absence of SA effects in our predation assay. Parasitoids and other natural enemies generally use volatile cues for detection of their host or prey, which cannot be mimicked by the clay caterpillars and may explain why there were no significant differences in the predation levels. Although attacks of model clay caterpillars were common, bite marks were generally caused by small arthropods, most likely by ants.

In contrast, we did not find any significant effect of JA application, on the arthropod community associated with the plots. Exogenous application of JA has been used in previous studies to mimic aboveground herbivory and is known to induce plant defense mechanisms (Thaler et al. 1996; Thaler 1999a, b; Van Dam et al. 2000; Heil et al. 2001; Stout et al. 1999; Boughton et al. 2005). Alteration of these plant hormones often yields a strong defense response that mimics a plant's natural response to herbivory (Gols et al. 1999; Bruinsma et al. 2007; Qiu et al. 2009). Jasmonic acid, when applied exogenously, activates a cascade of events inside plant tissues, that eventually leads to the production of secondary defense metabolites that may negatively affect insect performance (Thaler et al. 1996; Van Dam et al. 2000; Bruinsma et al. 2007; Qiu et al. 2009). Further, application may trigger the release of herbivore-induced volatile organic compounds, which can be used as host detection cues by natural enemies, such as predators and parasitoids (Gols et al. 1999; Kessler and Baldwin 2001). Although several studies have shown effects of JA application on the performance and abundance of insect herbivores and their natural enemies, even in the field, in this study it did not result in such effects. Our insect collection revealed that phloem

Fig. 4 **a** Effects of simulated herbivory by exogenous phytohormone (JA and SA) application, on mean ($\pm$ SE) soil levels of $\text{NO}_2 + \text{NO}_3$, and the relationship between **b** soil Iron and **c** soil Potassium and plot productivity. The dashed line represents the mean value in the control treatments

feeders were abundant in the field site during our sampling period. However, chewers, other than a very low number of grasshoppers, were virtually absent on our collection traps. This may be one explanation of the absence of effects of JA application on insect communities. We emphasize that we must be cautious to over-generalize the results found in this study. Although our study shows some consistent patterns in changes in plot-level insect communities, we stress that (flying) insects are often good dispersers. Although we separated our field plots by a barrier of 60 cm of unmown vegetation, there is a risk of non-independence in insect sampling. Future studies are needed to test whether these patterns hold true over a larger spatial scale.

We also measured treatment effects on several standard soil chemical parameters, as these may also be affected by changes in phytohormonal status, as has been shown for herbivory (Bardgett and Wardle 2003). As such, shifts in soil chemical parameters may lead to contrasting plant–soil

feedback responses (Van der Putten et al. 2013). We found that, generally, there was no effect of exogenous application of phytohormones on levels of the soil nutrients NH_4 , Fe, Mg, P, K, Zn, S or on soil organic matter content. However, we observed a slight decrease in nitrate and nitrite levels when JA was applied to the plant communities. One explanation for this may be that plants respond to herbivory by taking up more nitrogen from the soil to compensate for tissues lost to herbivores (Holland and Detling 1990; Knops et al. 2002). Alternatively, JA-induction in plants may lead to changes in soil communities, e.g., through altered rhizodeposition, which could cause changes in soil nitrogen levels. However, we have no evidence that the JA-induced changes in soil nitrogen altered performance of either the focal plants or total plot productivity.

Contrary to expectations, phytometer plants did not respond significantly to the phytohormones that were previously applied to the surrounding plant community. This

may indicate that JA and SA applications did not affect the composition of soil communities in this study, which is in contrast with previous studies that observe effects of manipulation of plant phytohormones in the soil microbial community (Carvalhais et al. 2013; Lebeis et al. 2015). The only measurable response we found in the plant was that *J. vulgaris* invested in root biomass in more productive plots, which is in line with the finding that *J. vulgaris* strongly responds to competition (e.g., Van de Voorde et al. 2012; Jing et al. 2015; Bezemer et al. 2018). *Jacobaea vulgaris* has been used as a model system to study biotic soil legacy effects that contribute to plant–soil feedbacks, and is highly responsive to soil conditioning by different plant species (Van de Voorde et al. 2012) and these responses have been ascribed to, at least in part, shifts in soil communities (Bezemer et al. 2013; Kos et al. 2015; Wang et al. 2019). This plant species, for example, responds differently to soil conditioning, depending on the presence or absence of aboveground herbivores on the plant that conditioned the soil (Kostenko et al. 2012; Bezemer et al. 2013). As such, we would expect that the test plant would have responded to potential shifts in soil microbial communities and is well-chosen to test for potential soil legacy effects.

We may speculate why we did not observe growth responses of the phytometer plants in response to exogenous JA and SA application. Soil legacy effects may be diluted in the field compared to the greenhouse because of the multitude of confounding factors that are present in the field not present in the glasshouse (Heinze et al. 2016; Heinze and Joshi 2018; Schittko et al. 2016; Heinen et al. 2018b). Insects feeding on the plant communities prior to or during the treatments may have caused levels of phytohormone induction to be leveled across plant communities. We have tried to mitigate these confounding effects by applying JA and SA relatively early in the season, when herbivory by insects is still naturally low. As a consequence, the time between the start of our treatments and transplantation of phytometer plants was roughly 6 weeks, which is relatively short in comparison with the more classical plant–soil feedback experiments, that usually condition soils for around 10 weeks. However, recent work by our group in a semi-field setup, started in the same month of the previous year, has shown that conditioning effects of plant species on the soil microbiome are detectable from as early as 1 month after transplanting into soils (Hannula et al. 2019). Finally, the effects of the drought and heatwaves may have had stronger impacts on soil biotic conditions than the phytohormonal treatments (Sheik et al. 2011; Jurburg et al. 2017). What our study shows is that testing plant–soil interactions in natural plant communities, although ecologically meaningful, is prone to high noise. Although some of this noise could be reduced in future work, for instance by planting standardized plant communities or individual plants in common garden experiments, instead of using natural vegetation, we could

also conclude that other driving environmental factors may outweigh the effects of plant–soil feedbacks under natural conditions.

In conclusion, we provide evidence that JA and SA were effective in invoking specific ecosystem responses in the field that are likely mediated via the plants. Although exogenous application of phytohormones resulted in altered interactions with insects aboveground and altered levels of soil nitrogen, we could not confirm the hypothesis that exogenous application of JA and SA to plant communities in the field can affect plant–soil feedbacks in phytometer plants. We argue that future experiments testing the effects of JA and SA on plant–soil and plant–insect interactions under natural conditions should standardize experimental conditions further, to get a better understanding of the role of plant phytohormonal pathways in ecosystem characteristics in the field. At the same time, given that anthropogenic global warming is generating more extreme conditions such as heat, droughts and short-term bursts of heavy rainfall across much of the biosphere waves (Meehl and Tebaldi 2004; Christidis et al. 2015), the results of future studies need to be framed within the context of ecological consequences of human-mediated global changes.

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