



The impact of air pollution on COVID-19 pandemic varied within different cities in South America using different models

Haining Huang¹ · Congtian Lin^{2,3} · Xiaobo Liu¹ · Liting Zhu^{1,3} · Ricardo David Avellán-Llaguno^{1,3} · Mauricio Manuel Llaguno Lazo⁴ · Xiaoyan Ai⁵ · Qiansheng Huang¹

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Abstract

There is a rising concern that air pollution plays an important role in the COVID-19 pandemic. However, the results were not consistent on the association between air pollution and the spread of COVID-19. In the study, air pollution data and the confirmed cases of COVID-19 were both gathered from five severe cities across three countries in South America. Daily real-time population regeneration (R_t) was calculated to assess the spread of COVID-19. Two frequently used models, generalized additive models (GAM) and multiple linear regression, were both used to explore the impact of environmental pollutants on the epidemic. Wide ranges of all six air pollutants were detected across the five cities. Spearman's correlation analysis confirmed the positive correlation within six pollutants. R_t value showed a gradual decline in all the five cities. Further analysis showed that the association between air pollution and COVID-19 varied across five cities. According to our research results, even for the same region, varied models gave inconsistent results. For example, in Sao Paulo, both models show SO_2 and O_3 are significant independent variables, however, the GAM model shows that PM_{10} has a nonlinear negative correlation with R_t , while PM_{10} has no significant correlation in the multiple linear model. Moreover, in the case of multiple regions, currently used models should be selected according to local conditions. Our results indicate that there is a significant relationship between air pollution and COVID-19 infection, which will help states, health practitioners, and policy makers in combating the COVID-19 pandemic in South America.

Keywords COVID-19 · Air pollution · Generalized additive model · Multiple linear regression · South America · Daily real-time population regeneration

Haining Huang and Congtian Lin contributed equally to this work.

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✉ Xiaoyan Ai
jxsfybjyaxy@sina.com

✉ Qiansheng Huang
qshuang@iue.ac.cn

¹ Center for Excellence in Regional Atmospheric Environment, Key Lab of Urban Environment and Health, Institute of Urban Environment, Chinese Academy of Sciences, Xiamen 361021, China

² Key Laboratory of Animal Ecology and Conservational Biology, Institute of Zoology, Chinese Academy of Sciences, Beijing 100101, PR China

³ University of Chinese Academy of Sciences, Beijing 100049, PR China

⁴ University of Franca/UNI-FACEF, Franca 14400160, Brazil

⁵ Jiangxi Provincial Key Laboratory of Birth Defect for Prevention and Control, Jiangxi Provincial Maternal and Child Health Hospital, 318 Bayi Avenue, Nanchang 330006, PR China

Introduction

In late December 2019, a novel coronavirus, named COVID-19, was first reported in Wuhan, Hubei Province, China (Darai et al. 2020; Lu et al. 2020; Wu et al. 2020). Acute Respiratory Syndrome Corona Virus 2 (SARS-CoV2) (Lu et al. 2020) is the pathogenic agent of COVID-19 and most of the infected had clinical manifestations of fever and shortness of breath (Chen et al. 2020). This epidemic has caused serious demographic changes and unemployment (Bashir et al. 2020b). It has been confirmed COVID-19 can be transmitted through direct contact (human-to-human) (Chan et al. 2020). A total of 20,871,160 patients with COVID-19 had been confirmed worldwide as of August 13, 2020, and 81 countries have more than 10,000 confirmed cases (<https://www.hopkinsmedicine.org/coronavirus>). South America accounts for about 27% of the world's confirmed cases, making it the region with the highest number of confirmed

cases. The confirmed cases in South America until August 13, 2020 are shown in Fig. 1.

Air pollution remains a major public health threat globally (Hashim et al. 2021). Previous results indicated that air pollution can increase the spread of diseases (Bell et al. 2004; Goings et al. 1989; Wei et al. 2019). Previous studies showed that droplets with virus can stay suspended in the air for a short time, and these particles may pose a threat of infection if they are inhaled by nearby persons. This approach makes it possible for people infected with COVID-19 to facilitate the spread of infection (Anfinrud et al. 2020; Meselson 2020). Lab experiments have demonstrated that the SARS-CoV-2 virus can survive in aerosols for days or weeks, making the virus susceptible to airborne contamination (Liu et al. 2020). Particulate matters such as PM₁₀ and PM_{2.5}, due to their small size, can easily penetrate into the lower respiratory tract and can carry the virus directly into the alveoli and tracheobronchial region (Qu et al. 2020). Several studies have proven that air pollutants act as a carrier to transmit virus reducing the level of immune system and therefore make human bodies more vulnerable to virus infection (Becker and Soukup

1999, Glencross et al. 2020, Xie et al. 2019, Xu et al. 2020). Air pollutants have been shown to affect the transmission and severity of respiratory viral infections including, but not limited to severe acute respiratory syndrome (SARS), the emergence of the Middle East respiratory syndrome (MERS), as well as SARS-CoV-2 (Cui et al. 2003; Domingo and Rovira 2020; Silva et al. 2014). It has been shown that air pollution is positively correlated with the mortality of SARS in China (Cui et al. 2003). The environment around us is filled with contaminants that can inadvertently expose humans to viruses (Daraei et al. 2020). Although risk factors for COVID-19 are still under investigation, it is possible that environmental factors, such as air pollution, may play a significant role in affecting the spread of the epidemic among the population.

In terms of SARS-CoV-2, multiple studies are showing the significant association between air pollution and the spread rate of the COVID-19. Several recent studies have shown that the risk is significantly higher for individuals contracting COVID-19 where they are exposed to environmental pollutants (Coccia 2021; Liu et al. 2021). Generalized additive models (GAM) showed that six air pollutants (PM_{2.5}, PM₁₀, SO₂, CO, NO₂, and O₃) were significantly related to the confirmed cases in 120 cities from Jan 23 to Feb 29, 2020 in China. Empirical estimates suggested that PM_{2.5} is a significant factor associated with the COVID-19 pandemic in the top 10 most affected states in the USA (Bilal et al. 2021b). In Europe, the most severely affected region is the same as that possessed the highest concentrations of PM₁₀ and PM_{2.5} (Martelletti and Martelletti 2020). Furthermore, most fatality cases occurred in the regions with the highest NO₂ concentration (Ogen 2020). Spearman correlation analysis indicated that PM_{2.5}, O₃, and NO₂ have a significant relationship with the outbreak of COVID-19 (Bilal et al. 2020). The relations were also confirmed in California, the USA, and India (Bashir et al. 2020c; Sharma et al. 2020). In South America, correlation analysis and wavelet transform coherence were used to explore the relationship between environmental pollution indicators and the spread of COVID-19. Results showed that PM₁₀, NO₂, CO, and O₃ are significant factors in the fight against the COVID-19 pandemic (Bilal et al. 2021a).

The impact of air pollution on the epidemic varies from study to study. Thus, the findings have been inconsistent and there were limited compelling reasons on the shape and magnitude of those relationships. Therefore, it is necessary to explore the effect of air pollution on the spread of COVID-19. Tracking the epidemic data and the dynamic variations of these values can help to estimate the spread of this emerging pandemic (Merl et al. 2009). Here, we assemble the datasets of the spread of the COVID-19 pandemic in five regions of South America. The time-dependent reproduction number (R_t) in each area was estimated to assess the expected number of secondary cases arising from a primary case infected during the t period (Thompson et al. 2019). The objective of this

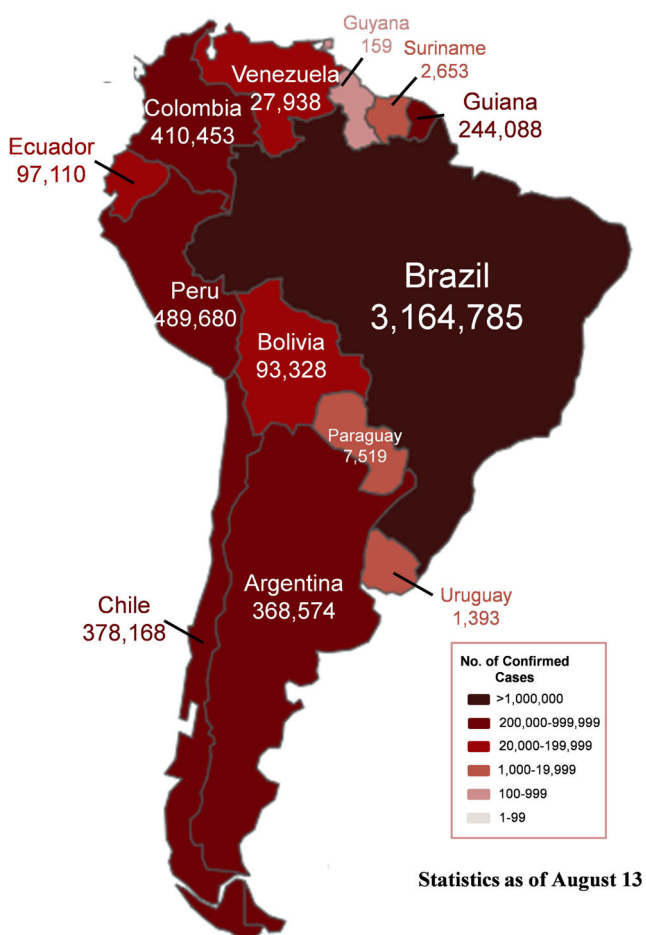


Fig. 1 Confirmed cases of COVID-19 across South America until 13 August, 2020. Data of coronavirus disease 2019 (COVID-19) from <https://www.jhu.edu/>

work is to assess the relationship of different air pollutants on the newly confirmed cases of COVID-19. To evaluate the impact of pollutants on epidemic spread more objectively and comprehensively, two frequently used models, generalized additive models (GAM) and multiple linear regression, were both applied to each city. And the results from both models were compared to explore the impact of air pollution on the spread of the COVID-19, in addition, we also compared the differences in the results of the two models.

Materials and methods

Database of air pollutants and COVID-19 infection

Five regions from three countries in South America, including Sao Paulo, Sao Jose dos Campos, and Vitoria in Brazil, Guayaquil in Ecuador, and Bogota in Colombia, were studied in this work. Time series data of air pollution including six major air pollutants PM_{2.5}, PM₁₀, O₃, nitrogen dioxide (NO₂), sulfur dioxide (SO₂), and carbon monoxide (CO) were obtained from real-time air quality index of Air Pollution in the World database (Data source: aqicn.org/data-platform/covid19/). This website uses the standard for air pollutants from the US Environmental Protection Agency (EPA). And the daily air quality index (AQI) data were then converted to mass concentrations (<https://www.airnow.gov/aqi/aqi-calculator/AQICalculator|AirNow.gov>). The data concerning the number of newly confirmed cases was collected directly from the National Health Department from March 28 to June 10, as it is shown in Table 1.

Estimation of the time-dependent reproduction number (R_t)

R_t , a time-dependent reproduction number, which can reflect the transmission of infectious diseases in the population (Cowling et al. 2010, Wallinga and Teunis 2004), was estimated with the "EpiEstim" package in the R software. Based on the research of the Chinese CDC (Li et al. 2020b), we set an offset gamma distribution with mean of 7.5 days and standard deviation of 3.4 days. The smoothing time was set to 10

days. The epidemic grows when R_t is above 1 and the outbreak will die out once R_t stays below 1. Cross-sectional analysis was performed to examine the spatial association between air pollutants and R_t of COVID-19, and longitudinal analysis was used to examine the temporal associations of air pollutants with R_t .

Statistical analysis

To determine the relationship between each air pollutants, and the correlation between air pollutants and the transmission of COVID-19 (R_t), we used Spearman correlation to assess the associations of air pollutants with R_t , with detection level $\alpha = 0.05$ (bilateral). Based on the analysis of correlation, two frequently used models, multiple linear regression and generalized additive models (GAM), were both used in the study. For multiple linear regression model, the number of R_t was used as dependent variables, and the daily air pollutants were selected as independent variables. The formula used was as follows:

$$Y = \alpha + \beta_1 X_1 + \beta_2 X_2 + \dots + \beta_n X_n$$

In this model, the outcome variable, Y , is thought to be a linear function of a set of predictor variables, where n is the number of predictor variables, α is a numerical constant that represents an intercept. β s stands for the partial regression coefficients of X , each β reflects that how Y will change with the X , which is associated with the β when all other X variables constant (Jaccard et al. 2006). Among them, X s stand for the parameters of air pollution that are significantly associated with R_t .

GAM, developed by Hastie and Tibshirani (Hastie and Tibshirani 1995), was also used to estimate the association between PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and R_t . The fitting of GAM uses nonlinear smoothing term, the regression equation of GAM to predict a regressed variable is shown below:

$$g(E(Y)) = \beta X_1 + \sum_{i=2}^p S_i(X_i)$$

where Y is the predicted values of the dependent variable, R_t ; X_i represents the levels of air pollution, independent variables, and S_i is the nonparametric smoothing function.

Table 1 The source of daily confirmed cases of COVID-19

Country	Regions	Data sources
Brazil	Sao Paulo Sao Jose dos Campos Vitoria	https://covid.saude.gov.br/
Ecuador	Guayaquil	https://coronavirusecuador.com/data/
Colombia	Bogota	https://www.ins.gov.co/Noticias/Paginas/Coronavirus.aspx

According to the different data distribution of dependent variables, different methods are used to fit the model. Popularly used distributions in GAM modeling are Normal, Gamma, and Poisson distributions (Ravindra et al. 2019). In this paper, we applied Poisson distributions to examine the moving average lag effect (7 days) of air pollutions on daily values of R_t of COVID-19 and all Poisson regression analyses were performed in R (version 3.6.2) with the “mgcv” package.

Results

Daily pollutant data

As shown in Fig. 2 and Fig. S1, the median concentration of particulate matter in Sao Jose dos Campos ($PM_{2.5}$, $11.040 \mu\text{g}/\text{m}^3$, PM_{10} , $19.440 \mu\text{g}/\text{m}^3$), O_3 (0.020 ppm), and CO (6.453 ppb) in Colombia, NO_2 (4.558 ppb) in Guayaquil and SO_2 (10.706 ppb) in Sao Paulo were the highest within the five cities, respectively. The concentrations of other pollutants are at similar levels through these five cities. According to Spearman’s correlation coefficient, there is a positive correlation between the six pollutants, most of which are extremely significant ($p < 0.01$), except O_3 , which has a negative correlation, or weak positive correlation (in Vitoria) with other

pollutants in three Brazilian cities (Fig. 3A–3C). In Bogota (Fig. 3D), there are strong positive correlations ($p < 0.01$) between each pollutant, except O_3/NO_2 , SO_2/CO . In Guayaquil (Fig. 3E), the correlation between any of the two pollutants is statistically significant ($p < 0.05$).

Epidemiological data in the selected regions

The calculated R_t (Fig. 4) values showed a gradual decline in all the five regions, particularly in Sao Jose dos Campos, where the peak was 5.56, and then went down to 1.16 on June 10. The R_t value of Guayaquil decreased from the peak of 1.72 to 0.24. By contrast, R_t in Vitoria fluctuated, and it remained above 1 until June 10, indicating that the epidemic situation in the region was still serious.

Model fitting

The results of GAM and multiple linear models are shown in Fig. 5, Figs. S2–S4, and Table 2, respectively. By establishing GAM models between the pollutant factors (explanatory variables) and the R_t response variables, the smooth regression function of explanatory variables is obtained, as well as the effect diagram of influencing factors on R_t (Fig. 5 and Figs. S2–S4). The results show that there is a nonlinear relationship

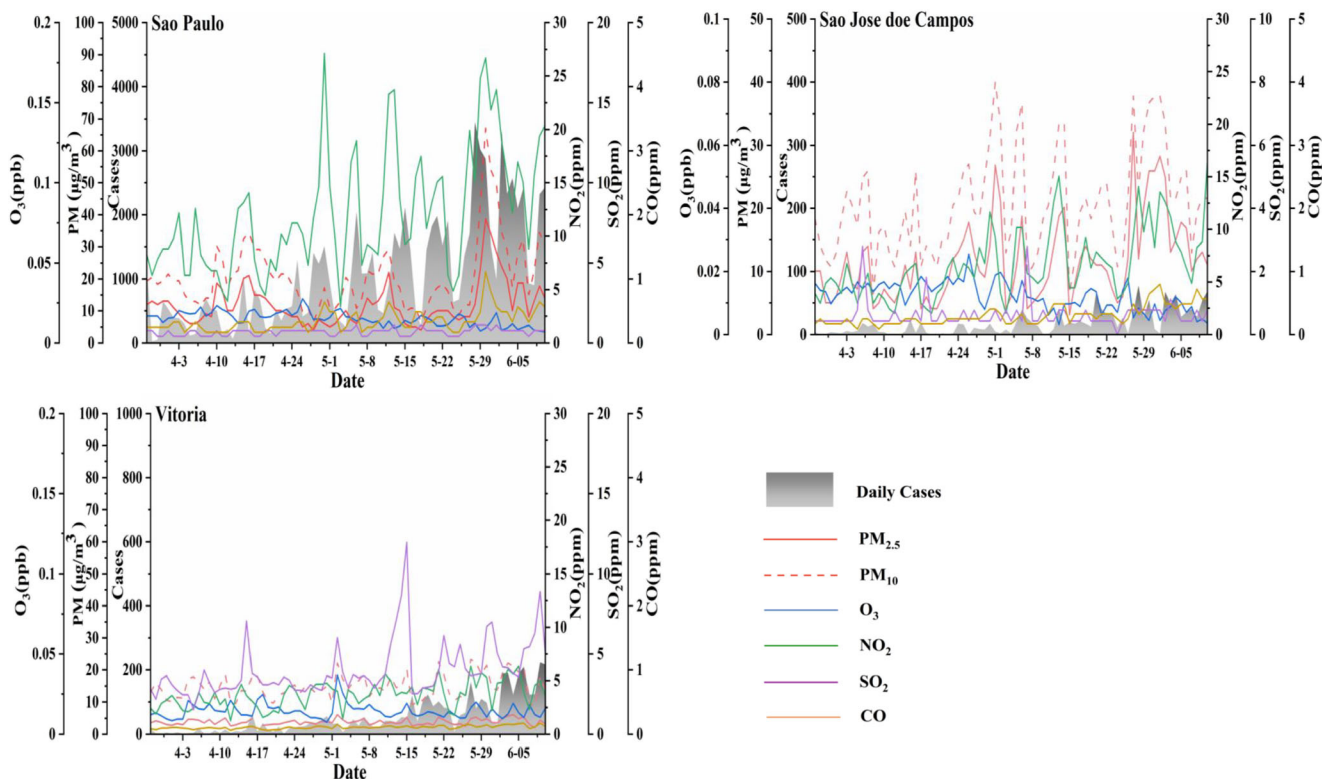


Fig. 2 Daily changes in the number of confirmed COVID-19 cases and air pollution in the selected regions. The gray areas indicate the number of daily confirmed cases. The colored lines represent the pollution changes

over the corresponding time, the red line represents $PM_{2.5}$, the dashed line represents PM_{10} , the blue line represents O_3 , the green line represents NO_2 , the purple line represents SO_2 , and the orange line represents CO

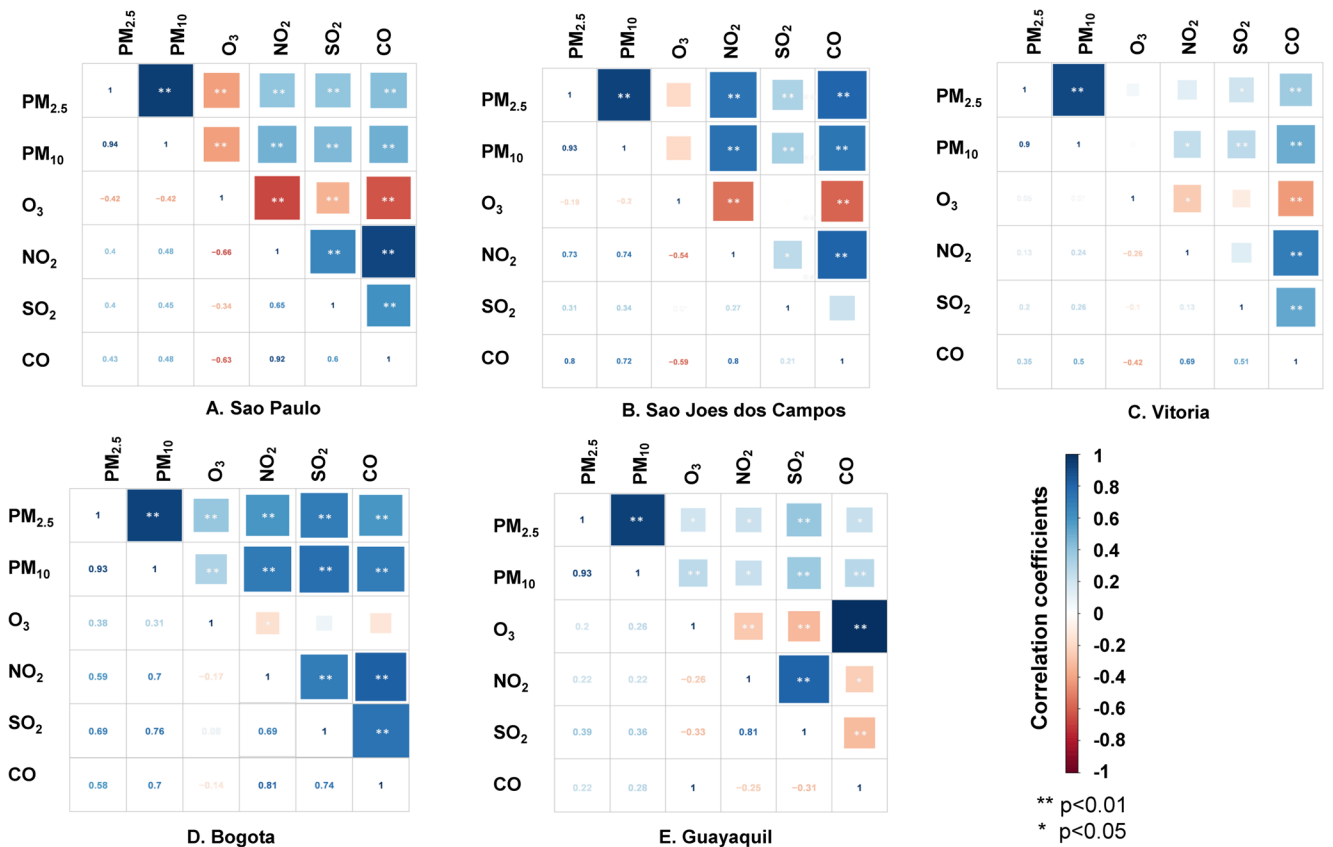


Fig. 3 Spearman correlation between air pollution and R_t in the five regions. The color gradient indicated Spearman’s correlation coefficients. The darker blue indicates a stronger positive correlation,

and darker red indicates a stronger negative correlation. Data significance was marked by * $p<0.05$, ** $p<0.01$

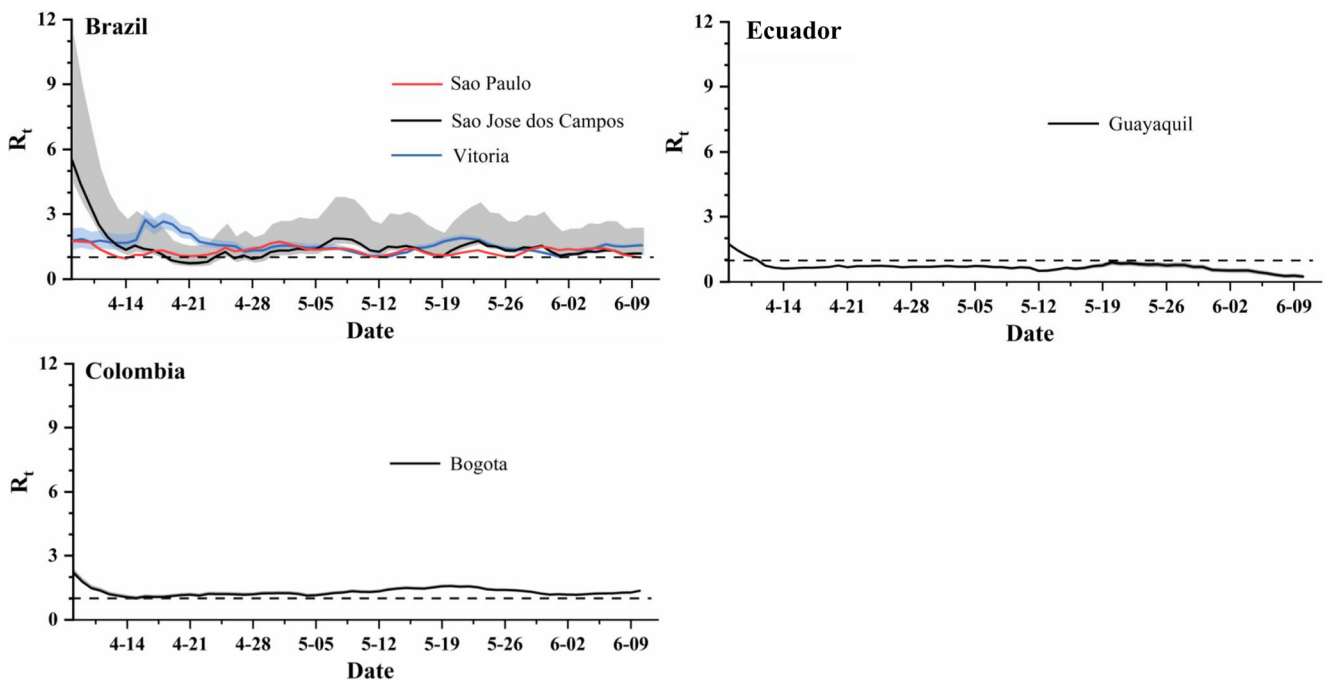


Fig. 4 Daily estimated distributions of the effective reproduction number R_t , based on selected epidemiological data for COVID-19 with 95% confidence intervals, where the dashed line represents the threshold of R_t

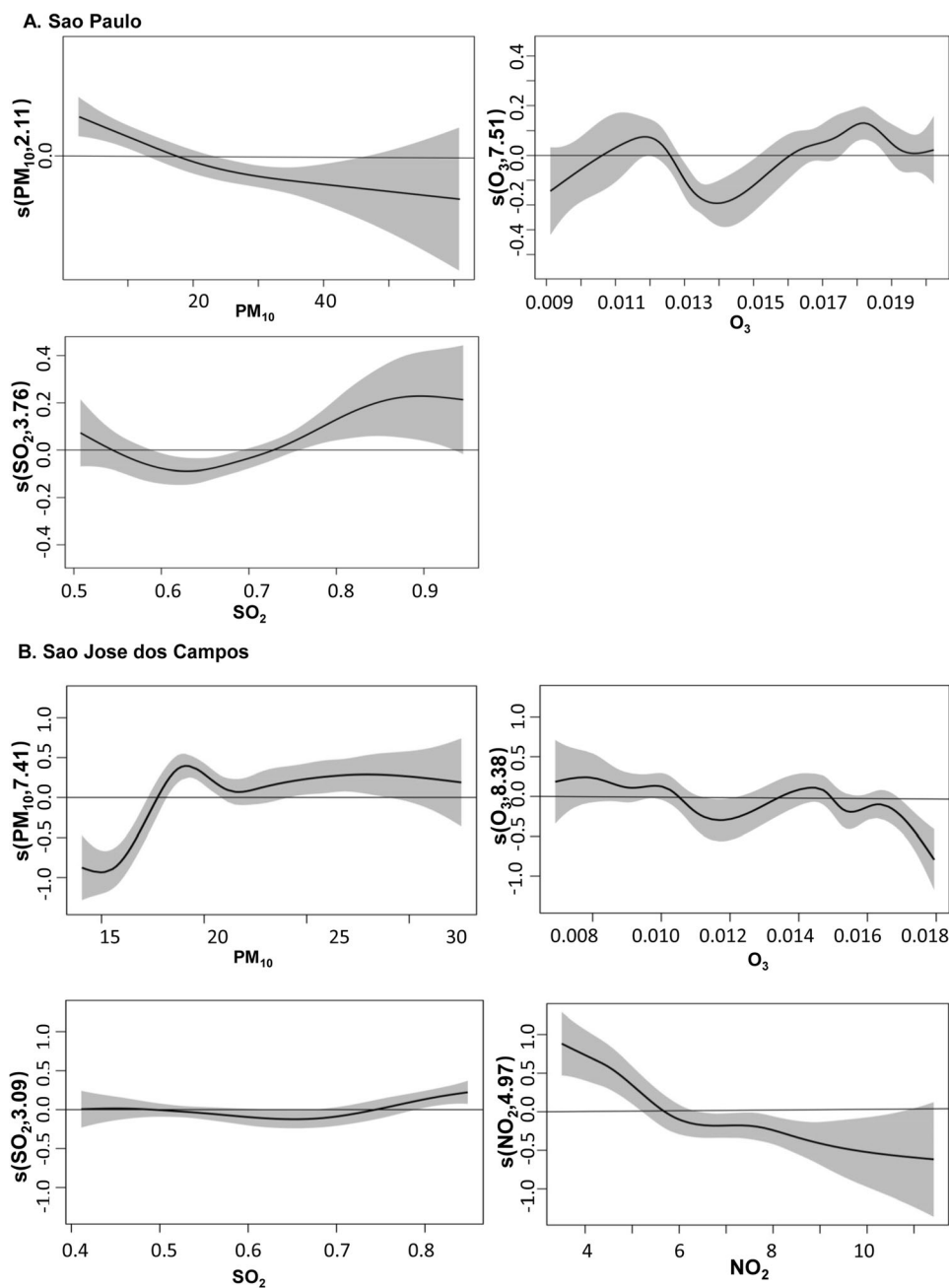


Fig. 5 The results of the GAM model for the effects of air pollutants on the variation of R_t . The gray areas represent the upper and lower limits of the confidence intervals for fitting additive functions, the solid lines represent the smooth fitting curves of R_t , and the horizontal coordinates

represent the measured values of the explanatory variables, ordinate represents the smooth fitting of explanatory variables to R_t , ordinate values in parentheses represent estimated degrees of freedom

between R_t and each explanatory variable in Sao Paulo (Fig. 5A), and R_t value decreases gradually with the increase of PM_{10} concentration. R_t increases monotonically when O_3 concentration is less than 0.012 ppm or between 0.014 ppm and 0.018 ppm. When SO_2 concentration is less than 0.6 ppb, R_t shows a slowly decreasing trend. However, R_t increases with the elevation of SO_2 when SO_2 concentration is higher than 0.6 ppb. In Sao Jose dos Campos (Fig. 5B), R_t shows an increasing

trend when PM_{10} concentration is between 15 $\mu\text{g}/\text{m}^3$ and 20 $\mu\text{g}/\text{m}^3$. And R_t shows a fluctuating downward trend with the increase of O_3 and NO_2 concentration, and a weak change with the increase of SO_2 concentration. In Victoria (Fig. S2), R_t shows a fluctuating downward trend with the increase of PM_{10} and NO_2 concentrations. When O_3 concentration is less than 0.012 ppm, R_t decreases gradually; when O_3 concentration is higher than 0.012 ppm, R_t changes relatively gently. R_t shows

Table 2 Summary of the models

Region	Model	Significant independent variables
Sao Paulo	GAM	PM ₁₀ , SO ₂ , O ₃
	Multiple linear regression	SO ₂ , O ₃
Sao joe dos Campos	GAM	PM ₁₀ , SO ₂ , NO ₂ , O ₃
	Multiple linear regression	/
Vitoria	GAM	PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃
	Multiple linear regression	SO ₂ , NO ₂
Bogota	GAM	PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃
	Multiple linear regression	PM ₁₀
Guayaquil	GAM	PM ₁₀ , SO ₂ , CO, O ₃

Multiple linear regression/

/, no significant independent variables in the model

a large fluctuating change when the SO₂ concentration is below 4.0 ppb and almost no significant change when it is higher than 4 ppb. *R_t* increases monotonically when CO concentration is higher than 0.11 ppb. In Bogota (Fig. S3), with the increase of PM₁₀ and NO₂ concentration, *R_t* shows a certain fluctuation, but the value does not change a lot. When the O₃ concentration is less than 0.004 ppm, *R_t* shows a monotonically increasing trend, while it decreases as the O₃ concentration rises above 0.008 ppm. When SO₂ concentration is less than 0.35 ppb, *R_t* tends to decrease slowly, while when it is higher than this concentration, *R_t* increases gradually. *R_t* shows a fluctuating rising trend with the increase of CO concentration. When CO concentration was higher than 0.42 ppb, there is no obvious change. In Guayaquil (Fig. S4), *R_t* decreases slowly when the concentration of PM₁₀ is lower than 17 μg/m³ and increases slowly when it is higher than 17 μg/m³. *R_t* increases when the O₃ concentration is less than 0.03 ppm and decreases above 0.03 ppm. *R_t* value shows a relatively slow change with the increase of SO₂ concentration. The correlation between CO concentration and *R_t* shows a certain linear relationship, and when CO concentration increases, *R_t* decreases monotonically.

However, based on the results of multiple linearity (Table 3), the magnitudes of β reflect the influence of the corresponding variable, there is a positive correlation between the *R_t* value and O₃ (β= 13.135) in Sao Paulo and a negative correlation with SO₂ (β= -0.320). In Vitoria, *R_t* was negatively correlated with NO₂ (β= -0.147) and SO₂ (β=-0.053).

There is a negative correlation between *R_t* and PM₁₀ in Bogota (β= -0.013). For Sao Jose dos Campos and Guayaquil, there is no linear correlation.

Discussion

In this study, we analyzed the relationship between COVID-19 infection and air pollution in five regions of South America. Our data spans a wider range of time and space and more types of pollutants than previous studies. And referred to the previous studies, two frequently used models, generalized additive models (GAM) and multiple linear regression, were both used for each city. At the same time, the results of the two models in the same region were different. According to our results, *R_t* responds to different air pollutants (PM_{2.5}, PM₁₀, O₃, SO₂, NO₂, CO) in different regions. Although the same significant factors were not obtained using the multiple linear regression model, however, the GAM model showed that PM₁₀ and SO₂ responded significantly to *R_t* in all regions. Previous study shows that COVID-19 confirmed cases is significantly positively associated with PM_{2.5}, PM₁₀, CO, and NO₂, and negatively associated with SO₂ by using the GAM model (Liu et al. 2021). And research on the spread of COVID-19 and environmental quality showed that PM₁₀, NO₂, CO, and O₃ are significant factors in the fight against the COVID-19 pandemic in South America (Bilal et al. 2021a).

Table 3 Statistical data of the multiple linear regression equation

Region	Model formula	R ²	Adjusted R ²
Sao Paulo	$Y_{Rt} = 1.163 + 13.135X_{O_3} - 0.320X_{SO_2}$	0.142	0.069
Sao Jose dos Campos	/	/	/
Vitoria	$Y_{Rt} = 2.120 - 0.148X_{NO_2} - 0.053X_{SO_2}$	0.306	0.246
Bogota	$Y_{Rt} = 1.257 - 0.013X_{PM_{10}}$	0.182	0.111
Guayaquil	/	/	/

Table 4 Comparison of correlational studies between air pollution and COVID-19 in various studies

Environment variable	Date range	Region	Model	Reference
PM _{2.5} , PM ₁₀ , NO ₂ , CO	Jan 26th to Feb 29th	Hubei, China	Linear regression	(Li et al. 2020a)
PM _{2.5}	Mar 1st to Apr 20th	New York City, America	Negative binomial regression	(Adhikari and Yin 2020)
PM _{2.5} , PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃	Jan 23th to Feb 29th	120 cities, China	GAM	
PM _{2.5} , PM ₁₀ , SO ₂ , VOC, CO, NO ₂ , Pb	Mar 4th to Apr 24th	California, America	Spearman and Kendall correlation	(Bashir et al. 2020a)
PM _{2.5} , PM ₁₀	Jan 15th to Feb 29th	Hubei, China	Spatial auto-correlation statistics	(Yao et al. 2020)
PM _{2.5} , PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃	Jan 25th to Feb 29th	Hubei, China	Multivariate Poisson regression	(Jiang et al. 2020)
PM _{2.5} , NO ₂	Feb to Mar	Italy	Pearson correlation	(Frontera et al. 2020)
PM _{2.5} , PM ₁₀ , NO ₂ , O ₃	Feb 24th to Jul 2nd	Germany	Spearman correlation	(Bilal et al. 2020)
PM _{2.5}	Mar 2nd to Sept 17th	the USA	Spearman and Kendall correlation	(Bilal et al. 2021b)
PM _{2.5} , PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃	Jan 22nd to Oct 8th	South American capital cities	Kendall correlation	(Bilal et al. 2021a)
PM _{2.5} , PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃	Jan 28th to May 31st	China	Regression discontinuity design	(Liu et al. 2021)
PM _{2.5} , PM ₁₀ , SO ₂ , CO, NO ₂ , O ₃	Mar 28th to Jun 10th	South America	GAM, multiple linear regression	This study

/, no model in the study

Our findings are partially consistent with these results. Currently, there are limited studies on air pollution and the spread of COVID, and conclusions are differing. There may be regional differences that could have a potential impact on the epidemic transmission, including variation in the timing and coverage of public health interventions (Dalziel et al. 2018). Due to different environmental conditions, even the impact of the same pollutant will vary. We pooled data from similar studies in the past, and summarized the types of pollutants, the study areas, the study time, and the fitting model used (Table 4). All results show that air pollutants are significantly correlated with the spread of COVID-19. Furthermore, there were differences in the results from different study areas, times of coverage, and models used.

Although currently there are some analytical findings, we offer the following limitations in our research work. Our findings do not reveal a clear effect of pollutants on the virus's ability to spread, but our data are broader in space and time than previous studies, and more diverse than most studies. This can give an indication that the effects of pollutants on disease may not be specific, and the results of GAM and multiple linear model in individual cities cannot be directly replicated in other regions. There are shortcomings in the currently used models. The conclusions drawn from a single fitting result cannot be directly applied, and more models can be used in future research.

There are some limitations of this study. First, there may be differences in the timing of the acquisition of epidemic data across regions. Thus, we encourage further research to analyze

the association of environmental indicators with COVID-19 transmission in a wider range of regions to provide more important insights. Second, there are many other models for this kind of prediction. In this paper, GAM and multiple linear models are adopted, and there may be other more appropriate models. Third, there are many other factors that contribute to outbreaks of COVID-19, such as individual behavior, government control measures, urban density, and population. In this paper, only air pollution is considered. Social intervention, as well as population immunity, may have a greater impact on the virus's ability to spread compared to air pollution (Baker et al. 2020). Future research needs to investigate the impact of social intervention, urban density, and population.

Conclusion

Over the past few months, the South American region has become the central region of the COVID-19 pandemic, as the infection rate of COVID-19 has been increasing. The current work investigates six environmental pollutants in South America and their association with the COVID-19 pandemic in the region. According to the experiments and analysis results, this study has come to the following conclusions: (1) R_t , which can reflect the spread of COVID-19, showed a gradual decline in all the five regions. (2) PM₁₀ and SO₂ responded significantly to R_t in all selected regions. Regulators should make better monitoring of these two pollutants. (3) The association between air pollution and the spread of COVID-19

differed in varied cities with specific statuses of air pollution. Inconsistent results were obtained from GAM and multiple linear regression model for one city. For example, in Sao Jose dos Campos, GAM revealed that the four air pollutants PM₁₀, SO₂, NO₂, and O₃ were significantly related to the spread of COVID-19. However, the multiple linear regression model showed that air pollution is not significantly related to the spread of COVID-19. There remains a need to optimize models to assess the contribution of air pollution to the COVID-19 pandemic. As well, more regions need to be studied to reveal the association of air pollution to the spread of COVID-19. Future research should take a comprehensive approach, including consideration of epidemiological aspects, socio-economic issues, and the different lockdowns and mobility restrictions imposed by each country.

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Availability of data and materials The datasets used and/or analyzed during the current study are available from the corresponding author on reasonable request.

Author contribution Haining Huang: conceptualization, methodology, visualization, writing. Congtian Lin: conceptualization, methodology, visualization, Writing. Xiaobo Liu: methodology, data collection. Liting Zhu: data collection, visualization. Ricardo David Avellán-Llaguno: data collection, critical revision. Mauricio Manuel Llaguno Lazo: data collection, critical revision. Xiaoyan Ai: study conception, drafting. Qiansheng Huang: study conception, analysis and interpretation of data, drafting.

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Declarations

Ethics approval Not applicable.

Consent to participate Not applicable.

Consent for publication Not applicable.

Competing interests The authors declare no competing interests.

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