



Optimization of Removal of Phosphate from Water by Adsorption Using Biopolymer Chitosan Beads

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Abstract The need for clean water is the most basic human right. Water scarcity will be one important environmental problem of all countries in the future. Phosphate is a harmful matter for public health and the environment. In this study, the removal of phosphate from water by chitosan, which is an environmentally friendly material, was investigated. Chitosan adsorbent spheres were prepared for phosphate separation from water by adsorption, which is a feasible method. The effects of phosphate concentration, adsorbent dosage, and operation time on the removal were investigated. The removal increased with acid concentration and adsorbent amount. The maximum adsorption capacity of chitosan beads is 87.26 mg/g. Adsorption behavior of the chitosan beads were examined by Langmuir and Freundlich isotherms and pseudo-first and second-order kinetic models. The adsorption process was optimized by the response surface method (RSM). Trial version of Design Expert® 12.0 was used in the study. It has been understood as a result of the RSM statistical analysis that higher phosphate removal values would

be obtained by increasing the amount of adsorbent. ANOVA analysis showed that adsorbent dosage had the biggest effect on removal of phosphate using chitosan beads prepared for adsorption.

Keywords Water · Chitosan · Phosphate removal · Adsorption · Optimization

1 Introduction

Due to climate change, pollution in lake and seawaters as a result of increasing water temperatures causes a serious threat to people and the environment. Phosphate is necessary for the growth of living things, but its excessive presence in the waters induces gigantism of algae, bacteria, and aquatic plants. Phosphate is used in different fields such as fertilizers, medicine, detergents, animal feeds, corrosion inhibitors, and cosmetics (Harech et al., 2022). Overdischarge of phosphate from domestic and industrial wastewater and agricultural activities to lakes and seas blocks sunlight. Failure of sunlight to reach the bottom of lakes and seas causes a serious environmental problem such as eutrophication (Xu et al., 2021, Banu et al., 2020, Chen et al., 2020, Omwene & Kobya, 2018). Thirty-two to seventy percent of the amount of phosphate results from detergents in the urban sewage water. In recent years, consumption habits have changed in the COVID-19 epidemic, and excessive use of detergent has caused eutrophication

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in freshwater resources and seawaters. It is evidenced that the formation of harmful algal bloom (HAB) in lakes and seas has increased due to climate change (Sha et al., 2021). HABs growing in water with excessive phosphate undo aquatic ecosystems irreparably. The threats posed by HABs are mucilage, water discoloration, depletion of oxygen at the bottom, and mass death of bios in aquatic ecosystems. Therefore, discharge of phosphate into the water systems should be prevented (Mackay et al., 2022).

The removal of phosphate from wastewater has been studied for years (Eltaweil et al., 2021). The conventional treatment methods such as biological and chemical techniques are restricted due to the costs of investment and operation, requiring high energy, and the formation of secondary pollutants like the sludge (Pap et al., 2020; Tao et al., 2020). For these reasons, environmentally friendly, sustainable, and innovative approaches like bioremediation, membrane techniques, ion exchange, and adsorption have been developed in recent years (Cheng et al., 2018). Adsorption that has a lower production cost and higher efficiency is preferred as an energy-friendly technique (Eltaweil et al., 2021; Shirazinezhad et al., 2021; Verma & Nadagouda, 2021; Zhao et al., 2020). Adsorbent selection in the adsorption method is an important point. A lot of adsorbents such as metal oxides (Shirazinezhad et al., 2021), graphene (Verma & Nadagouda, 2021), clay (Banu et al., 2020; Chen et al., 2020), active carbon (Han et al., 2020, Karthikeyan et al., 2020), nanocomposite (Yang et al., 2021; Zhou et al., 2014), biochar (Tao et al., 2020; Zhou et al., 2013), and chitosan (Liu et al., 2020, Pap et al., 2020, Mackay et al., 2022; Kong et al., 2023) are used for phosphate removal from water. In literature, it has been seen that the adsorption property of chitosan is increased by modifying it with different materials. The cost of adsorbent would increase with these processes. The adsorbents of biochar modified with chitosan have great availability as green material at lower costs (Zhou et al., 2013). There are studies for phosphate removal with cross-linked pure chitosan also (Filipkowska et al., 2014). In this study, the use of pristine chitosan and rising of adsorption efficiency with the RSM method are significant and notable in terms of sustainability and economy. In the literature, adsorbents are good but have limitations for trace amounts. It is requested

a higher adsorption efficiency for lower phosphate concentrations (Liu et al., 2020). Chitosan is a sustainable, nontoxic, biocompatible, economic, and environmentally friendly material and has high adsorption ability. Chitosan is like a shining material in different areas such as medical, environmental, industrial, pharmaceutical, and wastewater treatment (Eltaweil et al., 2021). Chitosan is a plentiful and low-cost material. Chitosan contains high amounts of amino and hydroxyl functional groups. These functional groups directly affect the adsorption capacity of chitosan, especially amino groups which are strongly active for the adsorption process. The electron pair on the amino groups of chitosan act as an adsorption center (Wang & Zhuang, 2017, Salehi et al., 2022).

Therefore, it draws attention (Filipkowska et al., 2014, Mackay et al., 2022, Liu et al., 2020, Oktor & Hilmioglu, 2021). In this study, adsorbent spheres were prepared in the laboratory. Adsorption tests were performed with 0.1 g, 0.3 g, and 0.5 g chitosan adsorbent spheres at 25 °C for 0.5 mg/L, 1 mg/L, and 1.5 mg/L phosphate concentrations. The operating conditions were determined by optimization with the response surface method (RSM) to achieve high removal rates in waters with different phosphate concentrations untested. Accordingly it has been understood that high adsorption efficiency can be obtained with chitosan for low phosphate concentrations.

2 Materials and Method

Chitosan, epichlorohydrin, NaOH, NaH_2PO_4 , and vanadate-molybdate reactives were supplied from Sigma-Aldrich. In this study, adsorptive beads were prepared with chitosan natural biopolymer to adsorb phosphate in the water. Before and after adsorption, phosphate concentration in the synthetic phosphate solutions was determined by UV–Vis spectrometer using vanadate-molybdate reagent. The adsorption process is optimized with RSM statistical analysis. Trial version of Design Expert® 12.0 was used in this study.

Adsorptive chitosan beads were prepared in the laboratory. 1% wt. chitosan solution is prepared by mixing with 1% wt. acetic acid. It was cross-linked with 0.25% wt. epichlorohydrin. After that, chitosan

sphere beads are prepared by dropping the polymer solution into 2 M NaOH solution. The beads are stirred in a 2 M NaOH solution for a long time to be stable. Then it is washed with pure water. The beads are dried at room temperature and then at 70 °C in the oven.

Synthetic phosphate solution is prepared with NaH_2PO_4 and water. The stock phosphate solution is obtained by dissolving NaH_2PO_4 salt in water. 1 M stock solution is prepared by dissolving a constant amount of solid salt in 1 L distilled water. Different solution concentrations (mg/L) are prepared by diluting this solution.

Solution concentration is determined by using a UV–Vis spectrometer. The solution is colored by vanadate-molybdate reagent, and the UV absorbance value is read at 400 nm. The absorbance-concentration calibration curve is prepared. Solution concentrations are determined with the curve.

Batch adsorption operations were made as described in the following procedure. The amount of solution was taken as 50 mL; the dosages of adsorbent were taken as 0.1 g, 0.3 g, and 0.5 g; and the concentrations of solution were taken as 0.5 mg/L, 1 mg/L, and 1.5 mg/L.

Adsorption experiments were made at 25 °C temperature during 45 min. The removal was calculated in Eq. 1.

$$\text{Removal (\%)} : (C_0 - C) * 100/C_0 \quad (1)$$

In Eq. 1, C_0 and C show the solution concentration before adsorption and after adsorption, respectively.

The adsorption behavior of the beads was investigated by adsorption isotherms and adsorption kinetic models. The adsorption isotherm models were investigated to approve the adsorption operation. Due to popularity, Langmuir and Freundlich models were studied in this study. Langmuir isotherm model equation is given in Eq. 2.

$$1/q_e = 1/q_m + 1/(bq_m C_e) \quad (2)$$

In Eq. 2, q_e is the adsorbed amount per adsorbent (mg/g), C_e is the equilibrium concentration of solution (mg/L), q_m is the maximum adsorbed amount (mg/g), and k_L and b are the constant. The constants are determined from slope and intercept of the $1/q_e$ versus $1/C_e$ graph. A dimensionless parameter k_L is given in Eq. 3 also.

$$k_L = 1/(1 + bC_0) \quad (3)$$

C_0 is the initial concentration of solution (mg/L) in Eq. 3. Freundlich isotherm model equation is given in Eq. 4.

$$\log q_e = \log k_F + (1/n) \log C_e \quad (4)$$

The terms in Eq. 4 are as follows: q_e is the adsorbed amount per adsorbent (mg/g), C_e is the equilibrium concentration of solution (mg/L), and k_F and n are the constants. The constants are determined from slope and intercept of the $\log q_e$ versus $\log C_e$ graph. At equilibrium condition, q_e (mg/g) is given in Eq. 5 (Tümör, 2018, Ilgen & Dulger, 2016).

$$q_e : (C_0 - C_e)(v/m) \quad (5)$$

C_0 is the initial concentration of solution (mg/L), C_e is the equilibrium concentration of solution (mg/L), v is the solution volume (l), and m is the adsorbent amount (gr) in Eq. 5. Adsorption mechanism is explained by kinetic model equations. Lagergren pseudo-first-order and pseudo-second-order models were investigated for the adsorption operation in this study. Lagergren pseudo-first-order model equation is as seen in Eq. 6.

$$\log(q_e - q_t) = \log(q_e) - (k_1/2.3030)t \quad (6)$$

In Eq. 6, q_e and q_t are adsorption capacity at equilibrium (mg/g) and adsorption capacity at any t time (mg/g), respectively. k_1 is a pseudo-first-order model constant (h^{-1}). Theoretical q_e and k_1 are obtained from the graph $\log(q_e - q_t)$ versus t as intercept and slope, respectively. Lagergren pseudo-second-order model equation is as seen in Eq. 7.

$$t/q_t = 1/(k_2 q_e^2) + t/q_e \quad (7)$$

In Eq. 7, q_e and q_t are adsorption capacity at equilibrium (mg/g), and adsorption capacity at any t time (mg/g), respectively. k_2 is a pseudo-second-order model constant (h^{-1}). Theoretical q_e and k_2 are obtained from the graph t/q_t versus t as slope and intercept, respectively (Ilgen, 2014; Tümör, 2018).

Optimization of an experimental process can be made by the response surface method. The response surface method (RSM) is a mathematical and statistical technique. Response surface methodology has been successfully used for many years (Aydar, 2018).

The RSM is a statistical method. Modeling is done by analyzing the response to independent variables that are called factors; the answers to different variables are determined by the model equation that is obtained by the method. It is ensured that the results are predicted without further experimentation. Therefore, time and process cost are decreased by optimization (Hilmioglu, 2022). The Box-Behnken design model is one method of the response surface method (RSM) and used widely (Hilmioglu et al., 2021). In this study, experimental data was examined by the Box-Behnken method.

The response surface method shows what the results would be under operation conditions where experimental work cannot be performed. The variables affecting the removal rate can be found with the RSM. In this study, variables as factors are concentration, adsorbent amount, and time, while the result as response is removal rate. The variable that most affects the removal is determined.

3 Results and Discussion

3.1 Characterization of Chitosan Beads

The photo of the prepared chitosan beads is given in Fig. 1.



Fig. 1 The images of chitosan beads

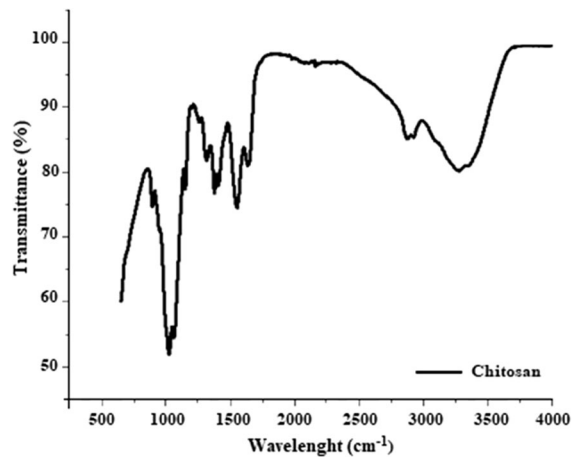


Fig. 2 FT-IR analysis for the chitosan beads

FT-IR and TGA analyses are given in Figs. 2 and 3.

Figure 2 shows the result of FT-IR analysis for the chitosan beads. The chitosan polymer characteristically contains hydroxyl, carbonyl, and amine groups. The peak in the 1150–1040 bands refers to the $-C-O-C-$ group. The peaks in the 1560–1590 bands indicate the presence of $NH-$ and NH_2 , respectively. The peaks in the 3400–3500 and 1415–1320 bands indicate the presence of $-OH$ and $-CH$, respectively. The peak around 1750 indicates the presence of the carbonyl group. The result of this analysis confirms the characteristic features of the chitosan when compared with the literature and proves that the chitosan

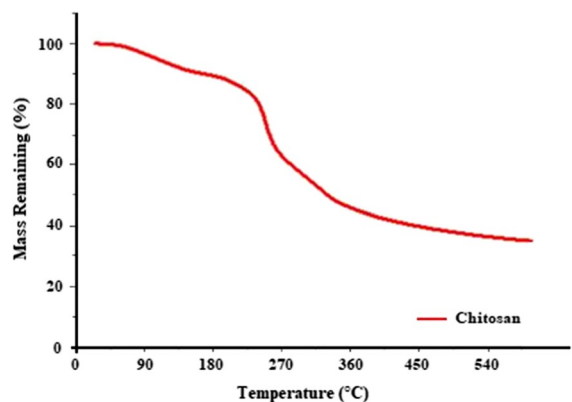


Fig. 3 TGA analysis for the chitosan beads

adsorbent did not undergo any agglomeration during the preparation stage.

Figure 3 shows the result of TGA analysis for the chitosan beads. Thermogravimetric analysis of the chitosan bead was performed with a Mettler Toledo brand device at room temperature to 600 °C and a heating rate of 10 °C/min. As a result of the analysis, it was seen that the first mass loss started at 98–80 °C. The fact that the mass loss started at such a high temperature is a very good result for the phosphate removal experiment. The usability of the prepared adsorbent is suitable for this experiment by concluding that no problems will be encountered for the chitosan beads used at the temperatures where phosphate removal occurs.

3.2 Adsorption Tests

Because of the electrostatic interaction between phosphate and chitosan, it is emphasized that adsorption at acidic pH is more suitable. It is stated that the adsorption efficiency decreased at alkaline pH (Banu et al., 2020, Liu et al., 2020, Eltaweil et al., 2021). High phosphate removal value was obtained with cross-linked chitosan at pH 4 (Filipkowska et al., 2014).

During phosphate adsorption, protonated amino groups are electrostatically attracted to negatively charged phosphate ions (Zhang et al., 2018). pH is an important factor that directly affects the form of adsorption. At pH 5, the amino groups contained in chitosan are protonated in acidic solutions, and the adsorbent becomes positively charged. Thus, electrostatic attraction forms between positively charged chitosan and phosphate anions. The adsorption of phosphate becomes easier. As the pH increases, the adsorbent surface becomes negatively charged. This causes a reduction in adsorption due to the repulsion between the adsorbent charged negatively and negatively charged phosphate anions (Rajeswari et al., 2015).

Adsorption tests were made at different pH values such as pH 2–11, and the highest removal was obtained at 4.95, the pH of the solution itself. Therefore, tests were carried out at the solution's own pH value of 4.95.

Figure 4 shows the relationship between time and removal. Adsorption was made for 3 h at 25 °C with 1 mg/L phosphate solution by using 0.5 g adsorbent. The highest removal rate was reached in 45 min.

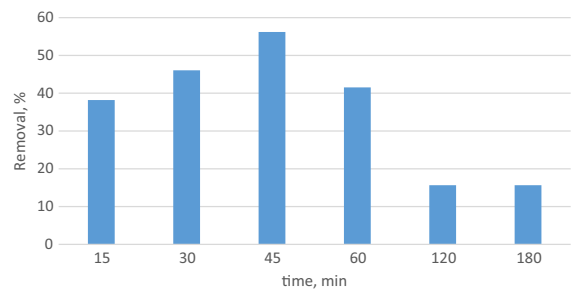


Fig. 4 Removal versus time

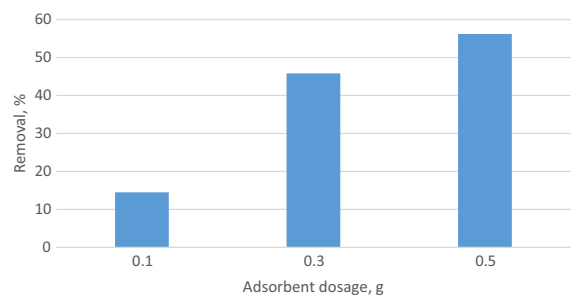


Fig. 5 Removal versus adsorbent dosage

After 45 min, removal decreased because the beads may have lowered their adsorption ability. Therefore, important removal was not observed in the 120th and 180th min. As a result, adsorption time was determined as 45 min.

Adsorbent dosage effect on removal was investigated. Adsorption was carried out for 45 min at 25 °C with 1 mg/L phosphate solution by using 0.1 g, 0.3 g, and 0.5 g adsorbent. The highest removal rate was reached by using 0.5 g adsorbent dosage. Figure 5 shows the relationship between adsorbent dosage and removal. As the amount of adsorbent increases, the removal rate also increases.

Since the accessibility to the empty areas on the adsorbent surface increased, the removal ratio increased as the adsorbent amount increased (Banu et al., 2020). To see the solution concentration effect on removal, adsorption was carried out for 45 min at 25 °C with 0.5 mg/L, 1 mg/L, and 1.5 mg/L phosphate solution by using 0.5 g adsorbent. The highest removal rate was reached with 1.5 mg/L solution concentration. Figure 6 shows the relationship between solution concentration and removal. Removal values obtained for 1 mg/L and 1.5 mg/L phosphate solution

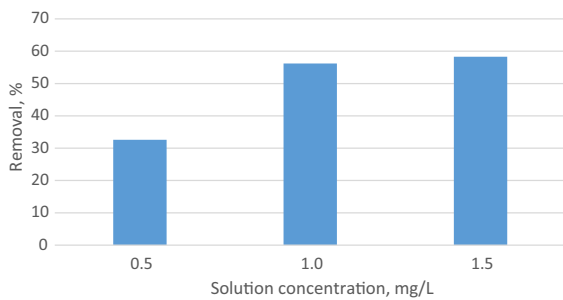


Fig. 6 Removal versus solution concentration

concentrations were approximate values. The removal increased with increasing solution concentration. In order to achieve high efficiency at low concentrations by chitosan, the adsorbent dosage should be increased. Increasing the amount of chitosan in the adsorbent beads is another option.

Chitosan is a nitrogen-containing biosorbent. It can bind and sorb anionic pollutants by means of various processes like hydrogen bonding, ion exchange, Van der Waals forces, and molecular interactions. These mechanisms increase the adsorption affinity and ability of chitosan for phosphate. The free amine groups in chitosan cause higher adsorption capacity. There is great interest in the positively charged adsorbent surface to adsorb negatively charged phosphate ions. Therefore, chitosan was chosen as adsorbent material for phosphate removal (Filipkowska et al., 2014; Liu et al., 2020). The removal mechanism between adsorbent chitosan and polluting phosphate ions is electrostatic attraction of phosphate ions with chitosan's amine groups.

3.2.1 Comparison of Results with Literature

In this study, the maximum adsorption capacity of chitosan beads is 87.26 mg/g. In 45 min, 56.2% removal was obtained for 1 mg/L phosphate solution using 0.5 g adsorbent.

Mohammadi et al. used biochar obtained from *Rosmarinus officinalis* leaves for phosphate removal from wastewater. They investigated the adsorption capacity depending on the phosphate concentration and the effect of PH. They obtained the maximum adsorption capacity of 78.24 mg/g and 67% removal ratio.

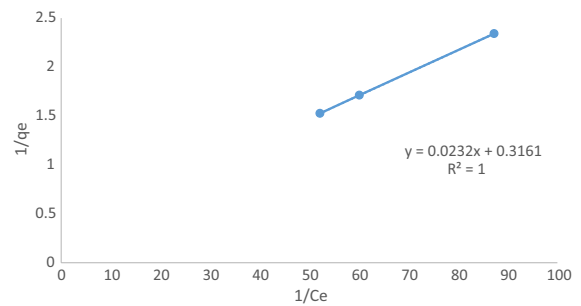


Fig. 7 Langmuir isotherm

Adsorption is pH dependent and increased in acidic environments (Mohammadi et al., 2021). Atnafu and Leta produced Fe_3O_4 clay polymer nanocomposites based on plasticized magnetic starch and used them for phosphate removal from aqueous solutions. The effects of solution pH, adsorbent concentration, and amount of adsorbent on phosphate removal were investigated. They achieved 45–95% removal efficiency at pH 6. They observed that phosphate adsorption increased when the pH was increased from 3 to 7. They observed that adsorption increased as the amount of adsorbent increased (Atnafu & Leta, 2021). Zhang et al. produced La/Ca-loaded adsorbents from modified wheat straw for phosphate removal. They obtained 84.72 mg/g adsorption capacity for phosphate removal and 85% removal ratio. They obtained a good adsorption capacity between pH 3 and 10 (Zhang et al., 2023).

3.3 Isotherms and Kinetic Models

Langmuir isotherm is given in Fig. 7.

From Langmuir isotherm, k_L was estimated as 0.0683. If this value is between 0 and 1, the adsorption isotherm suits the Langmuir isotherm model. b was estimated as 13.627 also.

Freundlich isotherm is given in Fig. 8.

From Freundlich isotherm, n value was found as 1.248. Because n is greater than one, the adsorption isotherm also suits the Freundlich isotherm model. k_F value was found as 0.0643. In Table 1 isotherm constants are seen.

Although the data seems to be suitable for both isotherms, it is understood that experimental values

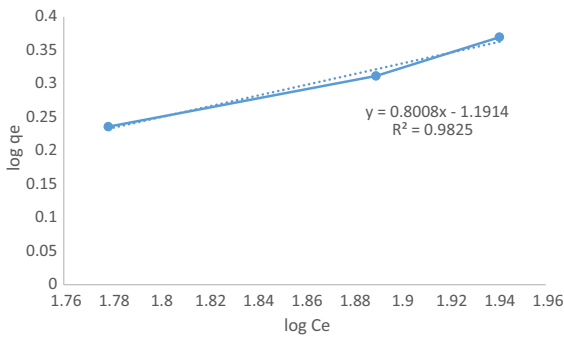


Fig. 8 Freundlich isotherm

Table 1 Isotherm constants

Langmuir			Freundlich		
k_L	b	R^2	k_F	n	R^2
0.0683	13.627	1	0.0643	1.248	0.9825

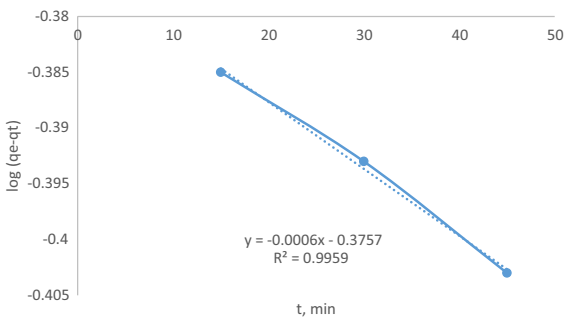


Fig. 9 Pseudo-first-order kinetic

are compatible with the Langmuir isotherm since the R^2 value is the bigger one. The pseudo-first-order kinetic is given in Fig. 9.

The calculated q_e value is close to the experimental q_e value, and the R^2 value is close to one. This situation indicates the suitability of the adsorption kinetics to the pseudo-first-order kinetics. The pseudo-second-order kinetic is given in Fig. 10.

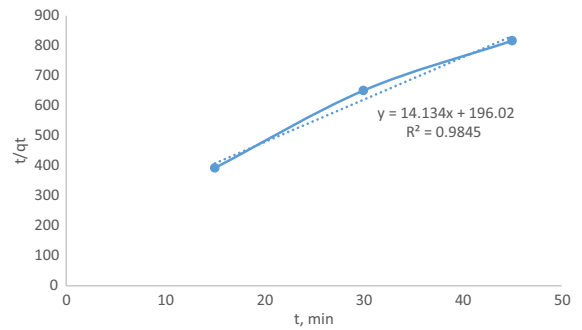


Fig. 10 Pseudo-second-order kinetic

Though the R^2 value is close to one, the calculated q_e value is not close to the experimental q_e value. This situation indicates that the adsorption kinetics are not proper to the pseudo-second-order kinetics. In Table 2, kinetic parameters are given.

The experimental and calculated q_e values were close to each other, and R^2 was bigger; these situations indicate that the data were suitable with the pseudo-first-order kinetic model.

3.4 Statistical Analysis Results

In this study, the factors were adsorbent dosage (A), solution concentration (B), and time (C). Response was adsorption removal (%). The Design Expert Program that used the Box-Behnken method suggested 17 experiments for the selected factors. Factors and responses obtained by adsorption experiments were entered to the program.

The model suggested by the program is given in Table 3. Linear model has the smallest p value and the biggest sum of squares. Therefore, it was selected for the following procedure.

In Table 3, F value of the model should be bigger than 4. In this respect, the linear model is a more suitable model. ANOVA analysis of the linear model is given in Table 4.

The factor with the highest F value has the highest effect on the response. The values of F , p , and sum of squares (SS) point to the significance of the model. When the SS rises, the significance of that factor rises. Because the adsorbent dosage has

Table 2 Kinetic parameters

q_e experimental	Pseudo-first-order			Pseudo-second-order		
	k_1	q_e theoretical	R^2	k_2	q_e theoretical	R^2
0.45	1.38×10^{-3}	0.421	0.9959	1.021	0.0707	0.9845

Table 3 Fitting models

Source	Sum of squares, SS	Degree of medium, df	Mean square	F value	p value	
Mean vs total	13,060.56	1	13,060.56			
Linear vs mean	3159.65	3	1053.22	49.92	< 0.0001	Suggested
2FI vs linear	120.08	3	40.03	2.60	0.1105	
Quadratic vs 2FI	104.02	3	34.67	4.84	0.0395	Suggested
Cubic vs quadratic	50.18	3	16.73			Aliased
Residual	0.0000	4	0.0000			
Total	16,494.48	17	970.26			

Values in bold indicate linear vs Mean and Quadratic vs 2FI

Table 4 ANOVA variance analysis for linear model

Source	Sum of squares, SS	Degree of medium, df	Mean square	F value	p value	
Model	3159.65	3	1053.22	49.92	< 0.0001	Significant
A—Adsorbent dosage (g)	1940.65	1	1940.65	91.98	< 0.0001	
B—Solution concentration (mg/L)	905.25	1	905.25	42.91	< 0.0001	
C—Time (minute)	313.75	1	313.75	14.87	0.0020	
Residual	274.28	13	21.10			
Lack of fit	274.28	9	30.48			
Pure error	0.0000	4	0.0000			
Cor total	3433.92	16				

Table 5 Fit statistics

Standard deviation	4.59	R^2	0.9201
Mean	27.72	Adjusted R^2	0.9017
C.V. %	16.57	Predicted R^2	0.8422
		Adequate precision	23.5293

the highest SS and F values, it is understood that the adsorbent dosage shows the biggest effect on removal of phosphate by adsorption using chitosan beads prepared. It is necessary for the value of p in the table to be less than 0.05 for model compliance. The suitability of the value of p indicates

that compatibility of the experimental data and the model data is at 95% confidence interval. It has been seen that the data obtained from linear models are 95% reliable. Table 4 also shows the probability values for each factor in the model. Factors with p values bigger than 0.10 are not suitable for the model. It can be said that all the factors adsorbent dosage, solution concentration, and time are favorable for adsorption operation. Fit statistics are given in Table 5.

The experimental results are evaluated with model results by the p value proposed by the program, the model regression coefficient R^2 , adequate precision, adjusted regression coefficient

R^2 , and predicted regression coefficient R^2 values. The cases that the p value is under than 0.05, the adequate precision that is representing the signal to noise ratio value is bigger than 4, and the difference there among the adjusted regression coefficient and the predicted regression coefficient is under than 0.2 indicate the fit of the model. The closer the R^2 value to 1, the more suitable the model. When Table 5 is evaluated, it is understood that the R^2 (0.9201) is closer to 1, the difference of adjusted R^2 (0.9017) and predicted R^2 (0.8422) is less than 0.2, and the adequate precision (23.5293) is bigger than 4. In this situation, the linear model proposed by the program can be said to be a quite proper model. The equation of the linear model proposed by the model is given in Eq. 8. The effects of all factors are seen in the equation.

$$\text{Removal}(\%) = 27.72 + 15.57 * A + 10.64 * B + 6.26 * C \tag{8}$$

The plot of diagnostics of predicted vs actual responses is given in Fig. 11.

The data values should not have large deviations for comparison of actual and predicted values. If the data is close to the black linear line, the actual and predicted values are close to each other. In Fig. 11, there is no excessive deviation. This indicates that the program presented a compatible model with experimental data. The plot of Cook’s distance versus run number is given in Fig. 12. The Cook’s distance value versus run number should be below 1. The Cook’s distance value shown in Fig. 12 is lower than 1. It is stated that the data have anticipated Cook’s distance range.

A 2D contour graph and a 3D surface graph can be drawn in the linear model. The plot in Fig. 13

Fig. 11 Diagnostic plots of predicted vs actual responses

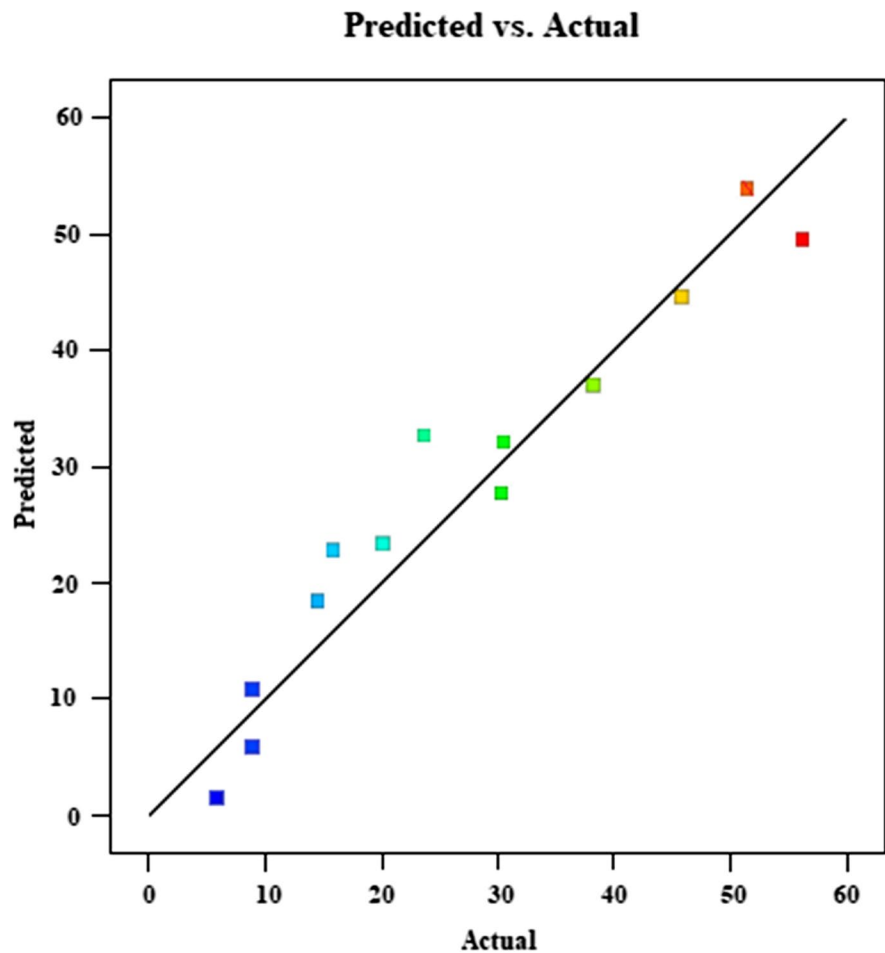
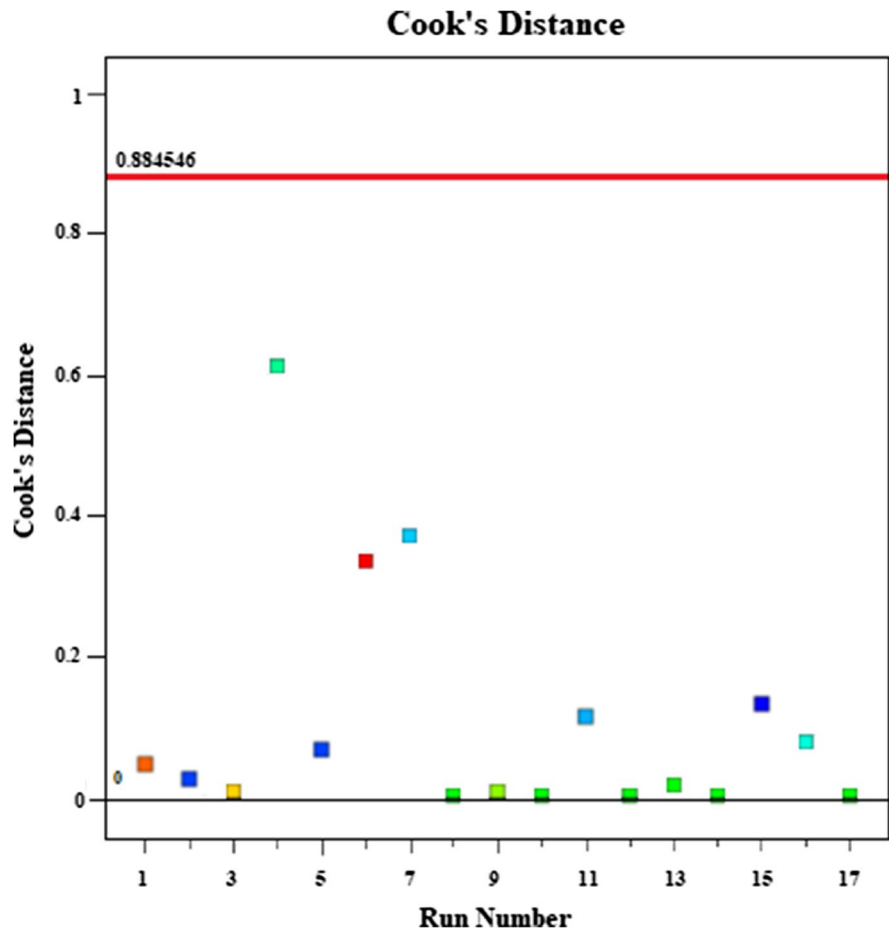


Fig. 12 Plot of Cook's distance vs runs



is the linear model contour graphic. The contour chart makes toning gradually. The relationship between adsorbent dosage and solution concentration on removal ratio were plotted. A two-dimensional surface plot is a two-dimensional contour plot of the two factors on responses. According to contour plot, as adsorbent dosage and solution concentration increase, removal in the adsorption increases also.

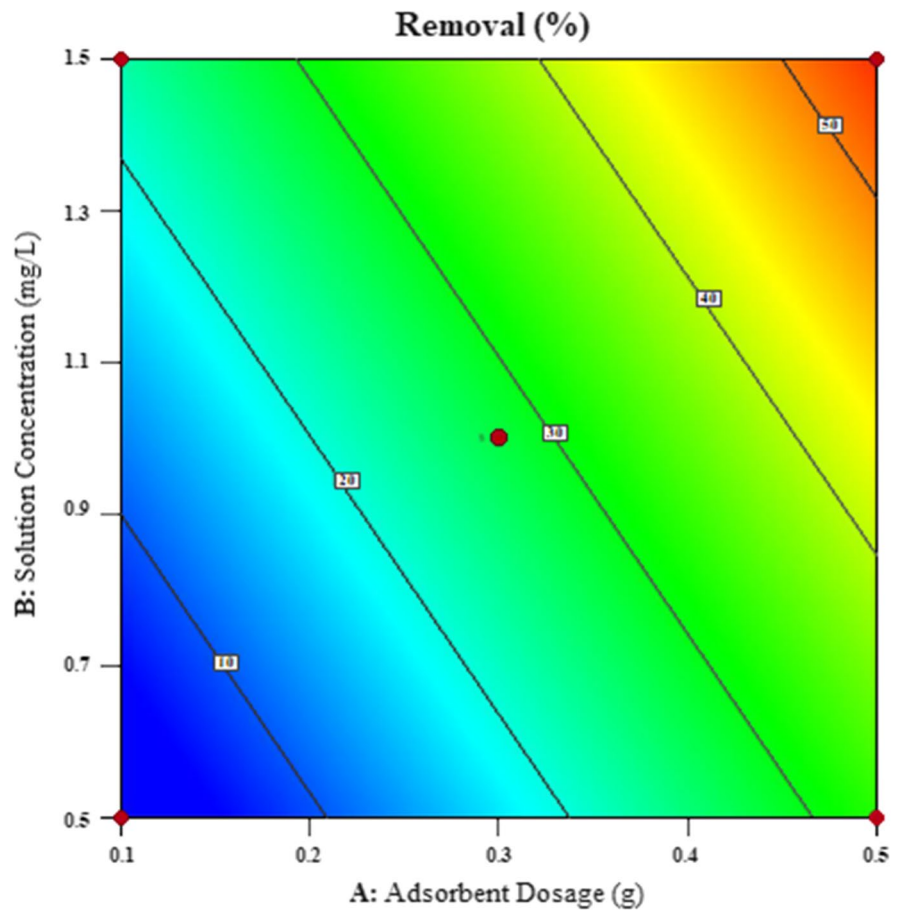
A 3-dimensional surface plot is a three-dimensional plot of the two factors on responses, while one factor is constant. A three-dimensional surface plot shows relation between factors. The dependence of a factor that is constant on other factors is seen, while the other two factors are changed. In Fig. 14, the change of adsorption removal was

observed by solution concentration and adsorbent dosage factors when the time factor was kept constant. Adsorbent dosage and solution concentration have the same effect on removal in 3D plot also as seen 2D plot.

It is seen that from Figs. 13 to. 14, the adsorption removal is enhanced by the solution concentration and adsorbent dose enhancement. A report of the model is presented in Table 6. In this report, the experimental data and the model data obtained in the program are compared. The values deviated from Cook's distance and DFFITS values are shown in color. Cook's distance and DFFITS values should be less than 1.

When DFFITS (difference in fits) are looked at, there is only one value in colorful in Table 6. That

Fig. 13 2D contour plot of adsorbent dosage and solution concentration on removal



means it exceeded the limit; it can be ignored because it could not put at risk the fit of the model. Cook's distance values are also below 1. It is seen that from the report table, the model is valid.

4 Conclusion

The phosphate removal from water by adsorption was done by chitosan beads prepared. Acid concentration and adsorbent dosage had effects on the removal of the phosphate.

The experimental data were suitable with the Langmuir isotherm model and pseudo-first-order kinetic model. The modeling was made with the Box-Behnken method; modeling has shown that

the linear model is the compatible model for the experimental data.

In this study, 56.2% highest removal was obtained for 1 mg/L phosphate solution in 45 min when 0.5 g adsorbent was used. According to the results obtained with RSM, the highest removal of phosphate is obtained as 88.49% for phosphate solution concentration is 1 mg/L and adsorbent dosage is 1 g, when the time is 45 min. It is determined that by optimization, the phosphate removal value would increase by 57.45% for the same phosphate concentration that is 1 mg/L at the same time that is 45 min, when the adsorbent dosage is increased by 100% from 0.5 to 1 g.

Fig. 14 3D surface plot of adsorbent dosage and solution concentration on removal

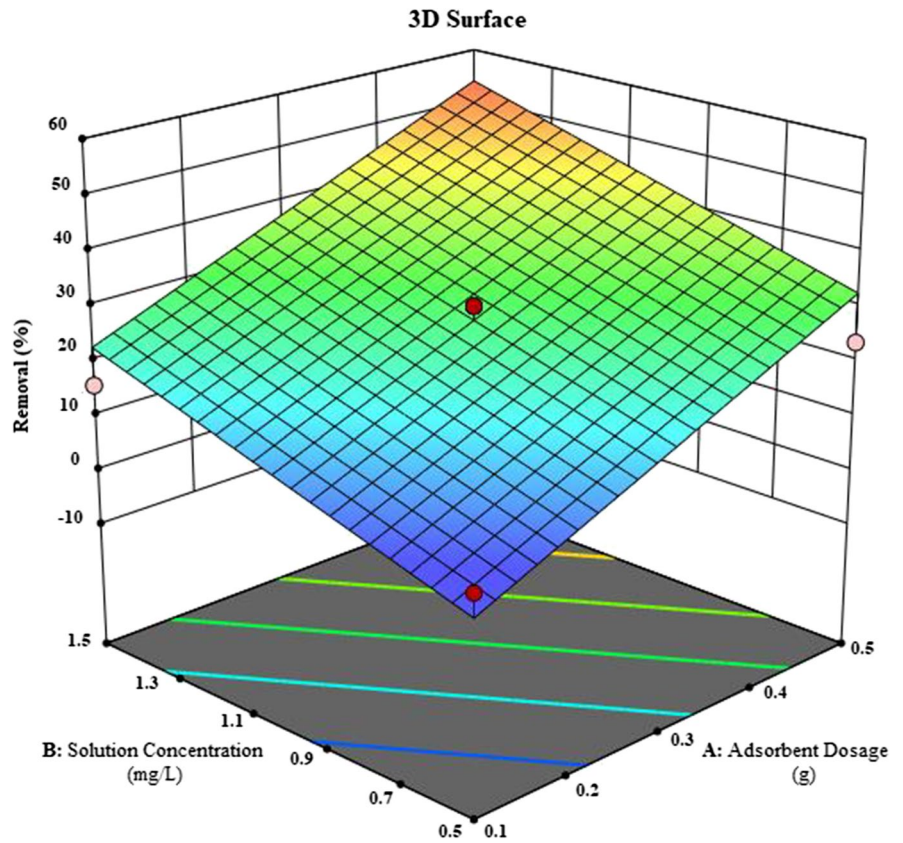


Table 6 Report on linear model

Run order	Actual value	Predicted value	Residual	Leverage	Internally studentized residuals	Externally studentized residuals	Cook's distance	Influence on fitted value DFFITS
1	51.40	53.93	-2.53	0.309	-0.663	-0.648	0.049	-0.433
2	8.90	10.82	-1.92	0.309	-0.502	-0.487	0.028	-0.326
3	45.80	44.62	1.18	0.309	0.310	0.299	0.011	0.200
4	23.70	32.66	-8.96	0.309	-2.345	-2.966	0.614	-1.983⁽¹⁾
5	8.90	5.88	3.02	0.309	0.791	0.779	0.070	0.521
6	56.20	49.56	6.64	0.309	1.740	1.909	0.338	1.276
7	15.80	22.78	-6.98	0.309	-1.828	-2.037	0.373	-1.362
8	30.30	27.72	2.58	0.059	0.580	0.564	0.005	0.141
9	38.20	37.03	1.17	0.309	0.306	0.295	0.010	0.197
10	30.30	27.72	2.58	0.059	0.580	0.564	0.005	0.141
11	14.50	18.41	-3.91	0.309	-1.023	-1.025	0.117	-0.685
12	30.30	27.72	2.58	0.059	0.580	0.564	0.005	0.141
13	30.50	32.09	-1.59	0.309	-0.417	-0.403	0.019	-0.270
14	30.30	27.72	2.58	0.059	0.580	0.564	0.005	0.141
15	5.70	1.51	4.19	0.309	1.098	1.108	0.135	0.741
16	20.10	23.34	-3.24	0.309	-0.849	-0.839	0.081	-0.561
17	30.30	27.72	2.58	0.059	0.580	0.564	0.005	0.141

Data Availability The datasets used and/or analyzed during the current study are available from the corresponding author upon reasonable request.

Declarations

Ethics Declarations Not applicable.

Conflict of Interest The authors declare no competing interests.

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