




# Decomposition and temperature sensitivity of fine root and leaf litter of 43 mediterranean species

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## Abstract

**Aims** Data on the decomposition of fine roots are scarce for the Mediterranean basin. This work aims to compare chemical traits, decomposition rate, and temperature sensitivity ( $Q_{10}$ ) for root and leaf litter of 43 Mediterranean species.

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**Methods** We carried out a two-years litterbag decomposition experiment using 43 fine roots litter and leaf litter types incubated in laboratory conditions at three different temperatures, i.e. 4 °C, 14 °C, and 24 °C. Litter was characterized for carbon (C), nitrogen (N), lignin and cellulose concentration, C/N, and lignin/N ratios.

**Results** Fine root litter had lower N content but higher lignin concentration, lignin/N, and C/N ratios compared to leaf litter. The decay rate of fine root litter was slower than leaf litter. For both tissues, the decay rate was negatively associated with lignin concentration, lignin/N, and C/N ratios but positively with N concentration.  $Q_{10}$  was higher for fine root than leaf litter, with a positive correlation with lignin while negative with N concentration.

**Conclusions** Our findings demonstrate a higher  $Q_{10}$  accompanied by a slower decomposition rate of fine root litter compared to leaf litter in Mediterranean ecosystems. These results must be considered in modeling organic C at the ecosystem scale.

**Keywords**  $Q_{10}$  · Chemical traits · C/N ratio · Lignin/N ratio · Nitrogen limitation.

## Introduction

Litter decomposition is a key process for biogeochemical cycles in terrestrial ecosystems. Decomposition controls the build-up of soil organic carbon (C) stock and the dynamics of major nutrients (Berg and McLaugherty 2014), thus affecting the plant diversity

and productivity along with associated microbial communities (Purahong et al. 2016). At both global and regional scales, climate, i.e. temperature and moisture, is the main factor controlling the rate of litter decomposition (Aerts 1997). However, at a finer ecosystem scale, the chemical traits of decomposing plant tissue become more important compared to climatic factors (Cornwell et al. 2008).

In general terms, the relationships between litter decomposition and temperature are fairly well established in the literature (e.g. Meentemeyer 1978; Hobbie 1996), with decomposition rates increasing in parallel with temperatures. However, the magnitude of temperature sensitivity of decomposing litter is still a debated issue, with contrasting results about the relationship between sensitivity and substrate recalcitrance (Hartley and Ineson 2008; Giardina and Ryan 2000). Theoretical and empirical evidence suggests that temperature sensitivity increases with substrate recalcitrance (Davidson and Janssens 2006; Hartley and Ineson 2008) i.e. litter with slow mass-loss is expected to increase its decay rate with rising temperature more compared to an already rapidly decomposing plant tissue. In this context, no studies have explicitly investigated the effect of temperature on mass loss of fine roots.

Fine roots (diameter < 2 mm) represent a substantial fraction of ecosystem C pool and, because of their rapid turnover, account for about 14–27 % of net primary productivity (McCormack et al. 2015). Freschet et al. (2013) estimated that fine root decomposition contributes to ~33 % of whole litter inputs in forest ecosystems and ~48 % in grasslands. However, fine root usually decomposes at a slower rate than leaves and, thus, provides a disproportionate contribution to soil C storage compared to leaf litter (Clemmensen et al. 2013; Sun et al. 2018). The slow decay rate of fine root litter can be related to its relatively low nitrogen (N) content coupled with larger lignin content compared to leaf litter (Zhang et al. 2008). N availability may limit litter decomposition when the C/N ratio lies above the threshold of ~30–35 (Taylor et al. 1989). The very low initial N content of some plant tissues like wood debris and coarse roots (Bradford et al. 2014) can, indeed, induce a microbial N limitation that slows down the decomposition rate. In this regard, the first global analysis of species-specific fine root decomposition rates (Silver and Miya 2001) recorded a strong negative effect of the C/N ratio and a positive effect of calcium (Ca). A

recent meta-analysis (See et al. 2019) confirmed the positive effects of N concentrations and the negative effects of lignin concentration on decomposition rates of fine root litters, while the positive effect of Ca was not confirmed. In addition to chemical traits, the type of mycorrhizal association, i.e. ectomycorrhizal, ericoid or arbuscular (AM), may affect decomposition but their effect has not yet been fully explored (Beidler and Pritchard 2017). For instance, the presence of the ectomycorrhizal fungal mantle could contribute to the slow decomposition of woody roots (Langley et al. 2006; Sun et al. 2018).

Although the number of published studies concerning fine root decomposition has substantially increased over the last two decades (See et al. 2019), fine root decay is still poorly understood compared to leaf litter. Systematic studies comparing leaf and root decomposition have been carried out in several ecosystems, including temperate forests (Hobbie et al. 2010; Freschet et al. 2012; Jo et al. 2016; Sun et al. 2018), tropical forest (Ostertag and Hobbie 1999; Cusack et al. 2009; Wang et al. 2010), and temperate grasslands (Vivanco and Austin 2006). For Mediterranean ecosystems, however, the study of Birouste et al. (2012) was the only one that compared the decomposition rate of fine root and leaf for 18 herbaceous species. Nonetheless, no studies are available for Mediterranean shrubs and trees. In this context, this work aimed to compare chemical traits, decomposition rate, and temperature sensitivity ( $Q_{10}$ ) for root and leaf litter of 43 Mediterranean species. We carried out a two-years litterbag decomposition experiment, in which the temperature sensitivity was assessed by incubating litterbags at three temperatures, i.e. 4 °C, 14 °C, and 24 °C. Leaf litter and fine roots were characterized for carbon, nitrogen (N), lignin, cellulose content, C/N, and lignin/N ratios. The specific targets of this work were: (i) to describe leaf litter and fine root features based on chemical traits, (ii) to assess the decomposition rate of 43 leaf litter and fine root types, (iii) to explore the relationships between litter chemical traits and the decay rate, and, (iv) to assess temperature sensitivity of leaf and fine root litter and their link along with tissue chemical traits. Based on the above considerations, three main hypotheses were tested: (1) decay rate of the fine root litter would be slower compared to leaf litter; (2) decay rate would be negatively correlated with lignin content and positively correlated with N concentration; (3) temperature sensitivity would be higher for fine root than leaf litter.

## Materials and methods

### Leaf litter and fine root collection

Forty-three species common in the Mediterranean basin were selected belonging to different growth forms (Supplementary Table S1). The species list includes native and invasive plants with nine deciduous woody (*Acer cappadocicum* subsp. *lobelii*, *Broussonetia papyrifera*, *Castanea sativa*, *Celtis australis*, *Fagus sylvatica*, *Fraxinus ornus*, *Ostrya carpinifolia*, *Populus nigra* and *Salix purpurea*), four evergreen woody (*Laurus nobilis*, *Myrtus communis*, *Pistacia lentiscus* and *Quercus ilex*), two N-fixing woody (*Emerus major* and *Spartium junceum*), nineteen forbs (*Acanthus mollis*, *Alcea rosea*, *Amaranthus retroflexus*, *Arum italicum*, *Chenopodium album*, *Conyza canadensis*, *Dittrichia viscosa*, *Foeniculum vulgare*, *Medicago sativa*, *Oxalis pes-caprae*, *Papaver rhoeas*, *Parietaria judaica*, *Plantago lanceolata*, *Smyrniolum olusatrum*, *Solanum nigrum*, *Tussilago farfara*, *Typha latifolia*, *Urtica dioica* and *Xanthium italicum*), four grass (*Arundo donax*, *Hyparrhenia hirta*, *Heteropogon contortus* and *Sorghum halepense*), three sedges (*Carex pendula*, *Juncus effusus* and *Scirpus holoschoenus*), and two vines (*Hedera helix* and *Rubus ulmifolius*). The selected species, representing 28 taxonomic families, are expected to cover a wide range of leaf litter and fine root chemical traits.

Plant litter materials were sampled across grasslands, shrublands, and woodlands located in the Campania Region, Southern Italy. Freshly abscised leaves of each species were collected from plants ( $N=10$ ) randomly selected at the sampling sites. Fine roots with a diameter  $< 2$  mm were collected from the top of the soil (20 cm). With this approach, we obtained a mixture of first- to third-order roots, allowing the exclusion of secondary structural roots (Jo et al. 2016). Herbaceous individuals were completely excavated using a trowel while, for woody species, roots were excavated and progressively tracked until the base of the plants to be certain of the species identity. Roots samples were washed with distilled water to remove all soil particles. After collection, leaf litter and fine root were air-dried in a ventilated chamber at 20 °C for 30 days until reaching a constant weight and then stored at room temperature in paper bags.

### Decomposition experiment

Plant tissue decomposition in field conditions is hierarchically controlled by litter chemistry and environmental conditions, i.e. water availability and temperature. In this study, we decomposed leaf and root under laboratory conditions to focus on the importance of litter chemistry on decay rate and reduce the influence of local climatic factors, e.g. fluctuation in water availability and temperature. To assess the relationship between litter chemistry and temperature sensitivity, leaves and fine roots were decomposed at three temperatures, i.e. 4 °C, 14 °C, and 24 °C that represent winter, spring/autumn, and summer conditions in many places along the Mediterranean basin.

The decomposition experiment was carried out accordingly to the litterbag method (Berg and McLaugherty 2014). Litterbags (10 × 10 cm, mesh size 1 mm) were filled with 2 g of leaf litter and fine root for every 43 species. Litterbags were placed, under laboratory conditions, over trays filled with soil collected from the site where most of the selected species naturally coexist (Cicerale 40°19' N 15°07' E, 186 m a.s.l.). The soil is sandy loam and its specific properties are reported in Supplementary Table S2.

The microcosms were placed in a growth chamber at three temperatures, i.e. 4 °C, 14 °C, and 24 °C. Microcosms were watered with distilled water every three days to maintain the soil at field capacity and avoid any drought effect on decomposition. Soil field capacity was previously determined equal to 18 % in volume for Cicerale soil using the methods reported in Bonanomi et al. (2017a). The entire experimental design included 86 litter types (43 leaf litters and 43 fine root litters), incubated at three temperatures with three replicates for each species. Additionally for each experimental factor six dates of retrieval from the beginning of the experiment (10, 30, 90, 180, 360, and 720 days) were fixed. Overall, 4,644 litterbags were prepared (86 litter types × 3 temperatures × 6 retrieval times × 3 replicates). At each retrieval date, litterbags were collected, oven-dried (30 °C until a constant weight was reached), soil debris carefully removed by brush, and then weighed to the nearest 0.001 g.

### Chemical analyses

Undecomposed leaf litter and fine root were ground to get a fine powder ( $< 1$  mm) and the chemical

determination was carried out in triplicate. Total C and N content were determined by flash combustion of micro-samples (5 mg) by a CN-elemental analyzer (Flash EA2000 Thermo). Cellulose and lignin contents were determined with the method described by Gessner (2005). Briefly, cellulose was quantified as a hydrolyzable fraction after the sulphuric acid digestion (loss after treatment with 72 % H<sub>2</sub>SO<sub>4</sub> for 3 h), while lignin was determined as a nonhydrolyzable fraction (loss upon ignition after the sulphuric acid digestion). All fractions are reported as ash-free dry mass. We highlight that lignin assessed with this method does not correspond to pure lignin, but it includes several other hydrolysis-resistant organic compounds like waxes, suberin, cutin, and condensed tannins (Berg and McLaugherty 2014). C/N and lignin/N ratios for each litter type were calculated.

#### Data analysis

Litter negative exponential decay constant ( $K$ ) was calculated according to Berg and McLaugherty (2014). The model equation was  $M_t = M_0 \cdot e^{-kt}$ , where  $M_0$  is the initial mass,  $M_t$  is the mass remaining after time  $t$  expressed as a proportion of the initial dry mass,  $K$ , the decay constant expressed in units  $d^{-1}$ ; thereafter,  $K$  can be derived for any sample set as the slope of a linear regression of  $\log_e(M_t/M_0)$  versus time (Salinas et al. 2011).

We used general linear models (GLMs) to test main and second-order interactive effects of plant litter types (43 species), temperatures of incubation (4 °C, 14 and 24 °C), and decomposition times (treated as a continuous covariate) on mass remaining percentage. Percentage mass remaining data were log-transformed before applying both GLM and Tukey HSD post-hoc tests. General linear models (GLMs) was also used to assess the effect of plant litter types, temperatures of incubation, and decomposition times on litter  $K$ . Differences in litter chemical traits (C, N, cellulose and lignin content, C/N and lignin/N ratios) and  $K$  were evaluated in all cases by paired t-test ( $P < 0.05$  and  $P < 0.01$ ). To explore the relationship between litter  $K$  with chemical traits, we calculated correlation according to non-parametric Spearman correlation rank methodology to avoid the bias of correlation between variables with non-normal distribution.

The temperature sensitivity of leaf litter and fine root decomposition was estimated by calculating the  $Q_{10}$

value (change in the rate of mass loss given by a 10 °C change in temperature) as follows:

$$Q_{10} = \left( \frac{K_{24^\circ C}}{K_{4^\circ C}} \right)^{\frac{10}{24^\circ C - 4^\circ C}}$$

Where  $K_{24^\circ C}$  and  $K_{4^\circ C}$  are the decay constant calculated with data obtained with litter incubation at 24 and 4 °C, respectively.

Pair-wise differences were statistically evaluated by paired t-test ( $P < 0.05$  and  $P < 0.01$ ). The software STATISTICA 10 (StatSoft Inc., USA) and  $R$  were used for statistical analyses (R Core Team 2017).

## Results

### Chemical traits of leaf litter and fine roots

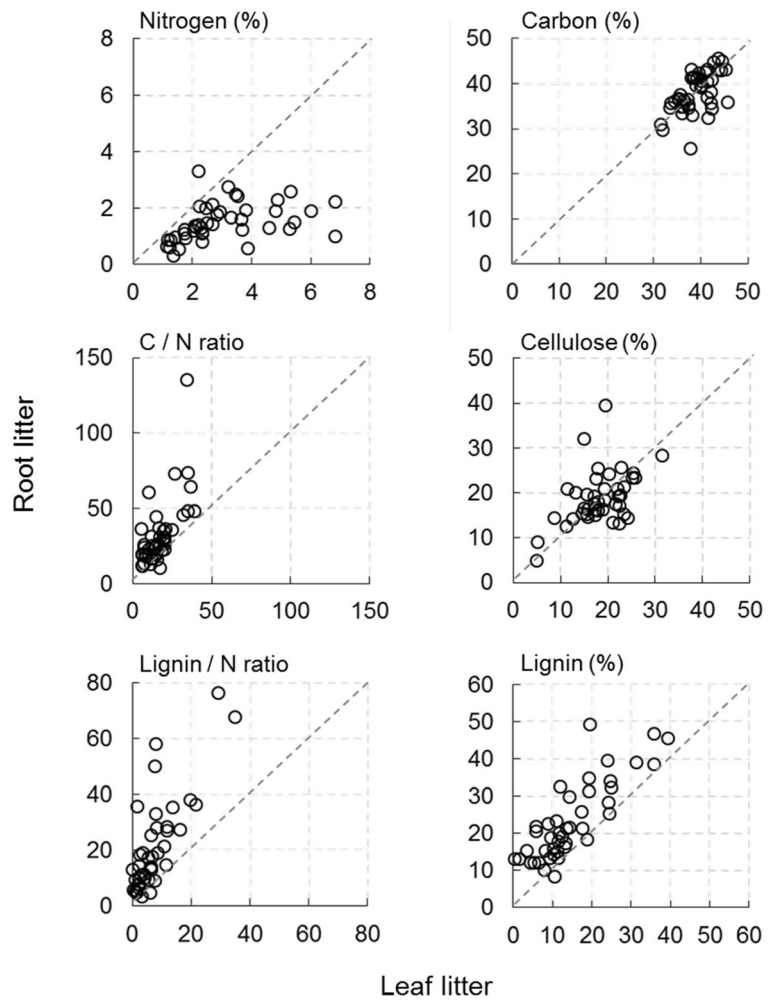
Chemistry of undecomposed leaf litter and fine root largely varied among species (Supplementary Table S3). Leaf litter had lower lignin content, lignin/N, and C/N ratio (paired t-test,  $P < 0.01$  for all cases) compared to root (Fig. 1). On the other hand, roots showed lower N content compared to leaf (paired t-test,  $P < 0.01$ ). No significant difference was recorded concerning the C (paired t-test,  $P = 0.14$ ) and cellulose (paired t-test,  $P = 0.46$ ) content of leaf and root.

C ( $r = +0.59$ ), N ( $r = +0.45$ ), cellulose ( $r = +0.47$ ) and lignin concentration ( $r = +0.82$ ) in leaf litter and the corresponding fine roots were positively correlated. Similarly, the C/N ratio ( $r = +0.67$ ) and lignin/N ratio ( $r = +0.79$ ) of leaf litter and the corresponding fine roots were also correlated.

### Decomposition of leaf litter and fine roots

Leaf and root decay rate ( $K$ ) largely varied among species and was significantly affected by the plant species, the temperature of incubation, and incubation time (Table S4). At low temperature (4 °C),  $K$  ranged from 0.69 (*A. italicum*) to 0.10 (*F. sylvatica*) for leaf and from 0.64 (*A. mollis*) to 0.02 (*A. lobelii*) for root (Supplementary Figure S1). At 14 °C of incubation,  $K$  ranged from 0.92 (*A. retroflexus*) to 0.48 (*L. nobilis*) for leaf and from 0.86 (*A. mollis*) to 0.07 (*S. holoshoenus*) for root (Fig. 2). At 24 °C,  $K$  was less variable among leaf litter with 19 species having a  $K > 0.90$  (Supplementary Figure S2). For root, instead, only three

**Fig. 1** Scatter plot reporting the comparison of chemical traits between leaf litter and fine roots of the 43 Mediterranean species. Raw data of chemical traits for each species are reported in Supplementary Table S3



species had  $K > 0.90$  (*P. lanceolata*, *R. ulmifolius*, and *X. italicum*) with the lowest decay rate recorded for the grass *C. hirtus* and the evergreen tree *Q. ilex* ( $K = 0.19$  in both cases).

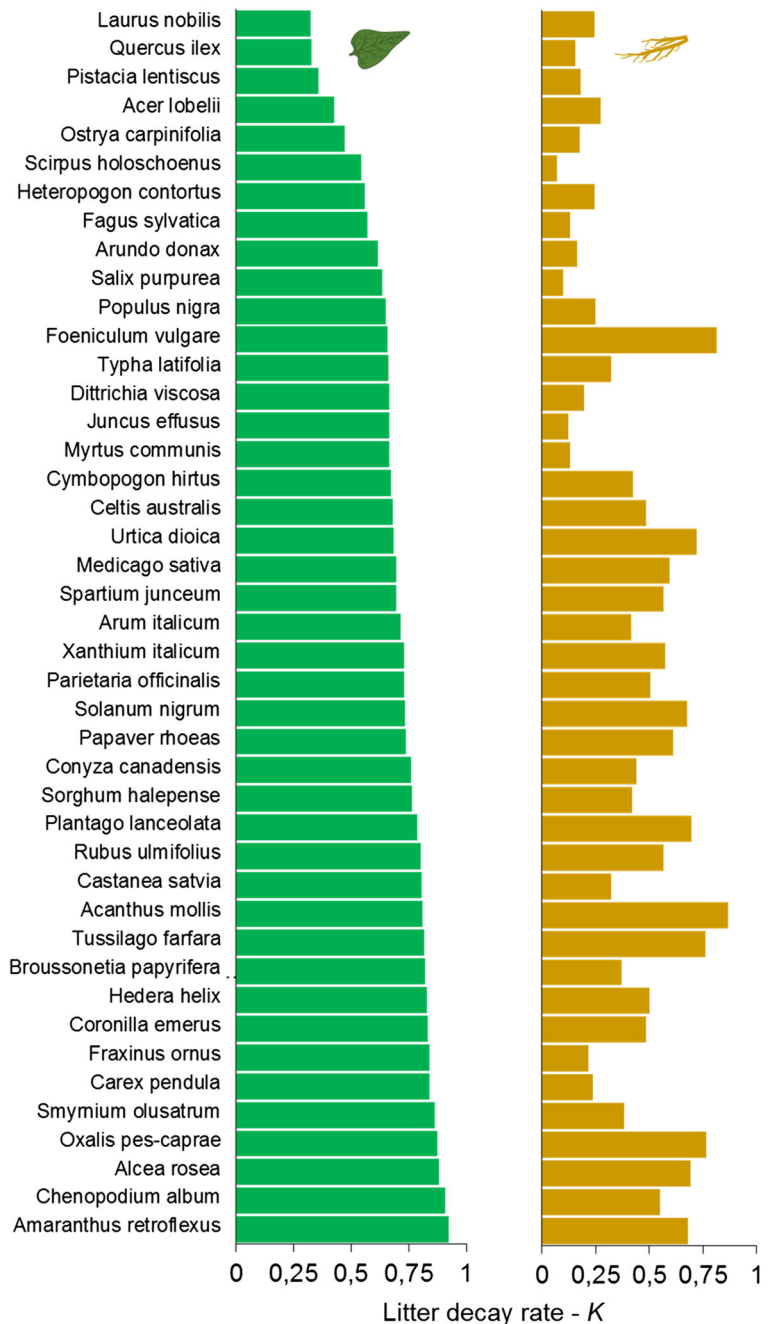
Across the 43 plant species, we found a slower mass loss for fine root litter compared to leaf litter at all incubation temperature. In detail, after two years of incubation at 4°, we found 35.31 and 56.57 % mass remaining for leaf litter and fine roots, respectively. At 14 °C the mass remaining was 10.01 % for leaf and 37.09 % for fine root, and only 2.55 % of leaf mass at 24 °C, while 23.72 % remained for fine root (Fig. 3a). Decay rate ( $K$ ) of leaf litters was faster than that of root litter at the three temperatures of incubation (paired t-test,  $P < 0.01$  in all cases) (Fig. 3b). Overall, root decay rate ( $K$ ) exceeds that of the corresponding leaf only in 12 cases out of 129 comparisons (9.3 % of the cases) (Fig. 3c).

#### Linking decay rate with woody debris traits

Considering the chemical parameters of undecomposed tissues, the decay rate ( $K$ ) showed a similar response for leaf and root with a trend of positive correlation with N and negative with lignin and carbon content, and lignin/N ratio (Fig. 4). Notably, the strength of the correlation changed with the temperature of incubation. The correlations were highly significant at 4 °C for lignin, C/N, lignin/N, and N but, at 14 °C and especially 24 °C they became weaker or even not significant. At 24 °C the correlation between  $K$  and lignin and lignin/N ratio remained significant for root litter but not for leaf litter. For N, instead, the positive correlation with  $K$  was significant only at 4 °C for leaf and at 4 and 14 °C for root litter. Finally, the correlation between the C/N ratio and  $K$  was significant only for the leaf litter at 4 °C of incubation.



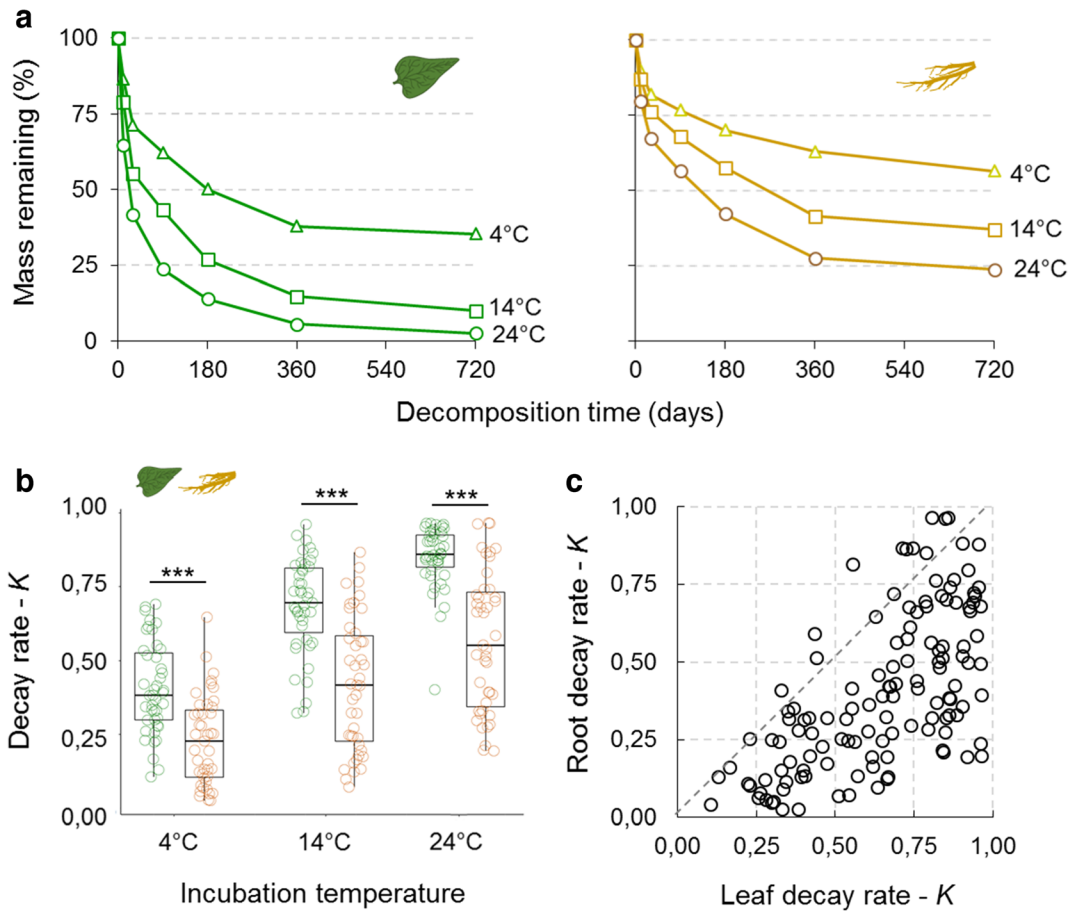
**Fig. 2** Decay rate (K) of leaf litter and fine roots of 43 species incubated at 14°C in laboratory conditions (statistics in Supplementary Table S4). Decay rate (K) recorded at 4°C and 24°C is reported in Supplementary Figure S1, S2



### Temperature sensitivity of leaf litter and fine root litter

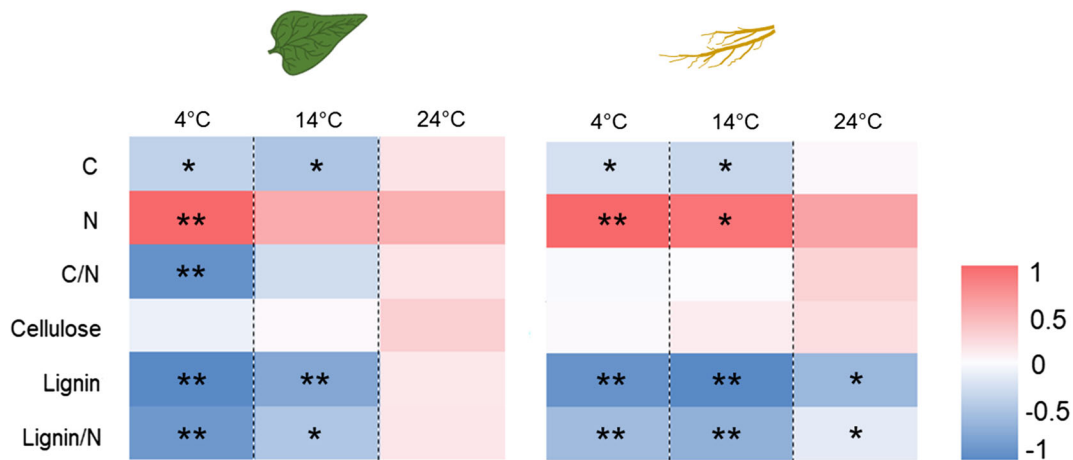
Mass loss significantly increased with temperature for litter types, with  $Q_{10}$  values (average  $\pm$  standard deviation) that were  $2.36 \pm 0.32$  for leaf litter and  $4.21 \pm 1.01$  for fine roots (paired t-test,  $P < 0.01$ ; Fig. 5a). Moreover,  $Q_{10}$  was higher for fine root than the corresponding leaf

litter in 35 out of 43 cases (81.40%) (Fig. 5b). For leaf litter,  $Q_{10}$  was positively correlated with N content and negatively with lignin and lignin/N ratio (Fig. 5c).  $Q_{10}$  of fine roots, instead, recorded only a highly significant negative correlation with lignin content. Finally,  $Q_{10}$  of both leaf litter and fine roots was negatively correlated with their decay rate (Fig. 5d).

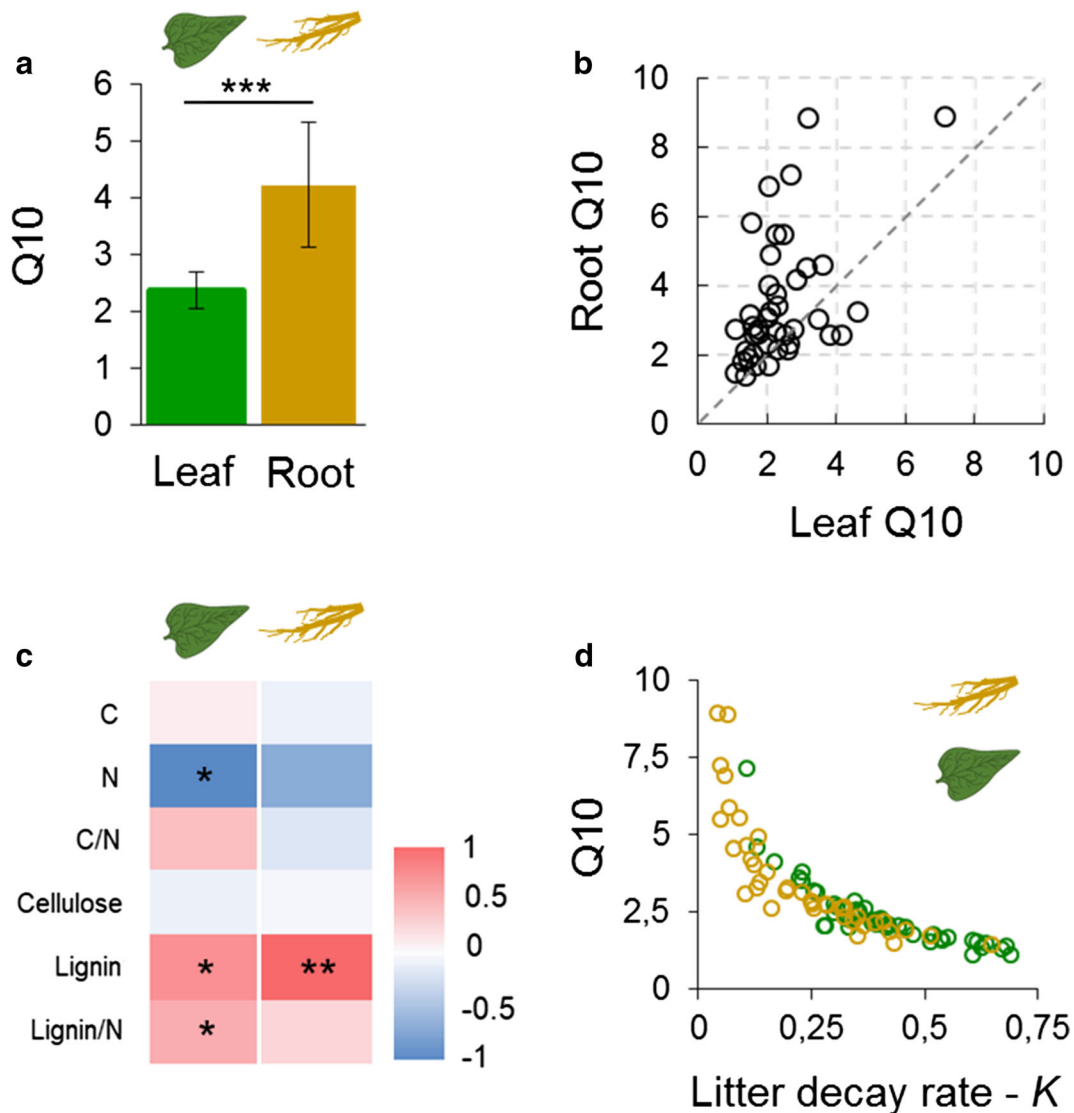


**Fig. 3** Decomposition of leaf litter and fine roots **a** Average percentage mass remaining for leaf litter and fine root mass of 43 co-occurring species incubated at three temperatures as a function of time (statistics in Supplementary Tables S5, S6); **b** pairwise

comparison of decay rate (K) of leaf and root litters incubated at three temperatures (Paired test-t, \*\*\* $P < 0.001$  in all cases), and (C) scatter plot reporting the decay rate (K) of leaf litter and fine roots of the 43 species



**Fig. 4** Heat-map reporting the correlations (Spearman's  $r$ ) between leaf litter and fine root chemical traits with decay rate (K). Asterisks indicate statistically significant correlation values, negative or positive (\*  $P < 0.05$ , \*\*  $P < 0.01$ )



**Fig. 5** Temperature sensitivity of leaf litter and fine roots of 43 Mediterranean species. **a** Q10 of leaf and fine root (Paired test-t, \*\*\* $P < 0.001$ ); **b** scatter plot of Q10 for all leaf and fine root types; **c** heat-map reporting the correlations (Spearman's  $r$ ) between litter

chemical traits and Q10 for leaf and fine root (asterisks indicate statistically significant correlation values, negative or positive; \*  $P < 0.05$ , \*\*  $P < 0.01$ ), and **d** the relationship between Q10 and decay rate ( $K$ ) of leaf litter and fine roots

## Discussion

Our experiment, based on forty-three paired leaf litter and fine root types demonstrated that the mass loss of root is consistently lower compared to leaf. In terms of litter chemistry, fine root litters had lower N content but higher lignin concentration, lignin/N, and C/N ratios compared to leaf litters. According to previous studies, the decay of plant tissues was negatively associated with lignin concentration, lignin/N, and C/N ratios but positively with N concentration. Notably,  $Q_{10}$  was higher

for fine root than leaf litter, and was positively correlated with lignin content and negatively with N concentration.

## Leaf litter and fine root decomposition

Fine roots of tested species decomposed at a slower rate than leaves at all incubation temperature. The decay rate of fine roots was lower than the corresponding leaf in 117 cases out of 129 comparisons (90.7 % of the cases). In this regards, our results confirmed the assumptions from previous studies that reported a slower degradation



for root litter than leaf litter (van Vuuren et al. 1993; Cusack et al. 2009; Xiong et al. 2013; Bonanomi et al. 2017b), although the opposite trend was also reported (Ostertag and Hobbie 1999). The limited mass loss suggested an important role of fine roots also for soil C sequestration and C and N dynamics in Mediterranean ecosystems. The slow decay of fine roots was probably due to chemical traits, mainly to a lower N concentration coupled with a higher lignin content of fine root litter as compared to corresponding leaf litter. Moreover, mass loss rates were negatively correlated with lignin content, lignin/N ratio but positively with N concentration. Correlation analysis highlighted the importance of N and lignin in controlling fine root decay rate and confirmed the well-established previous knowledge reported for both fine roots (See et al. 2019), and woody debris (Meentemeyer 1978; Weedon et al. 2009). In general terms, the low N concentration coupled with the high lignin content of fine root would induce an N limitation for decomposer microbes. C/N ratio of all decomposing substrate that lies above the threshold of  $\sim 30$  would be N limited (Taylor et al. 1989). In our study only 6 leaf litter types (13.9 %) has a C/N above 30 while 18 fine root litters (41.8 %) exceed this threshold, indicating that N limitation would be more important for fine root. Notably, previous studies demonstrated that microbe may require extra N from other sources like the underlying soil (Schimel and Hättenschwiler 2007; De Marco et al. 2008; Lummer et al. 2012; Berglund and Ågren 2012). In this regard, the present study probably underestimates the potential effect of N limitation because fine root litter was incubated on N rich soil from a woodland. In other words, fine root decomposing in nutrient-poor soils may exacerbate microbial N starvation and, thus, further slow their decay rate. Further studies carried out across the gradient of soil N availability could provide useful insights about the dynamics and impact of N limitation on fine root decomposition.

Our results supported the paradigm of lignin/N ratio control over litter decomposition for both leaf litter and fine roots. However, our experiment carried out across an ample temperature gradient i.e. from 4 to 24 °C, demonstrated that both lignin and N limitation are stronger at low temperature. The strong negative correlation between decay rate with lignin, lignin/N, and C/N ratio recorded at 4 °C became weaker or not significant at 24 °C. The same pattern was observed for N where the strong positive correlations observed at 4 °C became not significant at 24 °C. Surprisingly, when leaf litter was

incubated at 24 °C neither lignin/N ratio, C/N ratio, nor the concentrations of lignin and N were correlated with decay rate. These results suggest that both lignin and N concentration limit decomposition more intensely at low than high temperatures. In the case of lignin, it could be hypothesized that the complex enzymatic machinery required to decompose lignin is less efficient at low temperatures (Fioretto et al. 2000, 2018). Instead, the N limitation in N poor plant tissue could be more intense at low temperature because of the reduction of exogenous N input derived from slower mineralization of soil organic matter than at 24 °C (Stanford et al. 1973). In other words, at 24 °C the rapid mineralization from the underlying soil would provide a larger N source that could remedy the N limitation of N poor plant tissue. Overall, the increasing importance of lignin and N in controlling decomposition as temperature decreases deserves further studies.

Notably, the C/N ratio was negatively correlated with the decay rate of leaf litter, at least at low temperature, but not for fine root. Several studies criticized the C/N ratio, debating its capability to predict leaf (Hättenschwiler et al. 2011; Carteni et al. 2018) and fine root decomposition (Goebel et al. 2011). Instead, a recent study reported that the initial C/N ratio achieved very high and significant correlations with the decay rate of wood debris (Bonanomi et al. 2021). We suspect that the very high C/N ratio and wood debris with the consequent extreme N limitation would obscure the dependence of decomposition dynamics from this parameter; probably a more detailed description of C chemistry would be necessary. Furthermore, for both leaf litter and fine root, the need to define C-quality to predict decay rate seemed confirmed by the negative correlation shown between the lignin/N ratio and decay rate at all incubation temperatures. In this regard, our study is not in contradiction with the findings of Sun et al. (2018) where lignin content and lignin/N ratio were unrelated to the decay rate of first order roots of 35 woody species. Our study used a mixture of first- to third-order roots and so cannot contribute to the debate concerning the very low decay rate of the first order roots despite their low lignin and relatively high N concentration (Goebel et al. 2011; Beidler and Pritchard 2017). Further studies are needed to understand the recalcitrance of thin root tissues to decomposition focusing on root chemical, anatomical traits also with the extent and the type of mycorrhizal colonization (Langley et al. 2006). In this regards, the use of other chemical analysis such as  $^{13}\text{C}$

CPMAS NMR (Kögel-Knabner 2002), or anatomical traits like specific root length would be more informative in future studies (Liang et al. 2016).

#### Temperature sensitivity of decomposing leaf and fine root

Temperature sensitivity of plant litter is still a debated issue. Two field studies, carried out across ample latitudinal and latitudinal gradients, reported higher  $Q_{10}$  for leaf litter compared to root debris (Gholz et al. 2000; Song et al. 2021). In this study, instead, we found that roots increase their decay rate more than leaf as temperature rise, i.e. root decomposition is more temperature-sensitive than the corresponding leaf. This finding suggests that microbial as well as enzymatic activities, which are generally faster at higher temperatures, strongly interact with litter chemistry. Theoretical models and empirical evidence that suggest an increased sensitivity with substrate recalcitrance (Fierer et al. 2005; Hartley and Ineson 2008; Conant et al. 2008), although opposing evidence are available (e.g. Giardina and Ryan 2000; Liski et al. 1999). The decomposition process, like all chemical and biochemical reactions, are temperature-dependent. In this context, basic thermodynamics theory predicts that organic matter recalcitrant substrates, that usually resist decomposition, will be more sensitive to an increase in temperature (Bosatta and Ågren 1999). In detail, recalcitrant substrates require higher activation energy and, consequently, are more sensitive to temperature than labile organic matter (Davidson and Janssens 2006). According to this theory, our results based on 86 litter types demonstrate that the temperature sensitivity of decay litter decreases with increasing recalcitrance. In fact,  $Q_{10}$  was negatively correlated with initial N concentration and positively with lignin content, i.e. factors driving the decay rate. Besides, for both leaf litter and fine roots the  $Q_{10}$  was negatively correlated with litter decay rate: i.e. temperature sensitivity increase with decreasing litter decay. Fierer et al. (2005) using 24 litter types reported that as the overall quality of the organic C declined, litter decomposition became more sensitive to temperature. The authors also highlighted that litter temperature sensitivity would be different from that of soil organic matter because factors like chemical and physical protection in soil (Thornley and Cannell 2001) may potentially obscure any relationship between C quality and  $Q_{10}$ . Thus as highlighted in a recent study (Bonanomi

et al. 2017b), the temperature sensitivity of recalcitrant substrates could be obscured by N limitation, as well as by other constraints such as water scarcity, flooding, freezing, or physical and chemical protection of organic matter as previously reported. The higher temperature sensitivity of the fine root litter compared to leaf litter observed in this study contrast with several field observations (Gholz et al. 2000; Song et al. 2021). Therefore, further studies comparing laboratory and field decomposition are required before any proper integration of temperature sensitivity in modeling organic C at the ecosystem scale.

Finally, the direct comparison of our  $Q_{10}$  values with others reported in the literature should be done with caution because of the different methods used. Several laboratory-based experiments made to assess litter and soil temperature sensitivity to decomposition used short-term laboratory incubations and measured respiration rate. This approach had the advantage of eliminating the confounding effects of the sample temperature history and the litter quality does not change appreciably throughout these assays (Fierer et al. 2005). Instead, longer assay times as in the litter-bag experiment may lead to an underestimation of  $Q_{10}$  (Burke et al. 2003).

#### Conclusions

The data from this comparative study based on 43 plant Mediterranean species from 28 taxonomic families demonstrate that fine root litter decomposes at a slower rate than the leaf litter from the same species. Our results further indicate that fine roots had different chemistry, being more lignified with a lower N concentration compared to the corresponding leaf litter. Moreover, fine root decomposition is more sensitive to temperature with an average  $Q_{10}$  of 4.21 compared to 2.36 of leaf litter. According to the literature, we found that N, lignin, and lignin/N ratio well predict leaf litter and fine root decay but only at low temperature, with a poor predictive power at high temperature, suggesting that N and lignin limitation of decomposition is important only at low temperature. In contrast to leaf litter, the large interspecific variation in fine root decomposition cannot be predicted by the well-established C/N ratio. These findings, however, need future integration with field experiments before any integration in predictive frameworks and models of C and N biogeochemical cycling.

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#### Declarations

**Conflicts of interest/Competing interests** The authors declare no conflict of interest.

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