Transmission and terrestrial dispersal of non-native ectosymbionts on invasive crayfish

Rhiannon Hunt · John Rhidian Thomas · Joanna James · Joanne Cable

Received: 2 September 2017 / Revised: 11 May 2018 / Accepted: 11 May 2018 / Published online: 19 May 2018
© The Author(s) 2018

Abstract Symbionts are a fundamental component of biological systems, and their survival is highly dependent on transmission and host movement. Ectosymbionts of amphibious animals face the added challenge of having to survive dramatic environmental changes as their hosts cross ecosystem boundaries. Within freshwaters, crayfish are amongst the most widespread invasive species that readily disperse overland and are host to a wide range of ectosymbionts. Relatively little is known about the transmission of these ectosymbionts, including their ability to survive overland host migration. Here, we assessed terrestrial emigration and both inter- and intra-specific transmissions of *Xironogiton victoriensis*, a non-native branchiobdellidan (Annelida: Clitellata) recently found on invasive signal crayfish (*Pacifastacus leniusculus*) in the UK. These branchiobdellidans tolerated desiccation and did not alter host terrestrial behaviour. Transmission was rapid between natural signal and novel virile (*Orconectes* cf. *virilis*) crayfish hosts, with host interactions facilitating transmission.

Thus, branchiobdellidans can disperse via amphibious host behaviour and readily infect novel hosts. These traits facilitate symbionts' survival and provide access to additional dispersal pathways that are likely to aid transmission.

Keywords Invasive non-native species · Branchiobdellidans · *Pacifastacus leniusculus* · *Xironogiton victoriensis*

Introduction

Dispersal is a fundamental life history trait, offering opportunities for range expansion, genetic differentiation and increased fitness. As such, both population and evolutionary dynamics are influenced by a species' dispersal capacity (Ronce, 2007). In aquatic ecosystems, dispersal is usually limited to connected water bodies, with isolated habitats (such as ponds or lakes) only accessible via flooding events or anthropogenic activity (Hulme et al., 2008; Rahel & Olden, 2008). Amphibious behaviour, however, offers not only opportunities for foraging, reproduction and escape from unfavourable conditions, but terrestrial emigration to nearby water bodies (Sayer & Davenport, 1991). This has played an important role in the spread of invasive non-native species (Peterson et al., 2013; Marques et al., 2014; Ramalho & Anastácio,

Handling editor: María del Mar Sánchez-Montoya

R. Hunt · J. R. Thomas · J. James · J. Cable (✉)
School of Biosciences, Cardiff University,
Cardiff CF10 3AX, UK
e-mail: cablej@cardiff.ac.uk

J. James
National Fisheries Services, The Environment Agency,
Bromholme Lane, Brampton PE28 4NE, UK

2014; Fialho et al., 2016). Invasive species can detrimentally affect biodiversity, human health and industry in their introduced range (Mack et al., 2000), and determining their dispersal capacity is paramount for non-native species risk assessments (Johnson & Padilla, 1996).

Crayfish are amongst the most successful aquatic invaders, with multiple species now present in many European countries (Holdich & Pöckl, 2007; James et al., 2014; Kouba et al., 2014). Certain species have the capacity to disperse up to 1 km overland and move between isolated water bodies (Lutz & Wolters, 1999; Ramalho & Anastácio, 2014; Fialho et al., 2016), where they act as ecosystem engineers and interact with organisms on multiple trophic levels (Gherardi, 2007; Strayer, 2010; James et al., 2015a). Because of their success and impact, invasive crayfish are considered a major cause of biodiversity loss (Lodge et al., 2000; Manchester & Bullock, 2000). Non-native crayfish have further impacted their native counterparts through the introduction of co-existing symbionts (Lymbery et al., 2014). In Europe, invasion of North American crayfish has been aided by their ability to transmit crayfish plague (*Aphanomyces astaci* Schikora, 1906), which has caused high mortalities of susceptible native crayfish species (Holdich et al., 2014; Svoboda et al., 2016). Whilst *A. astaci* is arguably the most detrimental introduced crayfish symbiont, various crayfish ectosymbionts including branchiobdellidans, temnocephalans and ostracods have also been co-introduced globally (Gelder et al., 1999; Ohtaka et al., 2005; Aguilar-Alberola et al., 2012; du Preez & Smit, 2013). These introductions have almost certainly been instigated by anthropogenic activity. Whether further dispersal has been aided by host overland movement is unknown particularly as these ectosymbionts could be vulnerable to desiccation. Little is currently known about the ability of aquatic ectosymbionts to tolerate desiccation and spread alongside their hosts.

For this study, we focused on branchiobdellidans, which are ectosymbiotic annelids commonly found on crayfish throughout the Holarctic (Gelder, 1999). Their relationship with crayfish can vary across the symbiosis continuum (Lee et al., 2009; Brown et al., 2012; Thomas et al., 2016) and as such, they have the capacity to affect host invasion dynamics. Branchiobdellidans possess several characteristics associated with successful invaders, including a direct life cycle

(Young, 1966; Creed et al., 2015) and broad host range (Goodnight, 1940; although host preference does occur in some species, see Brown & Creed, 2004), which has led to invading branchiobdellidan species spreading to novel hosts (Sobecka et al., 2011; Vedia et al., 2014). Considering the number of non-native crayfish introductions into Europe, reports of introduced branchiobdellidans are relatively few. Whilst this may be due to lack of reporting, branchiobdellidans could be lost during terrestrial host emigration. Hosts may also exploit their environmental tolerances to reduce infection, for example infected insects, reptiles and fish choose to inhabit different temperatures (termed “behavioural fever”), which increases their immune response and decreases pathogen survival (Vaughn et al., 1974; Müller & Schmid-Hempel, 1993; Mohammed et al., 2016). The time crayfish spend on land and frequency of amphibious behaviour could therefore be affected by branchiobdellidan infection. Alternatively, non-native branchiobdellidans may tolerate terrestrial emigration but fail to establish due to host population bottlenecks encountered during the introduction phase of invasion.

Non-native branchiobdellidans [*Xironogiton victoriensis* (Gelder & Hall 1990)] were recently found on invasive signal crayfish [*Pacifastacus leniusculus* (Dana, 1852)] in the UK (James et al., 2015b). Compared to other freshwater annelids, *X. victoriensis* possess a relatively short reproduction time (cocoons hatch 10–27 days after laying) and can quickly colonise a host following transmission (Govedich et al., 2009; James et al., 2017). Like other *Xironogiton* spp., *X. victoriensis* are predominantly located on the crayfish chelae (Gelder et al., 2009; James et al., 2015b). This location may facilitate transmission as crayfish often lock chelae during aggressive interactions, but it could also promote desiccation during terrestrial host movements. Although branchiobdellidans are generally considered commensals, *X. victoriensis* reduces crayfish aggression and foraging efficiency, which could alter signal crayfish invasion success (James et al., 2015c). Here, we tested whether these aquatic ectosymbionts can tolerate amphibious host behaviour, which in turn could affect symbiont dispersal with potential consequences for host invasion dynamics. Specific aims were to (i) verify whether *X. victoriensis* can survive terrestrial host emigration, and if so; (ii) determine if terrestrial crayfish behaviour is affected by *X. victoriensis*; (iii)

examine transmission of *X. victoriensis* from the environment and between intra-specific and inter-specific crayfish pairs, and finally; (iv) assess how host interactions facilitate transmission.

Materials and methods

Animal origin and maintenance

Crayfish were captured throughout March–October in 2012–2016 via baited traps and manual searching. Animals were maintained in the lab for 2–4 months prior to the experiments described below. Signal crayfish infected with *Xironogiton victoriensis* were caught at the Gavenny River (Abergavenny, Wales, Grid Reference: SO308164), whilst uninfected, branchiobdellidan naïve signal crayfish were obtained from Dderw Farm Pond (Powys, Wales, Grid Reference: SO138375). Uninfected virile crayfish [*Orconectes virilis* (Hagen, 1870)] were collected from the River Lee (London, England, Grid Reference: TL370028). Although *X. victoriensis* have never been reported on virile crayfish, signal and virile crayfish co-exist in the UK (James et al., 2015d).

All crayfish were caught under licence (NT/CW065-C-652/5706/01) and held at Cardiff University (W C ILFA 002), where infected and uninfected crayfish were maintained separately at a density of 20 individuals per m² in tanks of dechlorinated water (14 ± 1°C). The tanks contained a pea gravel substrate (2 cm) and refugia (plant pots and plastic tubes) with no access to a terrestrial area. Crayfish were fed daily with bloodworm (*Tubifex* spp.) and frozen peas with 50% water changes performed weekly to maintain water quality. All experiments were conducted under these laboratory conditions in windowless rooms to prevent external lighting. Experiments 1 (Dispersal) and 4 (Intra-specific group transmission) were performed under a 12-h-light:12-h-dark photoperiod to provide equal periods of time for diurnal and nocturnal crayfish activity during behavioural observations. A 16-h-light:8-h-dark photoperiod was used during Experiments 2 (Environmental transmission) and 3 (Intra- and inter-specific dyad transmission).

Only healthy branchiobdellidan naïve crayfish in the intermoult stage were used in experiments, with all crayfish measured (carapace length, accurate to

0.1 mm) and sexed. Experiments 1 (Dispersal) and 4 (Intra-specific group transmission) were conducted using 189 L aquaria (0.5 m² area) stocked with groups of four crayfish (density within natural UK signal crayfish populations which can range from 2 to 34 crayfish per m²: Guan & Wiles, 1996; Bubb et al., 2004). These experiments were also recorded using 24-h infrared CCTV cameras (Sentient Pro HDA DVR 8 Channel CCTV, Maplin) with individual crayfish numbered using non-toxic yellow paint on the dorsal carapace (a region *X. victoriensis* does not inhabit; James et al., 2015b) for visual identification. In accordance with the Wildlife and Countryside Act 1981, signal and virile crayfish were humanely destroyed by freezing at – 20°C upon termination of experiments as they could not be returned to the wild due to their invasive status.

Individual *X. victoriensis* (identified according to James et al., 2015b) were carefully dislodged from stock crayfish after encouraging them to move using blunt forceps and placed onto a petri dish (experiment 1—dispersal), ceramic tile (experiment 2—environmental transmission) or new, branchiobdellidan naïve crayfish (experiments 3—Intra- and inter-specific dyad transmission and 4—intra-specific group transmission). Removal from the original host did not cause any visible damage to the worms, and they readily re-attached to the petri dish, tile or new host. Previous in vitro experiments show that worms removed using this method survive and successfully reproduce long-term (James et al., 2017). Only adult (> 3.0 mm), *X. victoriensis* that were visually healthy were used in experiments, and any *X. victoriensis* that detached 30 min following transfer were replaced. All other crayfish used in transmission experiments were sham infected to ensure they experienced the same handling procedure and period of time out of water. Crayfish were immediately used in experiments following artificial infection to ensure the same starting number of worms across experiments as host grooming can rapidly reduce worm numbers following infection. In addition, an acclimation period could have resulted in worm reproduction and/or crayfish moulting (James et al., 2017).

Experiment 1: dispersal

To investigate the maximum survival time out of water, *X. victoriensis* were removed from their host

and subjected to periods of dehydration at 15°C (59% RH) and 23°C (41% RH). Five individual *X. victoriensis* were added to a petri dish containing 10 ml water ($n = 10$ petri dishes per temperature treatment). After ensuring all worms were alive, active and firmly attached to the bottom of the petri dish, the water was gently poured from the petri dishes leaving the worms in situ, with excess water removed using an absorbent paper wick. Each petri dish was refilled with dechlorinated water following a set period of dehydration (1, 2, 4, 8, 15, 20, 25, 30, 40 or 50 min; times selected following preliminary trials). The worms were then left in water for 24 h at 15 or 23°C to allow them to rehydrate and animate, and then the number of worms alive in each petri dish was counted. In this study, we did not attempt to monitor *X. victoriensis* desiccation survival on the host as crayfish had a tendency to groom some of the branchiobdellidans from their chelae immediately after artificial infection (also reported previously by Farrell et al., 2013; Skelton et al., 2014).

The capacity of *X. victoriensis* to exploit new hosts in recipient water bodies following terrestrial host emigration was assessed by placing infected signal crayfish ($n = 4$) into an aquarium containing ramps (29 × 43 cm) that provided access to a terrestrial bridge (250 × 20.5 cm). The bridge was connected to another aquarium containing uninfected signal crayfish ($n = 4$) (Fig. 1). The ramps and bridge were lined with affixed pea gravel substrate to aid crayfish movement. Crayfish were left to acclimatise for 1 day with an opaque divider on the bridge preventing movement or visual exchange between the two aquaria. Following removal of the divider, individuals were screened for branchiobdellidans every morning

(09:00 h) until a crayfish in the uninfected population became infected (maximum of 7 days).

To determine whether *X. victoriensis* infection influenced how often signal crayfish left the water and the length of time spent on land (per exit and total time), infected ($n = 4$) and uninfected ($n = 4$) crayfish were placed into aquaria and given access to a 125 × 20.5 cm section of land. Crayfish were filmed for 24 h with individual terrestrial activity determined from analysis of video footage ($n = 10$ replicates, 80 crayfish).

Experiment 2: environmental transmission

To investigate indirect transmission of *X. victoriensis* from the environment to crayfish, 30 individuals of *X. victoriensis* attached to a ceramic tile (5 × 5 cm²) were placed in a 15-l aquarium with one uninfected signal crayfish ($n = 20$). The number of *X. victoriensis* on the crayfish was recorded after 1, 2, 4, 6, 8, 10, 12, 18 and 24 h.

Experiment 3: intra- and inter-specific dyad transmissions

To assess both intra- and inter-specific transmissions of *X. victoriensis*, artificially infected ‘donor’ signal crayfish (30 worms per individual) were placed in a 15-l aquarium at 14:00 h with either a ‘recipient’ signal ($n = 20$) or virile crayfish ($n = 20$) sex and size matched to within 10% carapace length (mm) of the donor. The number of *X. victoriensis* worms on the recipient crayfish was recorded after 1, 2, 4, 6, 8, 10, 12, 18 and 24 h. The experiment was repeated for intra-specific virile crayfish pairs ($n = 20$).

Fig. 1 Terrestrial dispersal of *Xironogiton victoriensis* experimental design. To determine how quickly *X. victoriensis* could transmit to separate host populations following overland host movement, crayfish were given free movement between tanks across a bridge

Insufficient virile crayfish were available to investigate virile-to-signal crayfish transmission.

Experiment 4: intra-specific group transmission

Xironogiton victoriensis transmission within a host population was assessed using groups of signal crayfish ($n = 10$ replicates, 40 crayfish). Crayfish ($n = 4$) were placed into an aquarium with a single uniform shelter large enough for all the crayfish ($29 \times 43 \text{ cm}^2$) and left to acclimatise for 3 days. After acclimatisation, an individual crayfish (44–48 mm carapace length) was experimentally infected with 154 *X. victoriensis* worms to match natural intensities (James et al., 2015b). This created a starting group of one infected donor crayfish and three uninfected recipient crayfish. Crayfish behaviour was then observed for 1 week. As branchiobdellidan transmission is linked to host contact (Young, 1966), all interactions between individual crayfish (aggressive and shelter sharing behaviours) were analysed via video recording. Intensity of *X. victoriensis* infection was quantified after 1 week for each individual. No branchiobdellidan reproduction could have occurred during the experiment as *X. victoriensis* cocoons take at least 10 days to hatch (under the same laboratory conditions used for these experiments: James et al., 2017).

Statistical analyses

Models were refined through stepwise deletion of insignificant terms and/or AIC comparisons with visual examination of model plots to check standardised residuals for normal distribution and homogeneity of variance (Crawley, 2007). Generalised Linear Mixed Models (GLMMs) were implemented using the glmmADMB package (Skaug et al., 2016) to include trial number as a random factor accounting for group effects. In all tests, the level of significance was taken as $P < 0.05$.

To assess whether temperature and (log) time significantly affected the proportion of worms that survived dehydration, we used a General Linear Model (GLM). For terrestrial emigration, a GLMM with a binomial error distribution and logit link function was used to compare successful transmission (yes/no) to the number of host crossings (total number normalised using log transformation). To determine

the influence of *X. victoriensis* intensity on the tendency of crayfish to leave the water, we also used a GLMM with a binomial error distribution and logit link function. Another GLMM with a gaussian error distribution and identity link function was then used to examine whether the duration of time crayfish spent on land (square root transformed, $n = 29$ as crayfish that did not leave the water were discounted) was affected by *X. victoriensis* intensity.

Transmission of *X. victoriensis* to uninfected hosts was analysed using two separate GLMs; The first model investigated the effect of transmission pathway (environment–signal, signal–signal, virile–virile, signal–virile) and crayfish size (mean pair carapace length) on the speed of transmission (log time to first worm transfer, 10 signal–signal pairs were excluded from this analysis as they were accidentally not checked at the 1-h stage). The second model determined the effect of transmission pathway and crayfish size on the $(\log x + 1)$ maximum number of worms transferred to each crayfish. Significant differences between transmission pathways were examined with a Tukey's Honest Significant Difference means comparison.

To determine if time spent interacting, sharing shelter, host size (carapace length), chelae size or sex influenced transmission of *X. victoriensis* in a group, a GLMM was used with a negative binomial error distribution and log link function. A GLM was also performed to examine whether time crayfish spent interacting with conspecifics was influenced by crayfish size or sex.

All statistical analyses were conducted in R statistical software v3.3.1 (R Core Team, 2016).

Results

Experiment 1: terrestrial emigration

Xironogiton victoriensis tolerated desiccation and survived significantly better at 15°C (40–50 min) than at 23°C ($> 15 \text{ min}$) ($F_{1,37} = 11.52$, $P < 0.001$). Terrestrial emigration frequency of signal crayfish between two connected water bodies significantly affected whether *X. victoriensis* transmission to the uninfected host population was successful ($\chi^2 = 12.22$, $\text{df} = 1$, $P < 0.001$). Terrestrial emigration of an infected crayfish to the uninfected

population resulted in successful transmission in 70% of trials after a maximum of 7 days, with the fastest transmission occurring after 1 day. The majority of crayfish that left their aquarium proceeded to cross over to the adjacent aquarium (90%, $n = 182$ instances of leaving the water with 163 full crossings) with the shortest crossing taking 2 min and the longest 14 min. The longest period of time spent on the bridge was 30 min; however, that individual did not enter the new aquarium and instead returned to its original aquarium. During the terrestrial behaviour trials, the tendency of signal crayfish to leave the water and duration of time spent on land was not significantly affected by *X. victoriensis* intensity. The total time individuals spent on land (calculated from multiple outings) ranged from 2 to 112 min with 36% of crayfish leaving the water ($n = 80$). The longest single period of time spent out of water by an individual crayfish was 58 and 41 min for uninfected and *X. victoriensis* infected crayfish, respectively.

Experiments 2 and 3: environmental, intra- and inter-specific dyad transmission

Table 1 summarises the results of *X. victoriensis* transmission from the environment to signal crayfish and between host pairs. Transmission within 24 h was 100% successful from the environment to signal crayfish, 95% for both intra-specific signal–signal and virile–virile, and 70% for interspecific signal–virile pairs. Success was significantly lower for interspecific signal–virile pairs compared to environment, signal–signal and virile–virile trials ($P < 0.05$). The time until first worm transfer and maximum number of worms transferred was significantly

dependent upon transmission route ($F_{3,59} = 3.67$, $P = 0.020$; $F_{3,76} = 28.59$, $P < 0.001$), with the fastest first worm transfer occurring between virile conspecifics and highest total number of worms transferred from the environment to signal crayfish. Crayfish size had no effect on either the speed of transmission or the number of *X. victoriensis* transferred.

Experiment 4: intra-specific group transmission

Following one week of co-habitating with infected crayfish, *X. victoriensis* intensity on recipient crayfish significantly increased with time spent sheltering ($\chi^2 = 17.10$, $df = 1$, $P < 0.001$) and interacting ($\chi^2 = 5.35$, $df = 1$, $P = 0.020$) (Fig. 2). In addition, male crayfish became infected with significantly more worms than female crayfish ($\chi^2 = 5.55$, $df = 1$, $P = 0.018$). Males also spent more time interacting with conspecifics (1130 min for males versus 653 min for females); however, this was not a significant difference. Crayfish size did not significantly affect *X. victoriensis* intensity or time spent interacting with conspecifics. By the end of the experimental period (1 week), all recipient crayfish had branchiobdellidan cocoons and 0–41 *X. victoriensis* worms.

Discussion

Symbioses are implicit in biological systems (Brooks, 2012); however, pathways of symbiont dispersal are often unknown. This study shows that branchiobdellidans possess several life-history traits which aid dispersal, establishment and spread. *Xironogiton*

Table 1 Transmission of *Xironogiton victoriensis*

Transmission route	Proportion (%) naïve hosts infected	Mean (range) time to first transmission (h)	Mean (range) total worms transmitted
Environment–signal	100	3.1 (1–12)	6.8 (3–12)
Signal–signal	95	6.6 (1–24)	1.6 (1–4)
Virile–virile	95	2.9 (1–18)	1.9 (1–4)
Signal–virile	70	7.2 (1–24)	2.4 (1–8)

Proportion of naïve hosts infected, time taken for transmission to first occur and total number of *X. victoriensis* transmitted from the environment to signal crayfish, and between pairs of conspecifics (signal to signal crayfish and virile to virile crayfish) and heterospecifics (signal to virile crayfish), ($n = 20$ trials per treatment group)

Fig. 2 Influence of signal crayfish *Pacifastacus leniusculus* behaviour: **a** host–host interactions and **b** sheltering sharing on *Xironogiton victoriensis* mean intensity. The number of individuals of *X. victoriensis* per host increased with time that

hosts spent interacting/sheltering with infected signal crayfish. Predictions were created using the final model, with mean values used for time spent interacting/sheltering and both sexes. Dotted lines represent standard error

victoriensis survived 40–50-min desiccation and infection did not alter terrestrial crayfish behaviour, as such, amphibious host movement provides an additional emigration opportunity. Furthermore, *X. victoriensis* readily infected both signal and virile crayfish with host interactions facilitating transmission.

To establish, ectosymbionts of amphibious invaders have to survive exposure to different environments when their hosts cross ecosystem boundaries. Ectosymbionts typically tolerate desiccation far less than their hosts; however, some species are capable of surviving overland migration. Ectoparasitic gyrodactylids are prone to desiccation but can survive on killifish (*Rivulus hartii* Boulenger, 1890), which migrate terrestrially (Sayer & Davenport, 1991; Cable et al., 2013), whilst aquatic leeches can be transported overland via crocodilians and waterfowl (Davies et al., 1982; Leslie et al., 2011). *X. victoriensis* is evidently capable of surviving translocation and dispersal due to the number of successful introductions and established populations across Europe (currently confirmed in seven countries; see James et al., 2015b).

In the present study, *X. victoriensis* survived terrestrial host movement and did not alter this behaviour, whilst also tolerating 40–50-min desiccation off the host. The terrestrial walking speed of signal crayfish is unknown, but adult male red swamp

crayfish (*Procambarus clarkii* Girard, 1852), another invasive North American species, walk on average 58 m h^{-1} (Ramalho & Anastácio, 2014). At this rate, we estimate *X. victoriensis* could survive terrestrial migration up to 43.5 m, although this does not take into account variable environmental conditions (Marques et al., 2014; Ramalho & Anastácio, 2014; Yoder et al., 2016). It is also possible that branchiobdellidans could tolerate longer periods of desiccation in vivo by retreating into host crevices or gill chambers, but for *X. victoriensis* this is unlikely given that this species is a chelae specialist (Gelder & Hall, 1990; James et al., 2015b). Regardless, it is likely that even based on our conservative estimates, branchiobdellidans would be able to survive natural crayfish overland dispersal, which can exceed 20 m (Puky, 2014; Ramalho & Anastácio, 2014). Cocoons present another opportunity for dispersal: oligochaete cocoons, for example, can have a higher desiccation tolerance than juvenile and adult worms (Holmstrup, 2001; Govedich et al., 2009). Therefore, even where juvenile/adult worms desiccate, cocoons may survive terrestrial dispersal and proceed to hatch and establish a viable population.

Following establishment, dispersal of non-native branchiobdellidans is likely promoted by their low host-specificity and the co-existence of multiple North American crayfish species (Koubat et al., 2014). In the present study, *X. victoriensis* transmitted readily

between signal and virile crayfish, which co-exist in the UK (James et al., 2015d). Intra-specific transmission occurred faster between novel virile crayfish hosts, compared to signal crayfish, potentially due to the higher aggression of this species and thus more frequent interactions (James et al., 2015d). The maximum number of worms transmitted, however, was not dependent upon the recipient host species. This supports previous reports indicating that whilst some branchiobdellidan species exhibit host preferences (Brown & Creed, 2004), most crayfish species are acceptable hosts (alongside certain crabs and shrimps; Gelder & Messick, 2006; Govedich et al., 2009; Niwa et al., 2014). *X. victoriensis* has been reported on invasive red swamp crayfish populations in Spain, likely due to cohabitation with infected signal crayfish (Vedia et al., 2014). Red swamp and virile crayfish are present in the UK together with five other non-native and one native crayfish species (James et al., 2014). In addition, other suitable crustacean hosts are present, for example, invasive Chinese mitten crabs (*Eriocheir sinensis* H. Milne Edwards, 1853), which *X. victoriensis* could exploit (Sobecka et al., 2011). Access to multiple hosts could, however, prompt a ‘dilution effect’ whereby infection on signal crayfish (i.e. the natural host of *X. victoriensis*) is reduced due to the increasing diversity of suitable hosts (Keesing et al., 2006).

As branchiobdellidan transmission is dependent on host contact (Manus, 1960; Young, 1966), crayfish behaviours that increase contact will evidently promote symbiont spread. Both host shelter sharing and, in particular, aggressive interactions (involving interlocking chelae, the primary niche of *X. victoriensis*; James et al., 2015b) positively increased *X. victoriensis* transmission in the current study. We also found male signal crayfish were infected with significantly more *X. victoriensis* than females. Whilst grooming behaviours can play a major role in the control of crayfish ectosymbionts (Farrell et al., 2014; Skelton et al., 2014), it is likely that this sex difference in branchiobdellidan intensity was due to a disparity in male/female interactions. Male crayfish are typically more aggressive (Ranta & Lindström, 1993) and, although not statistically significant, spent almost twice the time interacting compared to females.

Overall, we demonstrate that *X. victoriensis* can survive terrestrial host emigration, and this may facilitate the movement of these symbionts between

isolated water bodies. Given that *X. victoriensis* is capable of reducing signal crayfish aggression and foraging efficiency, these results have potential implications for the invasion dynamics of these highly successful invasive species. The spread of these branchiobdellidians in the UK could be further promoted by their propensity to transmit to novel non-native crayfish hosts, as demonstrated in the current study. As crayfish are keystone species and ecosystem engineers that interact with organisms on multiple trophic levels, the presence of these behaviour-altering symbionts may have ecosystem-wide consequences.

Acknowledgements We thank Amanda Chapman for laboratory assistance. The Cardiff University Morgan E Williams Helminthology Scholarship, the Natural Environment Research Council (studentship NE/L002434/1) (RH) and the Coleg Cymraeg Cenedlaethol (JRT) funded the project.

Compliance with ethical standards

Conflict of interest The authors declare that they have no conflict of interest.

Open Access This article is distributed under the terms of the Creative Commons Attribution 4.0 International License (<http://creativecommons.org/licenses/by/4.0/>), which permits unrestricted use, distribution, and reproduction in any medium, provided you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made.

References

- Aguilar-Alberola, J. A., F. Mesquita-Joanes, S. López, A. Mestre, J. C. Casanova, J. Rueda & A. Ribas, 2012. An invaded invader: high prevalence of entocytherid ostracods on the red swamp crayfish *Procambarus clarkii* (Girard, 1852) in the Eastern Iberian Peninsula. *Hydrobiologia* 688: 63–73.
- Brooks, W., 2012. The importance of symbioses in biological systems. *Journal of Marine Science: Research & Development* 2: 1–2.
- Brown, B. L. & R. P. Creed, 2004. Host preference by an aquatic ectosymbiotic annelid on 2 sympatric species of host crayfishes. *Journal of the North American Benthological Society* 23: 90–100.
- Brown, B. L., R. P. Creed, J. Skelton, M. A. Rollins & K. J. Farrell, 2012. The fine line between mutualism and parasitism: complex effects in a cleaning symbiosis demonstrated by multiple field experiments. *Oecologia* 170: 199–207.
- Bubb, D. H., T. J. Thom & M. C. Lucas, 2004. Movement and dispersal of the invasive signal crayfish *Pacifastacus*

- leniusculus* in upland rivers. *Freshwater Biology* 49: 357–368.
- Cable, J., G. A. Archard, R. S. Mohammed, M. McMullan, J. F. Stephenson, H. Hansen & C. van Oosterhout, 2013. Can parasites use predators to spread between primary hosts? *Parasitology* 140: 1138–1143.
- Crawley, M. J., 2007. Statistical modelling. In Crawley, M. J. (ed.), *The R Book*. Wiley, New Jersey: 323–386.
- Creed, R. P., J. D. Lomonaco, M. J. Thomas, A. Meeks & B. L. Brown, 2015. Reproductive dependence of a branchiobdellidan annelid on its crayfish host: confirmation of a mutualism. *Crustaceana* 88: 385–396.
- Davies, R. W., L. R. Linton & F. J. Wrona, 1982. Passive dispersal of four species of freshwater leeches (Hirudinoidea) by ducks. *Freshwater Invertebrate Biology* 1: 40–44.
- du Preez, L. & N. Smit, 2013. Double blow: alien crayfish infected with invasive temnocephalan in South African waters. *South African Journal of Science* 109: 1–4.
- Farrell, K. J., R. P. Creed & B. L. Brown, 2013. Preventing over-exploitation in a mutualism: partner regulation in the crayfish–branchiobdellid symbiosis. *Oecologia* 174: 501–510.
- Farrell, K. J., R. P. Creed & B. L. Brown, 2014. Reduced densities of ectosymbiotic worms (Annelida: Branchiobdellida) on reproducing female crayfish. *Southeastern Naturalist* 13: 523–529.
- Fialho, C., F. Banha & P. M. Anastácio, 2016. Factors determining active dispersal capacity of adult Chinese mitten crab *Eriocheir sinensis* (Decapoda, Varunidae). *Hydrobiologia* 767: 321–331.
- Gelder, S. R., 1999. Zoogeography of branchiobdellidans (Annelida) and temnocephalidans (Platyhelminthes) ectosymbiotic on freshwater crustaceans, and their reactions to one another *in vitro*. In Healy, B. M., T. B. Reynoldson & K. A. Coates (eds), *Aquatic Oligochaetes*. Springer, Netherlands: 21–31.
- Gelder, S. R. & L. A. Hall, 1990. Description of *Xironogiton victoriensis* n.sp. from British Columbia, Canada, with remarks on other species and a Wagner analysis of *Xironogiton* (Clitellata: Branchiobdellida). *Canadian Journal of Zoology* 68: 2352–2359.
- Gelder, S. R. & G. Messick, 2006. First report of the aberrant association of branchiobdellidans (Annelida: Clitellata) on blue crabs (Crustacea: Decapoda) in Chesapeake Bay, Maryland, USA. *Invertebrate Biology* 125: 51–55.
- Gelder, S. R., G. B. Delmastro & J. N. Rayburn, 1999. Distribution of native and exotic branchiobdellidans (Annelida: Clitellata) on their respective crayfish hosts in northern Italy, with the first record of native Branchiobdella species on an exotic North American crayfish. *Journal of Limnology* 58: 20–24.
- Gelder, S. R., L. McCurry & D. F. McAlpine, 2009. Distribution and first records of Branchiobdellida (Annelida: Clitellata) from crayfishes (Crustacea: Decapoda) in the Maritime Provinces of Canada. *Northeastern Naturalist* 16: 45–52.
- Gherardi, F., 2007. Understanding the impact of invasive crayfish. In Gherardi, F. (ed.), *Biological invaders in inland waters: profiles, distribution, and threats*. Springer, Netherlands: 507–542.
- Goodnight, C. J., 1940. Biology. In Goodnight, C. J. (ed.), *The Branchiobdellidae (Oligochaeta) of North American crayfishes*. The University of Illinois Press, Urbana: 65–66.
- Govedich, F. R., B. B. Bain, W. E. Moser, S. R. Gelder, R. W. Davies & R. O. Brinkhurst, 2009. Annelida (Clitellata): Oligochaeta, Branchiobdellida, Hirudinida, and Acanthobdellida. In Thorp, J. H. & A. P. Covich (eds), *Ecology and Classification of North American Freshwater Invertebrates*, 3rd ed. Academic Press/Elsevier, San Diego: 385–436.
- Guan, R. & P. R. Wiles, 1996. Growth, density and biomass of crayfish, *Pacifastacus leniusculus*, in a British lowland river. *Aquatic Living Resources* 9: 265–272.
- Holdich, D. M. & M. Pöckl, 2007. Invasive crustaceans in European inland waters. In Gherardi, F. (ed.), *Biological invaders in inland waters: profiles, distribution, and threats*. Springer, Netherlands: 29–75.
- Holdich, D. M., J. James, C. Jackson & S. Peay, 2014. The North American signal crayfish, with particular reference to its success as an invasive species in Great Britain. *Ethology Ecology & Evolution* 26: 232–262.
- Holmstrup, M., 2001. Sensitivity of life history parameters in the earthworm *Aporrectodea caliginosa* to small changes in soil water potential. *Soil Biology and Biochemistry* 33: 1217–1223.
- Hulme, P. E., S. Bacher, M. Kenis, S. Klotz, I. Kühn, D. Minchin, W. Nentwig, S. Olenin, V. Panov, J. Pergl, P. Pyšek, A. Roques, D. Sol, W. Solarz & M. Vilà, 2008. Grasping at the routes of biological invasions: a framework for integrating pathways into policy. *Journal of Applied Ecology* 45: 403–414.
- James, J., F. Slater & J. Cable, 2014. A.L.I.E.N. databases: addressing the lack in establishment of non-natives databases. *Crustaceana* 87: 1192–1199.
- James, J., F. M. Slater, I. P. Vaughan, K. A. Young & J. Cable, 2015a. Comparing the ecological impacts of native and invasive crayfish: could native species' translocation do more harm than good? *Oecologia* 178: 309–316.
- James, J., J. Cable, G. Richardson, K. E. Davidson & A. S. Y. Mackie, 2015b. Two alien species of Branchiobdellida (Annelida: Clitellata) new to the British Isles: a morphological and molecular study. *Aquatic Invasions* 10: 371–383.
- James, J., K. E. Davidson, G. Richardson, C. Grimstead & J. Cable, 2015c. Reduced aggression and foraging efficiency of invasive signal crayfish (*Pacifastacus leniusculus*) infested with non-native branchiobdellidans (Annelida: Clitellata). *Parasites & Vectors* 8: 596.
- James, J., J. R. Thomas, A. Ellis, K. A. Young, J. England & J. Cable, 2015d. Over-invasion in a freshwater ecosystem: newly introduced virile crayfish (*Orconectes virilis*) out-compete established invasive signal crayfish (*Pacifastacus leniusculus*). *Marine and Freshwater Behaviour and Physiology* 49: 9–18.
- James, J., K. E. Davidson, R. Hunt & J. Cable, 2017. Assessing the invasion potential of non-native branchiobdellidans: experimental studies of survival, reproduction and competition. *Knowledge & Management of Aquatic Ecosystems* 418: 35.
- Johnson, L. E. & D. K. Padilla, 1996. Geographic spread of exotic species: ecological lessons and opportunities from the invasion of the zebra mussel *Dreissena polymorpha*. *Biological Conservation* 78: 23–33.

- Keesing, F., R. D. Holt & R. S. Ostfeld, 2006. Effects of species diversity on disease risk. *Ecology Letters* 9: 485–498.
- Kouba, A., A. Petrusek & P. Kozák, 2014. Continental-wide distribution of crayfish species in Europe: update and maps. *Knowledge and Management of Aquatic Ecosystems* 413: 05.
- Lee, J. H., T. W. Kim & J. C. Choe, 2009. Commensalism or mutualism: conditional outcomes in a branchiobdellid–crayfish symbiosis. *Oecologia* 159: 217–224.
- Leslie, A. J., C. J. Lovely & J. M. Pittman, 2011. A preliminary disease survey in the wild Nile crocodile (*Crocodylus niloticus*) population in the Okavango Delta, Botswana. *Journal of the South African Veterinary Association* 82: 155–159.
- Lodge, D. M., C. A. Taylor, D. M. Holdich & J. Skurdal, 2000. Nonindigenous crayfishes threaten North American freshwater biodiversity: lessons from Europe. *Fisheries* 25: 7–20.
- Lutz, C. G. & W. R. Wolters, 1999. Mixed model estimation of genetic and environmental correlations in red swamp crawfish *Procambarus clarkii* (Girard). *Aquaculture Research* 30: 153–163.
- Lymbery, A. J., M. Morine, H. G. Kanani, S. J. Beatty & D. L. Morgan, 2014. Co-invaders: the effects of alien parasites on native hosts. *International Journal for Parasitology: Parasites and Wildlife* 3: 171–177.
- Mack, R. N., D. Simberloff, W. Mark Lonsdale, H. Evans, M. Clout & F. A. Bazzaz, 2000. Biotic invasions: causes, epidemiology, global consequences, and control. *Ecological Applications* 10: 689–710.
- Manchester, S. J. & J. M. Bullock, 2000. The impacts of non-native species on UK biodiversity and the effectiveness of control. *Journal of Applied Ecology* 37: 845–864.
- Manus, L. R. M., 1960. Some ecological studies of the Branchiobdellidae (Oligochaeta). *Transactions of the American Microscopical Society* 79: 420–428.
- Marques, M., F. Banha, M. Águas & P. Anastácio, 2014. Environmental cues during overland dispersal by three freshwater invaders: *Eriocheir sinensis*, *Pacifastacus leniusculus*, and *Procambarus clarkii* (Crustacea, Decapoda). *Hydrobiologia* 742: 81–93.
- Mohammed, R. S., M. Reynolds, J. James, C. Williams, A. Mohammed, A. Ramsubhag, C. van Oosterhout & J. Cable, 2016. Getting into hot water: sick guppies frequent warmer thermal conditions. *Oecologia* 181: 911–917.
- Müller, C. B. & P. Schmid-Hempel, 1993. Exploitation of cold temperature as defence against parasitoids in bumblebees. *Nature* 363: 65–67.
- Niwa, N., M. V. Archdale, T. Matsuoka, A. Kawamoto & H. Nishiyama, 2014. Microhabitat distribution and behaviour of Branchiobdellidan *Holtodrilus truncatus* found on the freshwater shrimp *Neocaridina* spp. from the Sugo River. *Japan. Central European Journal of Biology* 9: 80–85.
- Ohtaka, A., S. R. Gelder, T. Kawai, K. Saito, K. Nakata & M. Nishino, 2005. New records and distributions of two North American branchiobdellidan species (Annelida: Clitellata) from introduced signal crayfish, *Pacifastacus leniusculus*, in Japan. *Biological Invasions* 7: 149–156.
- Peterson, A. C., K. L. D. Richgels, P. T. J. Johnson & V. J. McKenzie, 2013. Investigating the dispersal routes used by an invasive amphibian, *Lithobates catesbeianus*, in human-dominated landscapes. *Biological Invasions* 15: 2179–2191.
- Puky, M., 2014. Invasive crayfish on land: *Orconectes limosus* (Rafinesque, 1817) (Decapoda: Cambaridae) crossed a terrestrial barrier to move from a side arm into the Danube River at Szeremle, Hungary. *Acta Zoologica Bulgarica* 57: 143–146.
- Rahel, F. J. & J. D. Olden, 2008. Assessing the effects of climate change on aquatic invasive species. *Conservation Biology* 22: 521–533.
- Ramalho, R. O. & P. M. Anastácio, 2014. Factors inducing overland movement of invasive crayfish (*Procambarus clarkii*). *Hydrobiologia* 746: 135–146.
- Ranta, E. & K. Lindström, 1993. Body size and shelter possession in mature signal crayfish, *Pacifastacus leniusculus*. *Annales Zoologici Fennici* 30: 125–132.
- R Core Team, 2016. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.
- Ronce, O., 2007. How does it feel to be like a rolling stone? Ten questions about dispersal evolution. *Annual Review of Ecology, Evolution, and Systematics* 38: 231–253.
- Sayer, M. D. J. & J. Davenport, 1991. Amphibious fish: why do they leave water? *Reviews in Fish Biology and Fisheries* 1: 159–181.
- Skaug, H., D. Fournier, B. Bolker, A. Magnusson & A. Nielsen, 2016. Generalized Linear Mixed Models using ‘AD Model Builder’. R package version 0.8.3.3.
- Skelton, J., R. P. Creed & B. L. Brown, 2014. Ontogenetic shift in host tolerance controls initiation of a cleaning symbiosis. *Oikos* 123: 677–686.
- Sobecka, E., G. Hajek & Ł. Skorupiński, 2011. Four pathogens found associated with *Eriocheir sinensis* H. Milne-Edwards, 1853 (Crustacea: Brachyura: Grapsidae) from Lake Dąbie (Poland). *Oceanological and Hydrobiological Studies* 40: 96–99.
- Strayer, D. L., 2010. Alien species in fresh waters: ecological effects, interactions with other stressors, and prospects for the future. *Freshwater Biology* 55: 152–174.
- Svoboda, J., A. Mrugała, E. Kozubíková-Balcarová & A. Petrusek, 2016. Hosts and transmission of the crayfish plague pathogen *Aphanomyces astaci*: a review. *Journal of Fish Diseases* 40: 127–140.
- Thomas, M. J., R. P. Creed, J. Skelton & B. L. Brown, 2016. Ontogenetic shifts in a freshwater cleaning symbiosis: consequences for hosts and their symbionts. *Ecology* 97: 1507–1517.
- Vaughn, L. K., H. A. Bernheim & M. J. Kluger, 1974. Fever in the lizard *Dipsosaurus dorsalis*. *Nature* 252: 473–474.
- Vedia, I., J. Oscoz, J. Rueda, R. Miranda, E. M. García-Roger, E. Baquero & S. R. Gelder, 2014. An alien ectosymbiotic branchiobdellidan (Annelida: Clitellata) adopting exotic crayfish: a biological co-invasion with unpredictable consequences. *Inland Waters* 5: 89–92.
- Yoder, J. A., J. A. Stoeckel, B. S. Helms, A. L. Lorenz & A. J. Jajack, 2016. Hygric stresses and strategies in maintaining the association between crayfish and ectosymbiotic worms across vastly different environments. *Symbiosis* 69: 141–150.
- Young, W., 1966. Ecological studies of the Branchiobdellidae (Oligochaeta). *Ecology* 47: 571–578.