

## Germination and anchorage of *Enteromorpha* spp. in sediments of the Wadden Sea

D. Schories & K. Reise

*Biologische Anstalt Helgoland, Wattenmeerstation Sylt; D-25992 List,  
Federal Republic of Germany*

**ABSTRACT:** Large quantities of filamentous green algae (*Enteromorpha* spp.) have regularly occurred on muddy and sandy tidal flats in Königshafen, on the island of Sylt (North Sea), since 1979 – covering the sediments in thick mats during the summer months. While spores of *Enteromorpha* were encountered in both mud and sand, germling formation was restricted to sand. However, mud snails (*Hydrobia ulvae* Pennant) were overgrown with small *Enteromorpha* filaments in both habitats, about 50 % of them at a muddy site and 20 % at a sandy one. Filaments, several cm in length and still adhering to the snails, became tangled into clusters. At the sandy site, with abundant *Arenicola marina* L., these clusters slid into the feeding funnels of lugworm burrows; the importance of this secondary anchorage is demonstrated by a field experiment. We suggest that the primary and secondary attachment of *Enteromorpha* filaments provided by benthic fauna is an essential step in the development of green algal mats on sedimentary tidal flats.

### INTRODUCTION

In earlier decades of this century, macroscopic green algae were rarely present on the sandy or muddy intertidal flats near the island of Sylt (Kuckkuck, 1896–1903 unpubl; Nienburg, 1927; Wohlenberg, 1937; Kornmann, 1952), but have regularly occurred in thick mats since 1979 (Reise, 1983; Reise et al., 1989). Within the last 25 years, excessive growth of green macroalgae has become an increasingly common phenomenon and a problem in sheltered bays (Sawyer, 1965; Buttermoore, 1977; Fitzgerald, 1978; Montgomery & Soulsby, 1981; Sfriso et al., 1987; Raffaelli et al., 1989). On tidal flats near the island of Sylt, several species of green algae, namely *Enteromorpha* spp., *Cladophora* spp., *Chaetomorpha sutoria* (Berkeley) Kornmann and *Ulva* spp. form dense mats during summer, which lead to dramatic changes on epi- and endofaunal abundance (Reise, 1983; Schories, 1991).

Eutrophication in estuaries and coastal waters is well-documented and may explain the extensive growth of certain macroalgae that take advantage of these conditions (Soulsby et al., 1982, 1985; Sfriso et al., 1987). Culture experiments with *Enteromorpha* spores demonstrated that the addition of sewage effluent to unpolluted seawater stimulates growth (Ford et al., 1983). Propagules of *Enteromorpha* species seem to germinate on any kind of solid substratum, including other plants. However, little is known about how *Enteromorpha* propagules can develop in muddy and sandy intertidal areas where tidal currents and wave action move the surface sediment to and fro. Linke

(1939) and Nienhuis (1970) mention that mollusc shells, dead or alive (i.e. *Cerastoderma edule* L.), are colonized by these algae, implying that algal cover of an intertidal area depends on shell abundance. More detailed information about possible substrata of *Enteromorpha* species occurring in the Wadden Sea is given by Koeman & van den Hoek (1982a, b, 1984).

Nienburg (1927) described the direct development of *Enteromorpha* germlings on sand grains, but he did not mention whether these germlings developed further into adult plants. Reise (1983) described how germlings of *Enteromorpha* spp., attached to sand grains, increased in length during summer, branched, and finally became plaited into tresses by tidal currents and wave action. Observations showed that these strands of green algae became anchored in the feeding funnels of the lugworm *Arenicola marina*, and hence resisted displacement by tidal currents.

The germling formation of *Enteromorpha* spp. in sandy and muddy intertidal areas is still a paradox. In this study, we describe the development of young *Enteromorpha* spp. stages on intertidal soft bottoms in more detail. We propose that benthic fauna can play an essential role in *Enteromorpha* development by providing primary and secondary anchorage.

## AREA AND METHODS

### Habitat

Investigations were conducted in Königshafen, a shallow tidal bay on the island of Sylt in the Northern Wadden Sea. Hydrography and macrofauna have been described by Wohlenberg (1937) and Reise (1985), macroalgae by Nienburg (1927) and Kornmann (1952), sediment by Austen (1990). 76 % of Königshafen Bay (total 4.8 km<sup>2</sup>) consists of sandy tidal areas and 9 % of muddy areas. Two study sites where algal cover develops during summer were investigated. One site ("Tonnenlegerbucht") is located in a narrow embayment between the dike and a sandy spit with predominantly muddy sand (organic content: 1.2 % of dry weight; silt content > 10 %; median of particle size: 406 µm). The relatively large median of sediment particles on the mud flat is explained by the vicinity of the sandy spit containing coarse sediment which, during windy days, drifts onto the flat. The location is at mid-tide level, highly sheltered, with a domestic sewage effluent that enters nearby. The other site ("Möwenberg-Watt") is located on an extensive sandy tidal flat at mid-tide level (organic content: 0.04 % of dry weight; silt content < 4 %; median of particle size: 324 µm).

### Sampling

7 samples of 5.3 cm<sup>2</sup>/0–0.5 cm depth were taken regularly from each site, to estimate the content of young *Enteromorpha* filaments (> 0.1 mm length) growing attached to sand grains. The sediment cores were transferred to Petri dishes and inspected under 50-fold magnification. In addition, samples of *Hydrobia ulvae* (6 samples of 100 cm<sup>2</sup> sediment from each site, sieved with a mesh of 1000 µm) were taken, and the number of snails with and without attached *Enteromorpha* germlings determined. Sampling was initiated on May 1st 1990 and continued until the end of August 1990. In the same year, green algal biomass was determined by taking algae from 6 plots (sampling size varied between 0.04 and 1 m<sup>2</sup> depending on green algal cover of the area) along permanent

transects at each site. The plants were washed in fresh water to remove the sediment, epifauna and salt from the algal filaments and then oven-dried for 72 h at 70°C.

Eighteen clusters of drifting, young *Enteromorpha* filaments, taken at random, were analysed in the middle of May 1992 at the same sites. *Enteromorpha* dry weight varied between 0.064 and 0.226 g cluster<sup>-1</sup>. The relation of sand grains to mudsnails as substrata of attachment was also recorded for these clusters of filaments.

### Experiment

At the sandy site, experimental removal of the lugworm *Arenicola marina* ( $144 \pm 25$  individuals m<sup>-2</sup>, SD) took place by inserting a sheet of gauze (mesh size 1 mm) horizontally into the sediment at 5 cm depth. The upper layer of sediment from fourteen 1-m<sup>2</sup> plots was removed with a shovel, the gauze unfolded and then covered with original surface sediment. Immediately afterwards the gauze was removed from the 7 plots which served as controls. The remaining 7 plots' gauze blocked the vertical shafts of lugworms, forcing them to leave sideways, while on the similarly disturbed controls they remained in their burrows. A pilot experiment showed that green algae never attached to the buried gauze. Thus, there was no need to implant a substitute for the gauze at the control plots.

The experiment was initiated on May 1st, and terminated on July 14th 1990. The effect of green algal cover was evaluated by taking photographs at regular intervals, and, with the help of a grid (4900 units), the surface cover of green algae on each plot was estimated. Algal cover from control and removal sites were compared with non-parametric tests (U-test from Wilcoxon, binomial sign test [Sachs, 1984]).

## RESULTS

### *Enteromorpha* germlings

Young plants of various macroscopic green algal genera (for example: *Cladophora*, *Monostroma*, *Percursaria*, *Rhizoclonium*, *Ulva*) were present at both sites, but remained rare in the samples. At no time during sampling were more than  $1.9 \pm 2.5$  percent of mudsnails at the muddy site overgrown with young plants of any genera mentioned above. Only species of the genus *Enteromorpha* were frequent. At the muddy site, a maximum of 50 small *Enteromorpha* filaments overgrew a single *Hydrobia* shell, but presumably there were still more spores and small germlings attached. At the sandy site, the abundance of *Enteromorpha* germlings growing on shells or sand grains was always much lower. There we counted a maximum of 18 *Enteromorpha* filaments growing on a single shell, and a maximum of 9 filaments on a single sand grain.

We could show that in May 1990 spores of *Enteromorpha* were attached to sand grains at each site. They were able to develop into young filaments when we kept them for a week in Petri dishes with sterilized seawater enriched with Erdschreiber medium. In the field, germination of spores attached to sand grains only occurred at the sandy site, and not the muddy site. On the other hand, development of spores on shells of the "mobile" substratum *Hydrobia ulvae* occurred at both sites. In contrast to the sandy site,

empty shells of mudsnails (with the exception of snails that had recently died) were never overgrown with *Enteromorpha* at the muddy site.

The highest numbers of snails or sand grains overgrown with *Enteromorpha* filaments were found in May at both sites (Fig. 1). At the sandy site, up to  $26\,000 \pm 14\,000$  sand grains  $\text{m}^{-2}$  were found overgrown with *Enteromorpha* filaments, respectively  $5700 \pm 5600$  of the snail shells  $\text{m}^{-2}$  were attached to *Enteromorpha* filaments. At the muddy site, none of the sand grains, but  $20\,300 \pm 5100$  snail shells  $\text{m}^{-2}$  were found with filaments.

The total abundance of *Enteromorpha* germlings varied in relation to the number of mudsnails present. *Hydrobia* densities  $\text{m}^{-2}$  were in the same magnitude at both sites:  $31\,000 \pm 4700$  to  $66\,700 \pm 31\,000$  at the muddy site, and  $24\,000 \pm 5000$  to  $52\,000 \pm 15\,000$  at the sandy site. Recruits of 1990 were not considered. The percentage of overgrown *Hydrobia* was much higher at the muddy site than at the sandy site (Fig. 2). Mean snail size at the muddy site was 3.3 mm in June and 3.9 mm in September; at the sandy site, the size was 2.7 mm and 3.0 mm, respectively. There was no relation between shell size and degree of colonization by *Enteromorpha* filaments.

### Clusters of *Enteromorpha* filaments

Throughout the summer, the filaments of *Enteromorpha* spp., attached to sand grains or *Hydrobia* shells, increased in length, but these were rarely present in our sediment or snail samples. In the latter two, most *Enteromorpha* filaments were very small (< 1 mm). Only at sheltered, muddy sites, did we occasionally observe filaments of up to 20 cm length attached to a mudsnail on the sediment surface. We assume that filaments > 1 mm easily become dislocated, together with their substratum (snail or sand grain), when subject to tidal currents. From the beginning of May onwards, we found clusters of them during low tide on the sediment surface. These clusters may be generated (1) by snails with attached filaments, which crawl on the sediment surface, thereby passively entangling other filaments with or without their substratum, (2) by tidal currents and waves which lift filaments together with the sand grains or snails up from the bottom into the water column where they become braided to clusters and tresses.

Similar to the results from sediment samples, algal filaments within drifting clusters were exclusively attached to living mudsnails at the muddy site, and to both snails and sand grains at the sandy site (Fig. 3). However, the proportion of snails to sand grains as substrata of attachment increased from sediment cores to clusters of algae at the sandy site, indicating a higher chance of snail-attached filaments becoming entangled in clusters.

### Green algal biomass

Extensive growth of *Enteromorpha* spp. was recorded from May to the end of August 1990 in Königshafen bay. Not only sand grains and mudsnails served *Enteromorpha* as primary substrata, but also polychaete tubes (i.e. built by *Lanice conchilega* Pallas) and cockles became overgrown, but this did not develop into a high algal cover in the bay. Only in one restricted area ( $5600 \text{ m}^2$ ) of it were *Enteromorpha* filaments (growing on the sandtubes of *Lanice*) able to form a dense mat ( $97 \pm 43 \text{ g dw m}^{-2}$ ; a total of 6 samples of  $0.25 \text{ m}^2$  were randomly taken) in August. At the muddy site, the algal mat was thickest in

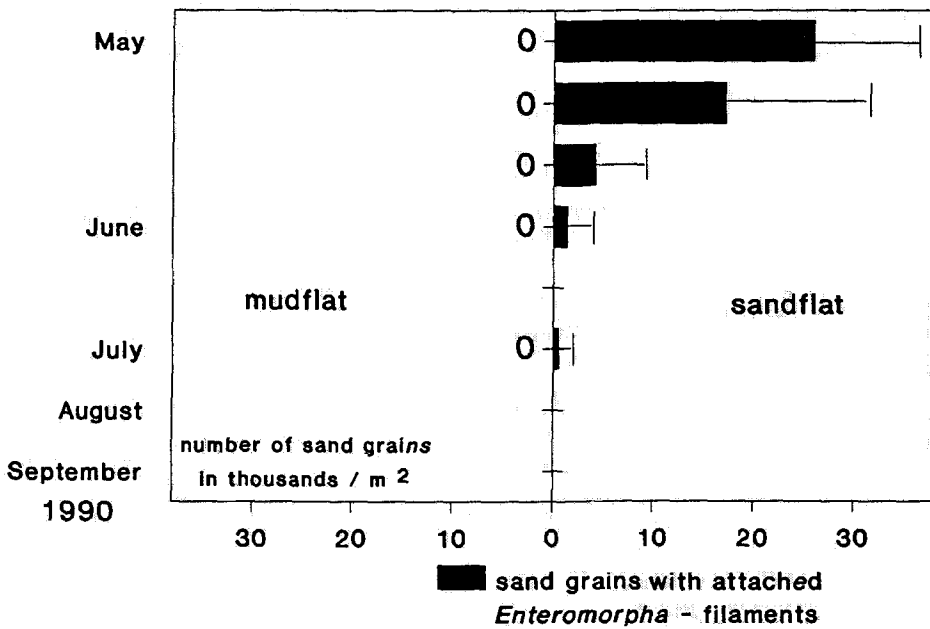
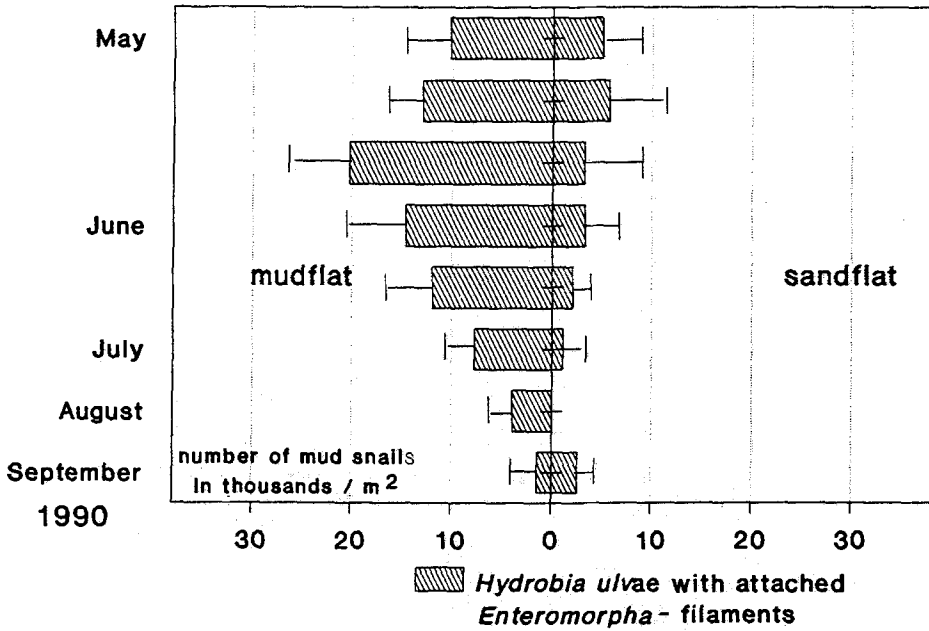


Fig. 1. Abundance (means  $\pm$  1SD) of sand grains (above) and *Hydrobia ulvae* (below) with attached *Enteromorpha* filaments at a muddy and a sandy site during summer 1990

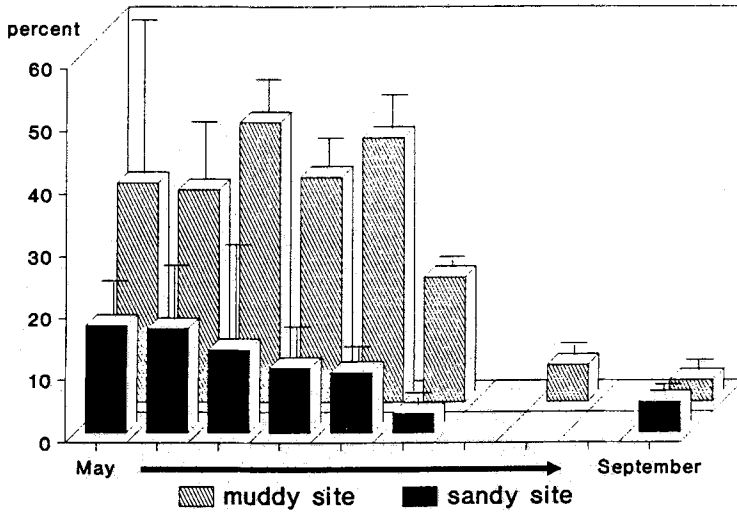


Fig. 2. Percentages of *Hydrobia ulvae* in mud and sand with attached *Enteromorpha* filaments during summer 1990 (means  $\pm$  1SD). The total number of snails found in each sample constitutes 100 %

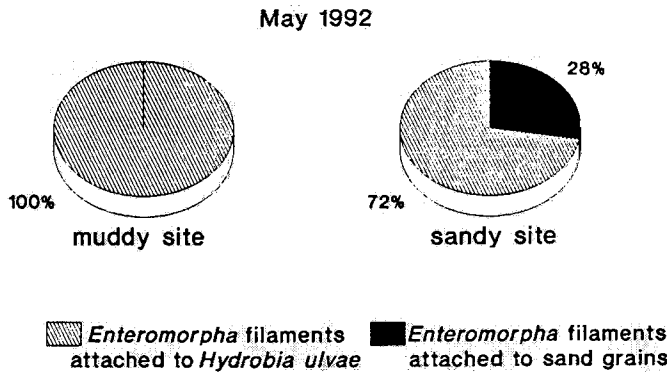


Fig. 3. At the muddy site, shells of living mudsnails, *Hydrobia ulvae*, served as the only substratum within clusters of young *Enteromorpha* filaments, whereas at the sandy site, sand grains were also overgrown by these algae. At both sites, substrata other than mudsnails or sand grains were not present within the clusters

June, when *Enteromorpha* biomass attained  $318 \pm 89 \text{ g dw m}^{-2}$ . There was no dense algal mat present at the sandy site. The surface cover with *Enteromorpha* tresses attained a maximum of  $42 \pm 18 \%$  (estimated from the same plots from which the algal samples were taken). This corresponds to an algal biomass of  $60 \pm 48 \text{ g dw m}^{-2}$ .

## Anchorage of drifting green algae

During ebb tide, drifting green algae become deposited in depressions of the sediment surface. On the sand flats in the Wadden Sea, these depressions are mostly funnels of lugworm burrows. Some of the algal filaments slide down the feeding funnels together with their attached substrata (sand grains and mudsnails). At the sandy site, we found *Enteromorpha* tresses vertically embedded in the sediment down to the depth of burrows. There were more algal tresses embedded in the sediment than there were lugworm feeding funnels, and only a few were actually encountered sticking out of functional head shafts. From this, it follows that lugworms abandon head shafts and funnels once these are clogged with algal tresses and must build new shafts and funnels in a new position. The consequence of this is that a single lugworm may anchor several drifting algal clusters and prevent them from further drifting. The result of this process is revealed by the lugworm exclusion experiment at the sandy site (Fig. 4). Algal cover

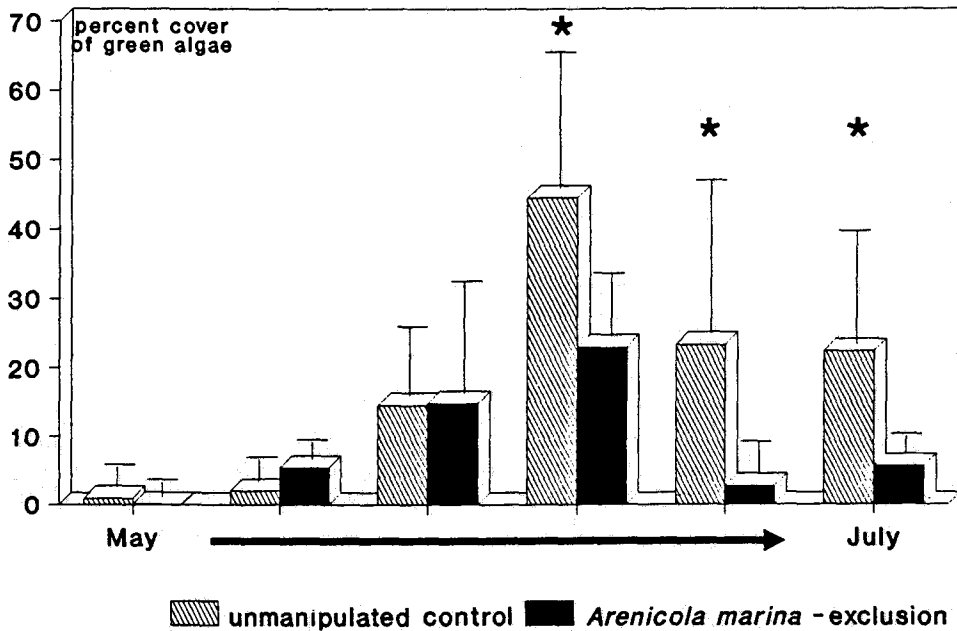


Fig. 4. Experimental exclusion of the lugworm *Arenicola marina* leads to a lower coverage with *Enteromorpha* spp. filaments during the growing season. Asterisks (\*) denote differences at the level of  $p < 0.10$  (Mann-Whitney U-test)

became significantly lower on plots without lugworms, compared with control plots with ambient lugworm densities. During the time of investigation, the number of lugworm casts in the controls did not change.

## DISCUSSION

The enhancement of green algal biomass on sandy and muddy bottoms of the shallow sea is generally ascribed to high nutrient loads in coastal waters, originating from

anthropogenic sources (Harlin & Thorne-Miller, 1981; Kautsky, 1982; Sfriso et al., 1987). High nutrient uptake rates and a storage of excess nutrients support an extensive growth of dense algal mats (Kautsky, 1982; O'Brien & Wheeler, 1987; Thomas & Harrison 1987). However, the occurrence of these green algae on soft bottoms is still a paradox, because one would have expected to find a firm substratum on the surface to allow for germination under suitable light conditions. Tidal currents and waves are likely to carry those thalli away which are not attached to solid objects. The present study provides some evidence of how this paradox is overcome by species of *Enteromorpha*.

We found abundant spores in both mud and sand of the intertidal zone. Germling formation occurred on sand grains of the surface layer, as described by Nienburg (1927), but only at a sandy site and not in an area with mud. Dayton (1975) showed that the total absence of *Hedophyllum* plants from a site where transplanted adults survived, was very likely due to siltation of early post-settlement stages. Moss et al. (1973) suggested that silt was a major factor inhibiting colonisation of *Himantalia*, and attributed this to reduced light preventing germination. Norton (1978) showed that an overlying layer of silt reduced light by 98 % and inhibited development of microscopic gametophytes. We also suggest that reduced light conditions on sand grains enveloped by silt were the main factor which inhibited the germling formation of *Enteromorpha* spp. at our muddy site. A consequence of this would be the absence of young *Enteromorpha* plants at muddy sites.

However, this is not the case, because hydrobiid snails provide an alternative substratum for germination. *Hydrobia ulvae* is a highly abundant component of the intertidal fauna on both mud and sand in the Wadden Sea (Linke, 1939; Reise, 1987), and elsewhere along European coasts (Walters & Wharfe, 1980). In the muddy area investigated in this study, *H. ulvae* turned out to be the only substratum suitable for germling formation of *Enteromorpha*. Thus, this snail may attain a key role in the germling formation, and later on in the development of green algae, on muddy bottoms in sheltered bays.

At the sandy site investigated, the relation of overgrown mud snails to sand grains changed from spring to summer. *H. ulvae* became more important as a substratum. This may indicate a disadvantage for green algal filaments attached to sand grains, once they have grown beyond a certain length. Presumably they become more easily dislocated and are taken away by the tidal currents, while filaments attached to hydrobiid snails remain. This suggests that also on sand flats *H. ulvae* may attain a key role in *Enteromorpha* development. Present investigations in Königshafen bay show that spores or microscopic stages of *Enteromorpha* spp. overwinter in mud and sand attached to the shells of *H. ulvae*. Nearly all mud snails were stocked with early post-settlement stages (Schories, in prep.).

We do not suggest a case of mutualism. In Petri dishes we observed *H. ulvae* feeding on spores of *Enteromorpha* spp. However, there was no evidence of *H. ulvae* grazing on young filaments from each other's houses or grazing on adult plants. Feeding experiments led Morrisey (1988) to conclude that *Enteromorpha* is not a preferred food source for *H. ulvae*. Nevertheless, *H. ulvae* may be highly abundant within algal mats (Nicholls et al., 1981; Soulsby et al., 1982). In none of our samplings from *Enteromorpha* clusters or algal mats did we find high numbers of empty *Hydrobia* shells or snails that had recently died with long *Enteromorpha* filaments. Snails attached to *Enteromorpha* filaments and entangled in clusters or mats were alive and engaged in epiphytic browsing, probably



feeding on microalgae and bacteria. Thus, the effects of green algae on *H. ulvae* may at best be neutral, and at times certainly negative, when algae become imbedded in the sediment, are washed ashore, or decay under anoxic conditions.

Some questions of the development of green algae on sediments were not considered in this study. It remains to be proven that those filaments found attached to snails are the same which later form thick algal mats. Overwintering filaments of *Enteromorpha* were found at the muddy site. These may also contribute to the formation of algal mats. At the sandy site, no overwintering filaments were found. Drift algae, originating outside the study area, may also contribute to the green algal mats observed.

*Enteromorpha* spp. are not only epizooic on *H. ulvae* on the tidal flats near Sylt. In particular, cockles *Cerastoderma edule* (L.) are often bearded at their siphonal end with *Enteromorpha* spp. In 1988 and 1989, dense cockle beds became entirely overgrown by green algae attached to cockle shells. Most of these cockles suffocated and died. On another site in Königshafen, *Enteromorpha* spp. germinated on the tube-caps of the abundant polychaete *Lanice conchilega* (Pallas), and subsequently developed into a coherent algal cover, until storms in August removed most of it (1990 and 1991).

On the sheltered mud flat investigated, *H. ulvae* provided a sufficient substratum for germination and attachment, and a thick mat of *Enteromorpha* filaments persisted from June to August. At the other site (the moderately-exposed sand flat), *H. ulvae* was suitable for attachment but did not provide sufficient anchorage to resist the water movements. As observed earlier by Reise (1983), and confirmed here by means of a field experiment, burrows of the lugworm *Arenicola marina* provide a secondary anchorage. Because of the widespread occurrence of *A. marina* on European tidal flats, we suggest this mode of anchorage to be of general importance in the development of green algal mats.

The algal strands sliding into feeding funnels are clearly a nuisance to the lugworms. This is evident from the fact that worms shift head-shafts and funnels into new positions, once the former are clogged by algae. This very fact amplifies the importance of lugworms for the algae, because in this way a single worm may anchor several algal strands. In the study area, we observed thick algal mats which had apparently been made possible by this mode of anchorage. When these mats of green algae start to decay, lugworms were observed abandoning the sediment. We found no evidence that lugworms feed on the green algae. Nor do fragments of green algae seem to be a suitable food for lugworms (Hylleberg, 1975; Rijken, 1979).

This is in contrast to nereid polychaetes. Woodin (1977) describes how *Nereis vexillosa* and *Platynereis bicanaliculata* attach pieces of drift algae to their tube-caps and subsequently feed on the growing thalli. She termed this "algal gardening". Another nereid worm, *N. brandti*, pulls large strands of green algae into its burrow to feed on them (Hylleberg & Henriksen, 1980). We observed *N. diversicolor* likewise pulling filaments of *Enteromorpha* into its burrow for subsequent consumption. Other polychaetes (Nereidae, Onuphidae) are known to decorate their tubes with drift algae (Pettiborne, 1963; Daly, 1973).

In the present study, we have described two cases where benthic invertebrates promote green algal development without receiving any clear benefit in return. The importance of *H. ulvae* as a primary, and *A. marina* as a secondary, means of attachment only became apparent with the advent of coastal eutrophication. In the absence of eutrophication as a precondition to the excessive growth of green algae, this non-trophic

link in the ecological web would be a curious coincidence, lacking any general significance.

*Acknowledgements.* We gratefully acknowledge the help of A. Albrecht. The investigation was part of the project "Grünalgenausbreitung im Wattenmeer (FKZ: 102 04 245)", supported by the Umweltbundesamt. It was also supported by the Ministry of Research and Technology (Publication No. 47 of the Wadden Sea Ecosystem Project).

#### LITERATURE CITED

- Austen, I., 1990. Geologisch-sedimentologische Kartierung des Königshafens (List auf Sylt) und Untersuchung seiner Sedimente. Dipl.-Arb., Univ. Kiel, 99 pp.
- Buttermoore, R. E., 1977. Eutrophication of an impounded estuarine lagoon. – Mar. Pollut. Bull. 8, 13–15.
- Daly, J. M., 1973. Behavioural and secretory activity during tube construction by *Platynereis dumerilii* Aud. and M. Edw. (Polychaeta: Nereidae). – J. mar. biol. Ass. U.K. 53, 521–529.
- Dayton, P. K., 1975. Experimental evaluation of ecological dominance in a rocky intertidal algal community. – Ecol. Monogr. 45, 137–159.
- FitzGerald, Jr., W. J., 1978. Environmental parameters influencing the growth of *Enteromorpha clathrata* (Roth) J. Ag. in the intertidal zone on Guam. – Botanica mar. 21, 207–220.
- Ford, G. S., Rees, R. L. G., Soulsby, P. G. & Lowthion, D., 1983. Nutrient removal trials and bioassay evaluation at Budds Farm Sewage Treatment Works, Havant. – Wat. Pollut. Contr. 82, 381–392.
- Harlin, M. M. & Thorne-Miller, B., 1981. Nutrient enrichment of seagrass beds in a Rhode Island coastal lagoon. – Mar. Biol. 65, 221–229.
- Hylleberg, J., 1975. Selective feeding of *Abarenicola pacifica* with notes on *Abarenicola vagabunda* and a concept of gardening in lugworms. – Ophelia 14, 113–137.
- Hylleberg, J. K. & Henriksen, K., 1980. The central role of bioturbation in sediment mineralization and element re-cycling. – Ophelia (Suppl.) 1, 1–16.
- Kautsky, L., 1982. Primary production and uptake kinetics of ammonium and phosphate by *Enteromorpha compressa* in an ammonium sulfate industry outlet area. – Aquat. Bot. 12, 23–40.
- Koeman, R. P. T. & Hoek, C. van den, 1982a. The taxonomy of *Enteromorpha* Link, 1820, (Chlorophyceae) in the Netherlands. 1. The section *Enteromorpha*. – Arch. Hydrobiol. (Suppl.) 63, 279–330.
- Koeman, R. P. T. & Hoek, C. van den, 1982b. The taxonomy of *Enteromorpha* Link, 1820, (Chlorophyceae) in the Netherlands. 2. The section Proliferae. – Cryptogamie (Algol.) 3, 37–70.
- Koeman, R. P. T. & Hoek, C. van den, 1984. The taxonomy of *Enteromorpha* Link, 1820 (Chlorophyceae) in the Netherlands. 3. The sections Flexuosae and Clathratae, and an addition to the section Proliferae. – Cryptogamie (Algol.) 5, 21–61.
- Kornmann, P., 1952. Die Algenvegetation von List auf Sylt. – Helgoländer wiss. Meeresunters. 4, 55–61.
- Linke, O., 1939. Die Biota des Jadebusen-Wattes. – Helgoländer wiss. Meeresunters. 1, 201–348.
- Montgomery, H. A. C. & Soulsby, P. G., 1981. Effects of eutrophication on the intertidal ecology of Langstone Harbour, U. K., and proposed control measures. – Prog. Wat. Tech. 13, 87–294.
- Morrisey, D. J., 1988. Differences in effects of grazing by deposit-feeders *Hydrobia ulvae* (Pennant) (Gastropoda: Prosobranchia) and *Corophium arenarium* Crawford (Amphipoda) on sediment microalgal populations. I. Qualitative aspects. – J. exp. mar. Biol. Ecol. 118, 333–342.
- Moss, B. L., Mercer, S. & Shearer, A., 1973. Factors affecting the distribution of *Himantalia elongata* (L.) S. F. Gray on the north-east coast of England. – Estuar. coast. mar. Sci. 1, 233–243.
- Nicholls, D. J., Tubbs, C. R. & Haynes, F. N., 1981. The effect of green algal mats on intertidal macrobenthic communities and their predators. – Kieler. Meeresforsch. (Sonderh.) 5, 511–520.
- Nienburg, W., 1927. Zur Ökologie der Flora des Wattenmeeres. 1. Teil. Der Königshafen bei List auf Sylt. – Wiss. Meeresunters. (Abt. Kiel) 20, 146–196.
- Nienhuis, P. H., 1970. The benthic algal communities of flats and salt marshes in the Grevelingen, a sea-arm in the south-western Netherlands. – Neth. J. Sea Res. 5, 20–49.

- Norton, T. A., 1978. The factors influencing the distribution of *Saccorhiza polyschides* in the region of Lough Line. – J. mar. biol. Ass. U. K. 58, 527–536.
- O'Brien, M. C. & Wheeler, P. A., 1987. Short-term uptake of nutrients by *Enteromorpha prolifera* (Chlorophyceae). – J. Phycol. 23, 547–566.
- Pettibone, M. H., 1963. Marine polychaete worms of the New England region. Part 1. – Bull. U.S. natn. Mus. 227, 1–356.
- Raffaelli, D., Hull, S. & Milne, H., 1989. Long-term changes in nutrients, weed mats and shore birds in an estuarine system. – Cah. Biol. mar. 30, 259–270.
- Reise, K., 1983. Sewage, green algal mats anchored by lugworms, and the effects on Turbellaria and small Polychaeta. – Helgoländer Meeresunters. 36, 151–162.
- Reise, K., 1985. Tidal flat ecology. Springer, Heidelberg, 191 pp.
- Reise, K., 1987. Distribution and abundance of small and juvenile macrofauna on the tidal flats in the Wadden Sea. In: Proceedings of the 5th International Wadden Sea Symposium. Ed. by S. Tougaard & S. Asbirk. The National Forest and Nature Agency & the Museum of Fisheries and Shipping, Esbjerg, 7–25.
- Reise, K., Herre, E. & Sturm, M., 1989. Historical changes in the benthos of the Wadden Sea around the island of Sylt in the North Sea. – Helgoländer Meeresunters. 43, 417–433.
- Rijken, M., 1979. Food and food uptake in *Arenicola marina*. – Neth. J. Sea Res. 13, 406–421.
- Sachs, L., 1984. Angewandte Statistik. Springer, Berlin, 522 pp.
- Sawyer, N., 1965. The sea lettuce problem in Boston harbour. – J. Wat. Pollut. Contr. Fed. 37, 1122–1133.
- Schories, D., 1991. Wechselwirkungen zwischen Grünalgen und Bodenfauna im Wattenmeer. Dipl.-Arb., Univ. Kiel, 95 pp.
- Sfriso, A., Marcomini, A. & Pavoni, B., 1987. Relationships between macroalgal biomass and nutrient concentrations in a hypertrophic area of the Venice Lagoon. – Mar. env. Res. 22, 297–312.
- Soulsby, P. G., Lowthion, D. & Houston, M., 1982. Effects of macroalgal mats on the ecology of intertidal mudflats. – Mar. Pollut. Bull. 13, 162–166.
- Soulsby, P. G., Lowthion, D., Houston, M. & Montgomery, H. A. C., 1985. The role of sewage effluent in the accumulation of macroalgal mats on intertidal mudflats in two basins in Southern England. – Neth. J. Sea Res. 19, 257–263.
- Thomas, T. E. & Harrison, P. J., 1987. Rapid ammonium uptake and nitrogen interactions in five intertidal seaweeds grown under field conditions. – J. exp. mar. Biol. Ecol. 107, 1–8.
- Walters, G. J. & Wharfe, J. R., 1980. Distribution and abundance of *Hydrobia ulvae* (Pennant) in the lower Medway Estuary, Kent. – J. moll. Stud. 46, 171–180.
- Wohlenberg, E., 1937. Die Wattenmeer-Lebensgemeinschaften im Königshafen von Sylt. – Helgoländer wiss. Meeresunters. 1, 1–92.
- Woodin, S. A., 1977. Algal "gardening" behaviour by nereid polychaetes: Effects on soft-bottom community structure. – Mar. Biol. 44, 39–42.