Helgoländer wiss. Meeresunters. 29, 503-523 (1977)

# On the dynamics of exploited populations of *Tisbe bolothuriae* (Copepoda, Harpacticoida)

# V. The toxicity of cadmium: Reponse to sub-lethal exposure

# M. Hoppenheit

Biologische Anstalt Helgoland (Laboratorium Sülldorf); Hamburg 55, Federal Republic of Germany

ABSTRACT: A total of 30 weekly exploited populations of the harpacticoid copepod *Tisbe* holothuriae Humes were exposed to 148 or 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup>, combined with exploitation rates of 10, 30, 50, 70 or 90% under conditions of surplus food supply at 22 °C and 30% S. During the observation period, most populations exhibited distinct U-shaped density trends. Addition of Cd<sup>++</sup> prolonged and reinforced the downward trends present in the initial periods of the experiments. If population densities, both in control and test populations, fell short of a certain value, the age structures shifted in favour of the nauplii. More pronounced reductions in population densities of the test populations), the effect of Cd<sup>++</sup> on population density was counteracted by an acclimation process. No relationship could be established between exploitation rate and observed effects. Erroneous assumptions on the adaptability of parametric methods had led to invalid statements on the dispersion of data published in previous papers of this series. These statements have been revised. Possible limitations of the significance of the results obtained are discussed in some detail.

## INTRODUCTION

In recent years attention has repeatedly been called to problems concerning pollution of aquatic environments by cadmium (cf. a recently published paper on economic and social aspects of cadmium pollution by Nobbs & Pearce, 1976, and literature cited by Hoppenheit & Sperling, 1977). Up to now, the majority of papers dealing with effects of cadmium on aquatic organisms have been concerned with acute exposures to the pollutant. However, today one frequently encounters the statement that knowledge of the concentration of a pollutant which kills  $50 \,^{0}/_{0}$  of test specimens in a given time is insufficient to assess an acceptable non-toxic concentration (cf. Sprague, 1971; Olla, 1974; Nobbs & Pearce, 1976). Studies on acute toxicity may have serious limitations because, for example, they ignore differences in the sensibility of a test species throughout several consecutive generations (Winner & Farrell, 1976), the possibility of the occurrence of adaptations (Stockner & Antia, 1976) and the impairment of reproduction at concentrations in which there is no effect on survival (Mount, 1968).

## M. Hoppenheit

Since cadmium has been proven to be highly toxic to aquatic organisms (cf. Anderson, 1950; Kumada et al., 1973; Baudouin & Scoppa, 1974; Vernberg et al., 1974, and literature cited by Hoppenheit & Sperling, 1977) and because of the obviously great need for data on the effects of chronic exposure, it appeared reasonable to conduct experiments with *Tisbe holothuriae* which has previously been shown to be an appropriate organism to study several components of its population dynamics in the laboratory (cf. Hoppenheit, 1975a, b; 1976).

## MATERIAL AND METHODS

All experiments were made in essentially the same manner as those decribed in previous papers (Hoppenheit, 1975a, b; 1976; Hoppenheit & Sperling, 1977).

The populations of *Tisbe holothuriae* Humes were maintained in 200 ml of nonaerated sea-water in 500-ml, wide-mouth, flat-bottom flasks at  $22 \pm 1^{\circ}$  C in a room having 12 h light per day. Fifteen populations were each exposed to 148 and 222 µg Cd<sup>++</sup> l<sup>-1</sup>. Both groups were subdivided into groups of three replicates each, which were exploited weekly at rates of 10, 30, 50, 70 or 90 % according to the table presented by Hoppenheit & Sperling (1977, p. 330). The specimens were fed dried mussel flesh (mantle) once a week following exploitation. At the time of weekly exploitation, 90 % of the contaminated medium was renewed. In samples of 10 ml drawn before the harvests, the numbers of nauplii and adults plus copepodids were determined.

For further details the reader is referred to previous publications of this series.

### RESULTS

In a study on the effects of  $Cd^{++}$  on weekly exploited populations of *Tisbe* holothuriae (Hoppenheit & Sperling, 1977) it has been shown that concentrations of 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> or less do not result in extinction of an exploited population.

The characteristics of distributions of population-density data in three populations exploited weekly at rates of 30, 50 or 90% (cf. Fig. 1) before and after the addition of 148  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> are given in Table 1. Density data obtained at the initial phase of population development are not taken into consideration. The distributions were found to be approximately normal and free of trend. Comparisons of the means by use of the t-test yielded significant differences between the time series before and after addition of Cd<sup>++</sup> ( $\alpha < 0.01$ ). No further approximately normal distribution together with independence of density data has been found at the same time in both parts of an experiment in the remaining 12 populations exposed to 148  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup>.

All populations exposed to 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> displayed deviations from normality and/or randomness to such an extent that no attempt has been made to test the significance of differences between time series belonging together. The results of experiments carried out at an exploitation rate of 10 % are shown in Table 2.

Six populations exploited at a rate of 10 % served as controls in order to distin-

guish disturbing influences of unknown and uncontrollable factors which possibly would have been able to mask or to simulate an effect in the period following addition of  $Cd^{++}$ . Five populations exhibited significant trends in population densities. Figure 2 presents the smoothed course of population density of a control experiment over a period of 70 weeks. The curve has been smoothed by moving an average of 3 items. The use of a new sea-water charge in Week 1 is followed by an increase of population density until Week 4. Owing to this effect it is assumed that the new charge of seawater, recently placed at our disposal by ship, had better properties than the stored



Fig. 1: Tisbe holothuriae. Courses of population densities in three populations at weekly removals of 30 (a), 50 (b) or 90% (c) of specimens before and after addition of 148  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> (indicated by an arrow)

water used before. A steady deterioration of "quality" of the water in subsequent weeks and an acclimation process were supposed to interact and to produce the Ushaped trend, which is exhibited more or less distinctly by all 36 populations under consideration. A detailed account of acclimation processes will be given in consecutive papers (Hoppenheit, in preparation). For the present, therefore, only an indication of the existence of such processes will be given.

As test populations were established 10 weeks earlier than controls the reaction to water from the renewed supply can be demonstrated more distinctly in the former. The courses of population densities in two populations exposed to 148 and 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup>, respectively, have been plotted in Figure 3. Again, the curves have been smoothed by moving an average of 3 items. Obviously the use of sea-water newly placed at disposal (Arrow I) resulted in a considerable population growth. In the population exploited weekly at a rate of 90 % the increase of population density is retarded.



µg Cd++ 1-1 before sity; Ni population sopulation density;	)/o (2)	1.8003 0.3373 30	0% before (1) and	nent C (2)	1.8274 0.5416 30
<i>rriae</i> exposed to 148 s; N population den s; N median of F variation	90 <sup>0</sup>	2.2693 0.2323 27	oited at a rate of 10 Fable 1)	Experir (1)	2.4207 0.2843 26
ons of <i>Tisbe holothu</i> umber of population 50% of water per w tion; V coefficient of	ion rate /0 (2)	2.0396 0.3747 30	<i>sbe holothuriae</i> expli- tion of symbols see	nent B (2)	1.7187 0.5560 30
hree exploited populatic ed in Tables 1 to 9: k nu density at renewal of 5 uency; s standard deviat	Exploitat 50°	2.3046 0.2820 25	Table 2 ee populations of <i>Ti</i> : 2d <sup>44</sup> I <sup>-1</sup> . (For explana	Experii (1)	2.3773 0.2324 26
l of specimens in thr tion of symbols used ek; N2 population de ple; p relative frequei	/0 (2)	2.0185 0.3618 30	l of specimens in thr addition of 222 µg C	1t A (2)	1.8732 0.6091 30
ities prior to remova ion of Cd <sup>++</sup> . Explana 00% of water per we n size of a saml	30 <sup>0</sup>	2.3750 0.1646 27	ities prior to remova	Experimer (1)	2.3994 0.2932 26
Mean population dens (1) and after (2) addit density at renewal of 9	Characteristics	log N slog N n	Mean population dens	Characteristics	log N slog N n

148 / 1 1-11 t Tieles Table 1



For comparison of tendencies in control and test populations Figure 4 was prepared. Since the test of Dixon for the analysis of extreme values (cf. Sachs, 1968; p. 278 f.), which is rather insensitive to deviations from normality and homogeneity of variance, showed no significant differences between population densities either in



Fig. 4: Tisbe holothuriae. Courses of mean population densities (moving averages of 6 items) of pooled control ( $\bullet$ ) and test populations ( $\bigcirc$ ) exploited weekly at a rate of 10% before and after the addition of Cd<sup>++</sup> (indicated by an arrow)



Fig. 5: Tisbe holothuriae. Frequency distributions of densities of pooled control populations (c) and populations exposed to 148 (a) 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> (b). Exploitation rate: 10 %

the six control populations or in the six test populations exposed to 148 or 222  $\mu$ g Cd<sup>++</sup>l<sup>-1</sup> the data of either group were pooled and average-population densities were calculated for every point of time. The resulting curves of means were smoothed by moving an average of 6 items to obtain a distinct picture of the essential events. Figure 4 illustrates that the downward trend was prolonged and reinforced in the test populations after addition of sub-lethal amounts of Cd<sup>++</sup>. While in the controls



l) and after (2		(2)	0.76	61.8 9.7 30
8 µg Cd <sup>++</sup> l <sup>-1</sup> before (1	Q/ 0/2	(1)	0.59	50.6 13.1 27
<i>buriae</i> exposed to 14 abols see Table 1)	on rate	(2)	0.74	60.1 8.9 30
pulations of <i>Tisbe holot</i> . (For explanation of syn	Exploitation	(1)	0.53	46.7 10.4 25
three exploited popul addition of Cd <sup>++</sup> . (F		(2)	0.68	56.3 10.1 30
rcies of nauplii of		(1)	0.51	45.6 8.4 27
Mean relative freque		Characteristics	ď	$\frac{\sin^{-1} Vp}{\sin^{-1} Vp}$

Dynamics of exploited populations of Tisbe holothuriae

Mean relative frequencies of nauplii in control populations of <i>Tisbe bolothuriae</i> exploited at a rate of $10^{0/6}$ for the periods before and after the addition of Cd <sup>++</sup> to test populations. (For explanation of symbols see Table 1)
---

Characterístics	Υ	В	Experiment C	D	щ	ц	
		Period before	addition of Cd <sup>++</sup> to te	st populations			
d	0.40	0.45	0.41	0.48	0.46	0.49	
$\sin^{-1} V_{\rm p}$	38.5	41.9	39.5	43.5	42.7	44.3	
$s_{sin^{-1}} Vp$	13.6	15.0	14.3	11.5	13.9	12.9	
- u	25	25	25	25	25	25	
		Period after ad	ldition of Cd <sup>++</sup> to test	populations			
ď	0.67	0.68	0.71	0.69	0.69	0.68	
$\sin^{-1} Vp$	55.3	55.9	57.6	56.9	56.8	55.9	
Ssin <sup>-1</sup> Vp	8.6	7.7	8.7	8.3	9.7	7.9	
, u	30	30	30	30	30	30	

# M. Hoppenheit

	10 %
	٩.
	rate
	c)
	at
le 5	exploited
Tab	bolothuriae
	Tisbe
	of
	pulations

d after (2) addition of 148 and
1) ar
Mean relative frequencies of nauplii in populations of <i>Tisbe bolothwiae</i> exploited at a rate of $10^{0/6}$ before (1 222 $\mu$ g Cd <sup>++</sup> I <sup>-1</sup> , respectively. (For explanation of symbols see Table 1)

	1	
ent C (2)	0.73 61.0 12.6 30	0.69 56.3 13.4 30
Experime (1)	0.49 44.3 11.6 26	0.53 47.0 9.8 26
t B (2)	[++ ] <sup>-1</sup> 0.73 59.6 13.3 30	-1 0.66 55.3 11.2 30
Experimen (1)	Addition of 148 µg Cd 0.51 45.4 8.7 26	Addition of 222 µg Cd <sup>++</sup> 1 0.49 12.9 26
hent A (2)	0.64 53.8 <b>16.3</b>	0.66 55.4 13.1 30
Experim (1)	0.50 45.3 12.8 26	0.47 43.3 11.1 26
Characteristics	$\frac{p}{\sin^{-1} V_P}$	$\frac{p}{s_{8in^{-1}}Vp}$

ſ

an inversion of the trend appears between the 45th and 46th experimental weeks, the inversion is delayed in the test populations and occurs between the 53rd and 54th weeks. After the 60th week mean-population densities are equal: The effect of addition of 148 or 222  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> has evidently been balanced by an acclimation process within the 23 weeks or 20 generations. Generation time has been found to be 7.9 days at 22° C (Hoppenheit, 1976).

The distributions of pooled values of population densities of test and control populations for the period of addition of Cd<sup>++</sup> to test populations are shown in Figure 5. All distributions exhibit negative skewness. As some inhomogeneity by clustering as a consequence of pooling has to be expected, the X-test of van der Waerden & Nievergelt (1956) was applied to prove the differences between the central tendencies. Mager & Theil (1974) showed that the application of this test can be profitable in the case of inhomogeneity of data. The comparisons yielded  $\alpha < 0.001$  (two-tailed) at exposure to 148 µg Cd<sup>++</sup> l<sup>-1</sup> and  $\alpha < 0.002$  when 222 µg Cd<sup>++</sup> l<sup>-1</sup> were added.

Previous experiments have yielded a close inverse relationship between population density and relative frequency of nauplii (Hoppenheit, 1976). Therefore, it can be expected to find a downward trend in density linked to an increasing relative frequency of nauplii, and vice versa. From Figure 6a, which presents the course of variation of density both in nauplii and adults plus copepodids of the control population shown in Figure 2, it appears that the ratio of nauplii to adults plus copepodids is altered in favour of the nauplii at low population densities. The changes in age structure can be taken directly from the plot of relative frequencies of nauplii in Figure 6b. The population exhibits a distinct downward trend of density in both the nauplii and the adults plus copepodids in the first phase of the experiment before addition of Cd\*+ to the test populations. The downward trend in density is accompanied by a decrease in the relative frequency of nauplii until the 21st experimental week. At this time, density of nauplii reached its lowest value (cf. Fig. 6a). While the entire population displays further decrease of density until the 33rd week (cf. Fig. 2) the relative frequency of nauplii increases to a value of 0.75. The fraction of nauplii remains enlarged almost to the end of the observation period when an increase of population density results in lower values.

A similar course of variation of relative frequency in nauplii has also been observed in the test populations of Figure 1 exhibiting no significant trends in population densities in either part of the experiments. As in the control the relative frequencies of nauplii started to rise some weeks before the beginning of the test period. The inspection of the variation of the population structure seems to provide a more distinct indication of the existence of a trend in population density than the use of statistical tests.

To stabilize the variances all relative frequency data of nauplii of experiments presented in Figure 1 were transformed into  $\sin^{-1} l/p$  (Table 3). At weekly exploitation of 90 %/0 of the specimens the transformed data exhibited a significant trend, which could be successfully eliminated. The comparisons of mean relative frequencies of nauplii of the periods before and after addition of 148 µg Cd<sup>++</sup> l<sup>-1</sup> by the t-test (by the X-test in the case of non-normally distributed data at 30 %/0 weekly exploitation) showed significant differences in each case. Weighted analyses of variance (weighted

		ewal		
(2)	3 55.7 12.4 0.22 90	ekly ren	u	72 36 72
2 µg Cd <sup>++</sup> l <sup>-1</sup>		buriae at wee	deviation $\overline{\mathrm{N}_2}$ and $\overline{\mathrm{M}_2}$ )	27 42 47
(1)	3 44.7 10.3 0.23 78	Tiske holot.	Mean ( (in <sup>0</sup> / <sub>0</sub> of	26 38 43
		lations of ee Table 1	$\tilde{\mathbf{N}}_{\mathbf{z}}$	184 136 100
+ [-1 (2)	3 58.1 13.5 0.23 90	loited popul of symbols se	$\mathbf{N}_{\mathbf{z}}$	189 151 114
48 µg Cd+	4	ans in exp	a	108 54 108
(1)	3 45.0 11.0 2.2 78	Table 7 eans and mediat water. (For expl	$\frac{\text{deviation}}{N_1}$ and $\hat{N}_1$ )	31 26 34
(2)	6 56.4 8.5 0.15 180	rom the mea of culture w	Mean de (in %0 of <u>N</u>	29 27 32
trol		densities f 0 or 50 %	Ň1	274 284 244
(1) Cont	6 41.7 12.5 0.30 150	f population of 9	N	306 280 260
		viations of	А	9 6 3
Characteristics	$\frac{k}{s \sin^{-1} V p}$	Mean absolute dev	Exploitation rate	0.10 0.50 0.90

by sample size), which resulted in the same significance levels, have been performed additionally to check the results obtained by the t-test, which should not be applied if sample size varies widely.

In an earlier paper on the influence of rate of weekly water renewal on population density of exploited populations of *Tisbe holothuriae* (Hoppenheit, 1975a), a reduction both in density and relative frequency of nauplii at reduced water renewal rate was reported. There too the comparisons of mean values of transformed relative frequencies before and after reduction of water-renewal rate were performed by means of the t-test (cf. Hoppenheit, 1975a; Table 4). As in the cases mentioned above weighted analyses of variance yield the same levels of significance. The decrement of the relative frequency of nauplii found at a reduced rate of water renewal is in concordance with present observations on decrease of relative frequency of nauplii in the period up to the 21st week (cf. Fig. 6b), provided that an increasing deterioration of quality of water supply has taken place in this period. Only if – as in the present experiments – the population density falls short of a certain value, the age structure of a population changes in favour of the nauplii. This behaviour may protect a population against extinction and renders possible the acclimation to impaired environmental situations by providing enough material to the selective agent.

Tables 4 and 5 present the mean relative frequencies of nauplii of control and test experiments at 10 % weekly exploitation. About 40 % of the time series of transformed data exhibit trends and/or deviations from normality. Differences between mean values of populations belonging to the same group and the same experimental period are small and no outliers were found by the test of Dixon (cf. Sachs, 1968; p. 278 f.). Therefore, the sin<sup>-1</sup> Vp were pooled (Table 6). No significant deviations from normality could be established in the pooled data by the chi-square test. The comparisons of the means by the t-test yielded significant differences between the periods before and after the addition of Cd<sup>++</sup> both in control and test populations (a < 0.001). As not all conditions for the use of the t-test were satisfied (variances are not constant) weighted analyses of variance have been computed which provided the same levels of significance.

From Table 6 it appears that the more intense reduction of population density in populations exposed to  $Cd^{++}$  (cf. Fig. 4) does not result in a larger elevation of the sample fraction of nauplii. But judging by the coefficients of variation there seems to be a higher variability of data in test populations following addition of  $Cd^{++}$  (0.23 and 0.22, respectively, compared with 0.15).

## DISPERSION OF DATA

# Addendum to previously published papers

Doubts of the correctness of previously made statements on dispersion of data entailed a verification of conclusions drawn on dispersion of population data in foregoing publications (Hoppenheit, 1975a, b; 1976\*). From this checking it followed

<sup>\*</sup> Erratum: In this article the  $\gg$  in line 29 of page 119 should read  $\geq$  and not as printed.

Tisbe bolo-	u	72 36 72	e at weekly	п	72 36 72
pulations of )	viation $\overline{V_2}$ and $\overline{N_2}$ )	35 35 33	e holothuria	viation and Ñ2)	51 72 56
exploited po s see Table 1	Mean de (in %0 of <u>N</u>	32 33 46	tions of Tisb	Mean de (in %) of N	46 64 50
nedians in o of symbol	Ř.	119 101 45	ited popula	Ñ2	53 39 50
neans and n explanation	N2	130 106 54	ns in exploi	aN <sup>a</sup>	59 45 60
from the n vater. (For e	ц	108 54 108	9 s and media	u u	108 54 108
 copepodids of culture <del>v</del>	viation $\frac{1}{1}$ and $\tilde{N}_1$	37 31 47	Table 9 n the means	riation and Ñ <sub>1</sub> )	48 43 40
f adults plus : 90 or 50%	Mean de (in % of N	34 30 40	nauplii froi	$\frac{1}{(in ^{9/6} of N_1)}$	42 43 37
densities of renewal of	Ž1	134 128 77	densities of	N,	136 144 151
population e at weekly	$\overline{N_1}$	148 136 90	population	IN1	158 144 170
viations of thuria	<del>کر</del> .	७७७	riations of	ĸ	6 3 6
Mean absolute de	Exploitation rate	0.10 0.50 0.90	Mean absolute dev	Exploitation rate	0.10 0.50 0.90

Dynamics of exploited populations of Tisbe holothuriae

that erroneously-made assumptions on the adaptability of parametric methods have to some extent led to invalid statements. The first paper of this series (Hoppenheit, 1975a) reported some aspects of effects of change in water renewal rate. In that publication a parametric test has been used to prove differences between coefficients of variation of untransformed population density data (cf. Tables 1 to 3 of that paper). It has been concluded that the reduction of the set point of population density by a decrease in rate of water renewal is accompanied by a reduction of



Fig. 7: Tisbe holothuriae. Cumulative relative frequencies of the productions of eggs per egg sac in populations exploited at rates of 10 (●), 70 (○) or 90% (△)

the range of fluctuation of the controlled condition, so that the precision of adjustment expressed in percentage remains essentially unchanged. This conservative decision has to be revised. Tables 7 to 9 present an impression of the situation. The mean absolute deviation from the mean and the median expressed as percentage has been used as a measure of the precision of adjustment. As can be taken from the tables the precision of adjustment remains about equal in the case of  $10 \, ^0/_0$  weekly exploitation. At exploitation rates 50 or 90  $^0/_0$  it becomes poorer with reduction of population density following a reduction of water renewal from 90 to 50  $^0/_0$  weekly. A comparison of the mean deviation values contained in Tables 8 and 9 shows that with decreasing population density the precision of adjustment is more reduced in the population of nauplii than in the population of adults plus copepodids.

In the populations of Figure 1 of the present paper calculations of mean deviations have also yielded a distinct reduction of precision of adjustment of population density in the second halves of the experiments. Therefore, except in the case of small exploitation rates a higher variability of data at reduced population density caused by a deterioration of water properties has to be expected.

No evidence has been found of a decreasing variability in the intrinsic rate of natural increase with increasing rate of exploitation as stated previously (Hoppenheit, 1975b). From ln  $N_t = \ln N_0 + rt$  where  $N_0$  is the population density following exploitation,  $N_t$  the population density before the next exploitation and r the intrinsic rate of natural increase, it follows that r is normally distributed if  $N_t$  or  $N_0$  are log-normally distributed (cf. Koch, 1966). The proof of the r-values, the means of which are shown in Table 5 in Hoppenheit (1975 b), according to the proposal of



Fig. 8: Tisbe holothuriae. Cumulative relative frequencies of nauplii hatching per egg sac in populations exploited at rates of 10 (●), 70 (○) or 90 % (△)



Fig. 9: Tisbe holothuriae. Cumulative relative frequencies of specimens per egg sac reaching the adult stage in populations exploited at rates of 10 (●), 70 (○) or 90 % (△)

Mager & Mager (1974) by use of probability paper, has yielded approximate normal distributions and no outliers. Thus, a transformation of values (cf. Hoppenheit, 1975b, Table 5) is needless and misleading. As tests for trend have shown no significant deviations from randomness in the time series presented in Figures 3 to 7 in Hoppenheit (1975b), the parametric test provided by Hartley (cf. Sachs, 1968; p. 480 f.) to prove the homogeneity of several variances of samples of equal size has been applied which

resulted in no significant heterogeneity. As no inverse relationship exists between the exploitation rate and variability of intrinsic rate of natural increase, it cannot be stated (cf. Hoppenheit, 1975a) that the relation between birth and death rate (r equals birth rate minus death rate) and thereby the relation between the regulating variables mortality and fertility are subjected to relatively higher fluctuations at lower exploitation rates. As is to be expected from density data (cf. Hoppenheit, 1975b; Table 1) the variability in the r-values is slightly less at moderate exploitation rates and it has to be stated that birth and/or death rates are subjected to slightly higher fluctuations at low *and* high exploitation rates.

In Hoppenheit (1976, p. 114) it was stated that the variability of numbers of nauplii and adults produced per egg sac decreases with increasing exploitation rate and that these findings correspond with the formerly found reduction of variability in the rate of increase. As has been demonstrated above, the former assumption of a reduction of variability of the rate of increase with increasing exploitation has to be revised. In Figure 7 the cumulative relative frequencies of the productions of eggs per egg sac are presented for the experiments at an exploitation of 10, 70 or 90 % weekly (cf. Hoppenheit, 1976; Table 4). Curves at the exploitation rates of 10, 30 or 50 % were found to be almost identical. Due to practical difficulties the desirable multiple comparison of the dispersions could not be made. Single comparisons by Hagas' dispersion test (cf. Lienert, 1973; p. 388 f.) showed 7 significant higher dispersions of data at the higher exploitation rate among the 10 possible combinations ( $\alpha < 0.01$ , two-tailed). From inspection and the test results it may be concluded that increasing exploitation rate yields a higher dispersion in numbers of eggs produced per egg sac. In nauplii hatching per egg sac (Fig. 8) single comparisons have shown significant increase in dispersion in 90 % exploitation as compared to the other rates  $(\alpha < 0.01)$ . In numbers of specimens per egg sac reaching the adult stage (Fig. 9) significant differences in dispersion ( $\alpha < 0.01$ ) were only found between 90 % exploitation on the one hand and 10 and 50  $^{0}/_{0}$  on the other.

The increase of dispersion in the production of eggs per egg sac with increasing exploitation rate is evident in numbers of nauplii hatching per egg sac only at 90  $^{0}$ / $_{0}$  weekly exploitation. In the numbers of adults produced per egg sac differences in dispersion are minor and hardly detectable.

## DISCUSSION

The results reported in the preceding paper on responses to lethal exposure (Hoppenheit & Sperling, 1977) and those in the present paper indicate that *Tisbe* holothuriae tolerates relatively high concentrations of Cd<sup>++</sup> in comparison with other crustaceans such as *Daphnia magna* (Anderson, 1950; Biesinger & Christensen, 1972), *Crangon septemspinosa, Pagurus longicarpus* and *Palaemonetes vulgaris* (Eisler, 1971), *Tigriopus japonicus* (D'Agostino & Finney, 1974) and larvae of *Uca pugilator* (Vernberg et al., 1974).

With few exceptions, unfortunately, it has not to date been possible to breed sensitive and ecologically important marine species in the laboratory. In the present case too, an easy-to-breed organism has been studied, which may be less vulnerable to pollution. Moreover, it has to be taken into account that the stock populations were maintained for several years in the laboratory before experiments were started. In the culture water high concentrations of  $NO_2^-$  have been found (Hoppenheit, 1975a) indicating that the populations are able to tolerate a high degree of pollution from decomposing organic material. This water burden – besides other factors – may act as a selective force contributing to a change of the genetic structure of the populations tested. Hence, their responses may have been modified. Since it has been demonstrated by the present paper that an acclimation process may be terminated in a relatively short period of time, one can scarcely expect to deal with populations comparable to those occurring in their natural habitat.

The experiments were performed at a salinity of 30 ‰ and a temperature of  $22^{\circ}$  C. Generally, high temperatures seem to increase the amount of Cd<sup>++</sup> taken up by crustaceans whereas increasing salinity reduces the uptake (O'Hara, 1973a, b; Hutcheson, 1974; Vernberg et al., 1974, 1977, Jones, 1975). Dissolved oxygen concentration can be vigorously reduced at high population densities (Hoppenheit, 1975a). The mode of action of dissolved oxygen in experiments on the toxicity of Cd<sup>++</sup> is not unique. In fishes, Middaugh et al. (1975) and Voyer et al. (1975) report an increased sensitivity at lower concentrations of dissolved oxygen. Findings by Clubb et al. (1975) in aquatic insects have shown an increased amount of the metal taken up.

In the preceding paper (Hoppenheit & Sperling, 1977) experiments were reported in which 3 populations each were exposed to 500 and 1125  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> at a removal rate of 5 % of specimens *five* times a week to detect acute mortality. After a reduction of population density to 2 specimens per 10 ml one population exposed to 1125  $\mu$ g Cd<sup>++</sup> l<sup>-1</sup> recovered from acute intoxication and reestablished initial population density within 3 days. The possibility, therefore, has to be taken into consideration that the exploitation regime has an influence on the capacity of a population to compensate for deleterious stresses.

As the importance of modifying factors can scarcely be overestimated and may account for much of the variation in toxicity found in the literature (cf. Sprague, 1970), more efforts should be made to assess the possible interaction of environmental factors and poisonings.

In studies on the toxicity of transition metals – especially when performed under static conditions as in the present case – the question arises as to what extent the metal reacts chemically or is adsorbed by particulate matter and to which fraction of the initially added metal the test organisms are actually exposed. Thus the relatively low sensitivity found in *Tisbe holothuriae* could be the consequence of an adulteration by the experimental conditions and may be attributed to a reduced amount of Cd<sup>++</sup> actually available to the test organisms because the metal has partly been captured by compartments of the elected experimental system which are not in interaction with the copepods. Consequences to be drawn from the partitioning of the toxicant in the experimental system will be discussed in some detail in a subsequent paper (Sperling & Hoppenheit, in preparation).

### M. Hoppenheit

As it has been found that survival time of the populations is related to the initially adjusted concentration of  $Cd^{++}$  (Hoppenheit & Sperling, 1977), the scale given by the different designated concentrations has been used as a reference in describing the obtained results. Since the amount of  $Cd^{++}$  to which the copepods are actually exposed has to be expected to depend on the type of the experimental system elected, it cannot be excluded that deviating results are achieved when the manner of experimentation is changed.

Although the significance of the results obtained may be limited by uncertainties with reference to number and importance of factors determining the experimental conditions, investigations on population dynamics of *Tisbe holothuriae* should be continued because this species complies in an outstanding way with the requirements needed to gain more insight into the regulating mechanisms operating at the population level under the influence of toxicants.

Acknowledgements. The author thanks Dr. K.-R. Sperling for helpful suggestions and discussions, Mrs. H. Rade and Mrs. K. Bote for their assistance in the experiments, Mr. G. Nagel for support in solving statistical problems, Miss M. Blake for help in preparation of the manuscript as well as Miss R. Kleinfeld and Mr. J. Marschall for drafting the figures.

## LITERATURE CITED

- Anderson, B. G., 1950. The apparent thresholds of toxicity to *Daphnia magna* for chlorides of various metals when added to Lake Erie water. Trans. Am. Fish. Soc. **78**, 96–113.
- Baudouin, M. F. & Scoppa, P., 1974. Acute toxicity of various metals to freshwater zooplankton. Bull. environm. Contam. Toxicol. 12, 745–751.
- Biesinger, K. E. & Christensen, G. M., 1972. Effects of various metals on survival, growth, reproduction, and metabolism of *Daphnia magna*. J. Fish. Res. Bd Can. 29, 1691–1700.
- Clubb, R. W., Gaufin, A. R. & Lords, J. L., 1975. Synergism between dissolved oxygen and cadmium toxicity in five species of aquatic insects. Environm. Res. 9, 285-289.
- D'Agostino, A. & Finney, C., 1974. The effect of copper and cadmium on the development of *Tigriopus japonicus*. In: Pollution and physiology of marine organisms. Ed. by F. J. Vernberg & W. B. Vernberg. Acad. Press, New York, 445-463.
- Eisler, R., 1971. Cadmium poisoning in *Fundulus heteroclitus* (pisces: cyprinodontidae) and other marine organisms. J. Fish. Res. Bd Can. 28, 1225-1234.
- Hoppenheit, M., 1975a. Zur Dynamik exploitierter Populationen von Tisbe holothuriae (Copepoda, Harpacticoida). I. Methoden, Verlauf der Populationsentwicklung und Einfluß der Wassererneuerung. Helgoländer wiss. Meeresunters. 27, 235-253.
- 1975b. Zur Dynamik exploitierter Populationen von Tisbe holothuriae (Copepoda, Harpacticoida). II. Populationsdichte, Alterszusammensetzung, Wachstum und Ausbeute. Helgoländer wiss. Meeresunters. 27, 377-395.
- 1976. Zur Dynamik exploitierter Populationen von Tisbe holothuriae (Copepoda, Harpacticoida). III. Reproduktion, Geschlechtsverhältnis, Entwicklungsdauer und Überlebenszeit. Helgoländer wiss. Meeresunters. 28, 109–137.
- & Sperling, K.-R., 1977. On the dynamics of exploited populations of *Tisbe holothuriae* (copepoda, harpacticoida). IV. The toxicity of cadmium: Response to lethal exposure. Helgoländer wiss. Meeresunters. 29, 328-336.
- Hutcheson, M. S., 1974. The effect of temperature and salinity on cadmium uptake by the Blue crab, *Callinectes sapidus*. Chesapeake Sci. 15, 237-241.
- Jones, M. B., 1975. Synergistic effects of salinity, temperature and heavy metals on mortality and osmoregulation in marine and estuarine isopods (crustacea). Mar. Biol. 30, 13-20.

- Koch, A. L., 1966. The logarithm in biology. 1. Mechanisms generating the log-normal distribution exactly. J. theor. Biol. 12, 276–290.
- Kumada, H., Kimuru, S., Yokote, M. & Matida, Y., 1973. Acute and chronic toxicity, uptake and retention of cadmium in freshwater organisms. Bull. Freshwat. Fish. Res. Lab., Tokyo, 22, 157-165.
- Lienert, G. A., 1973. Verteilungsfreie Methoden in der Biostatistik. Hain, Meisenheim, 1, 1–736.
- Mager, P. P. & Mager, R., 1974. Das Wahrscheinlichkeitsnetz. I. Prüfung auf Normalverteilung und Analyse eines Mischkollektives. Gegenbaurs morph. Jb. 120, 560–574.
- & Theil, S., 1974. Zur statistischen Auswertung von Reaktionszeiten, die während der Ausarbeitung einer bedingten Fluchtreaktion (avoidance conditioning) erhalten wurden. Biometr. Z. 16, 49-54.
- Middaugh, D. P., Davis, W. R. & Yoakum, R. L., 1975. The response of larval fish, *Leiosto-mos xanthurus*, to environmental stress following sublethal cadmium exposure. Contr. mar. Sci. 19, 13-19.
- Mount, D. I., 1968. Chronic toxicity of copper to fathead minnows (Pimephales promelas, Rafinesque). Wat. Res. 2, 215-223.
- Nobbs, C. L. & Pearce, D. W., 1976. The economics of stock pollutants: The example of cadmium. Int. J. environm. Stud. 8, 245-255.
- O'Hara, J., 1973a. Cadmium uptake by fiddler crabs exposed to temperature and salinity stress. J. Fish. Res. Bd Can. 30, 846-848.
- 1973b. The influence of temperature and salinity on the toxicity of cadmium to the fiddler crab, Uca pugilator. Fish. Bull. U.S. 71, 149–153.
- Olla, B. L. (Ed.), 1974. Behavioral measures of environmental stress. In: Marine Bioassays. Workshop Proceedings. Marine Technology Society, Washington, D. C., 1-31.
- Sachs, L., 1968. Statistische Auswertungsmethoden. Springer, Berlin, 671 pp.
- Sprague, J. B., 1970. Measurement of pollutant toxicity to fish. II. Utilizing and applying bioassay results. Wat. Res. 4, 3-32.
- 1971. Measurement of pollutant toxicity to fish. III. Sublethal effects and "safe" concentrations. Wat. Res. 5, 245-266.
- Stockner, J. G. & Antia, N. J., 1976. Phytoplankton adaptation to environmental stresses from toxicants, nutrients, and pollulants – a warning. J. Fish. Res. Bd Can. 33, 2089–2096.
- Vernberg, W. B., Coursey, P. J. de & O'Hara, J., 1974. Multiple environmental factor effects on physiology and behavior of the fiddler crab, *Uca pugilator*. In: Pollution and physiology of marine organisms. Ed. by F. J. Vernberg & W. B. Vernberg. Acad. Press, New York, 381-425.
- Kelly, M. & Johns, D. M., 1977. Effects of sublethal concentrations of cadmium on adult *Palaemonetes pugio* under static and flow-through conditions. Bull. environm. Contam. Toxicol. 17, 16-24.
- Voyer, R. A., Yevich, P. P. & Barszcz, C. A., 1975. Histological and toxicological responses of the mummichog, *Fundulus heteroclitus* (L.) to combinations of levels of cadmium and dissolved oxygen in a freshwater. Wat. Res. 9, 1069-1074.
- Waerden, B. L. van der & Nievergelt, E., 1956. Tafeln zum Vergleich zweier Stichproben mittels X-Test und Zeichentest. Springer, Berlin, 34 pp.
- Winner, R. W. & Farrell, M. P., 1976. Acute and chronic toxicity of copper to four species of Daphnia. J. Fish. Res. Bd Can. 33, 1685-1691.

Author's address: Dr. M. Hoppenheit Biologische Anstalt Helgoland (Laboratorium Sülldorf) Wüstland 2 D-2000 Hamburg 55 Federal Republic of Germany