

Spatio-temporal changes in biomass carbon sinks in China's forests from 1977 to 2008

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Forests play a leading role in regional and global carbon (C) cycles. Detailed assessment of the temporal and spatial changes in C sinks/sources of China's forests is critical to the estimation of the national C budget and can help to constitute sustainable forest management policies for climate change. In this study, we explored the spatio-temporal changes in forest biomass C stocks in China between 1977 and 2008, using six periods of the national forest inventory data. According to the definition of the forest inventory, China's forest was categorized into three groups: forest stand, economic forest, and bamboo forest. We estimated forest biomass C stocks for each inventory period by using continuous biomass expansion factor (BEF) method for forest stands, and the mean biomass density method for economic and bamboo forests. As a result, China's forests have accumulated biomass C (i.e., biomass C sink) of 1896 Tg (1 Tg=10¹² g) during the study period, with 1710, 108 and 78 Tg C in forest stands, and economic and bamboo forests, respectively. Annual forest biomass C sink was 70.2 Tg C a⁻¹, offsetting 7.8% of the contemporary fossil CO₂ emissions in the country. The results also showed that planted forests have functioned as a persistent C sink, sequestering 818 Tg C and accounting for 47.8% of total C sink in forest stands, and that the old-, mid- and young-aged forests have sequestered 930, 391 and 388 Tg C from 1977 to 2008. Our results suggest that China's forests have a big potential as biomass C sink in the future because of its large area of planted forests with young-aged growth and low C density.

bamboo forests, biomass carbon stock, carbon sink, forest inventory, economic forests, natural forests, planted forests

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Forests play a dominant role in regional and global carbon (C) cycles because they contain 85%–90% of total terrestrial vegetation biomass and annually exchange up to 90% of total terrestrial ecosystem C with the atmosphere through photosynthesis and respiration [1–12]. Forests become sources of atmospheric CO₂ when disturbed by human and natural disturbances, but they can also function as C sinks to

sequester or conserve significant quantities of C during re-growth after disturbance [11,13–16]. Because of their importance in the C cycles, the Kyoto Protocol clearly suggested increasing C sequestration through afforestation and reforestation to meet greenhouse gas emission targets during commitment periods for signatory countries of the United Nations Framework Convention on Climate Change (UNFCCC) [14].

Estimating broad-scale biomass C stocks and their dy-

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namics has been one of the important issues in the study of the global C cycle. In recent decades, forest inventory data have been widely used to estimate C stocks and fluxes for the world's forests at a regional or national scale, because these data are generally collected across the whole area of the forests and designed to be statistically valid [3,5,7,15,17,18]. For example, based on U.S. Forest Inventory and Analysis (FIA) data, Brown and Schroeder [15] estimated that eastern U.S. forests annually accumulated $\sim 174 \text{ Tg C}$ ($1 \text{ Tg} = 10^{12} \text{ g}$) during the late 1980s and early 1990s and offset about 13% of the C emissions from fossil fuels and cement manufacturing in the United States during the early 1990s [19]. Similarly, Europe's terrestrial ecosystems have absorbed 7%–12% of European anthropogenic CO_2 emissions [7].

Located in the eastern margin of Eurasia, China ranks the fifth in its forest area in the world [20] and encompasses various forest biomes, from boreal forests in the north to the subtropical/tropical evergreen broadleaf forests in the south [21], which provides a unique area to study the regional forest C cycle. In addition, with the implementation of national afforestation and reforestation programs since the late 1970s, such as the Three-north Protective Forest Program, the Natural Forest Conservation Program, and the Wetland Restoration Program, forest ecosystems in China are thought to have significantly contributed to the regional and global C sinks in the past several decades [3,11,18,22–32]. Based on national forest inventory data, Fang et al. [3] estimated changes in the national C stocks of living forest biomass between 1949 and 1998, concluding that China's forests had functioned as a C sink at a mean annual accumulation rate of 0.021 Pg C ($1 \text{ Pg} = 10^{15} \text{ g}$) during the study period, mainly due to forest expansion and regrowth. One decade later, Pan et al. [11] updated these inventory data and found that the magnitude of China's forest biomass C sink almost doubled from $0.060 \text{ Pg C a}^{-1}$ in the 1990s (1990–1999) to $0.115 \text{ Pg C a}^{-1}$ in the 2000s (2000–2007). While these studies have contributed to understanding of the role of China's forests in the regional and global C budget, they lack systematic and detailed analysis on changes in forest biomass C stocks with different zonal forest type, age class, and different province. Therefore, based on the up-to-date national forest inventory data since the late 1970s, this study is to explore: (i) the temporal changes in biomass C stocks for China's forests and among forest stands, economic and bamboo forests, and (ii) the changes in biomass C stocks in the forest stands, and planted and natural forests, with forest type and age class.

1 Data and methods

China's forest inventory data for the period of 1977–1981, 1984–1988, 1989–1993, 1994–1998, 1999–2003, and

2004–2008 were used in this study. According to the definition of China's forest resource inventory, China's forest was categorized into three groups: forest stand (including planted and natural forests), economic forest (woods with the production of fruits, edible oils, drinks, flavorings, industrial raw materials, and medicinal materials as the main purposes), and bamboo forest. However, the forest inventory provides different information for these different forest groups: it documents the areas and timber volumes by dominant tree species and by age class in each province for forest stands, as well as the respective areas and volumes of natural forests and plantations for each province, and only the forest area in each province for economic and bamboo forests. Note that due to the lack of data, forest in Hong Kong, Macao and Taiwan was not included in this study.

Three methods, including the mean biomass density method, the mean biomass expansion factor (BEF, ratio of forest biomass to timber volume) method, and the continuous BEF method (the function of BEF as timber volume), are commonly used to estimate forest biomass C stocks across the world [33,34]. Using China's forest inventory data between 1984 and 2003, Guo et al. [34] compared these three methods and concluded that the continuous BEF method was most reasonable and accurate to estimate regional forest biomass when there are sufficient forest inventory data, as well as field measurement data. Therefore, we estimated forest biomass C stocks for each inventory period by using the continuous BEF method for forest stands, and the mean biomass density method for economic and bamboo forests. It should be noted that a ratio of 0.5 was used in this study to convert biomass to C stock.

1.1 Forest stands

Because forest inventory data only document information on area and timber volume of forest stands, it is necessary to develop allometric relationships between forest biomass and forest timber volume to obtain BEF for each forest type [17,23,33]. Previous studies have suggested that the value of BEF varies with forest age, site class, stand density and quality [3,17]. Fang and his colleagues [3,14,15,17,23,35,36] have derived a simple reciprocal equation from direct field measurements to express the BEF-timber volume relationship for a specific forest type (eq. (1), Table S1 in Supporting Information), and one can easily scale up to calculate regional forest biomass based on direct field measurements and forest inventory data by using these equations [37]. For details about this method, see Fang et al. [3,18] and Guo et al. [34].

$$\text{BEF} = a + b/x, \quad (1)$$

where x is the timber volume per unit area, a and b are constants for a specific forest type, and BEF is the biomass expansion factor.

It should be noted that since 1994, the definition of forest

stands in China's forest resource inventory has been changed from >30% canopy coverage to $\geq 20\%$ canopy coverage. In order to make our results comparable among different periods, Fang et al. [18] analyzed the 1994–1998 inventory data which provided both criteria (20% and 30% canopy coverage), and found the robust linear relationships between the two criteria for the forest area and biomass C stocks at the provincial level. In this study, we modified their equations with power function relationships (eqs. (2) and (3)) to obtain better estimates.

$$\text{AREA}_{0.2} = 1.290 \times \text{AREA}_{0.3}^{0.995} \quad (R^2 = 0.996, n = 30), \quad (2)$$

$$\text{CARBON}_{0.2} = 1.147 \times \text{CARBON}_{0.3}^{0.996} \quad (R^2 = 0.999, n = 30), \quad (3)$$

where AREA and CARBON are the forest stand area (10^4 ha) and biomass C stock (Tg C) in a province, respectively; subscripts 0.3 and 0.2 represent the criterion of >30% and $\geq 20\%$ canopy coverage, respectively.

In addition, because early forest inventory data only documented total area and volume of natural and planted forests for each province without detailed information on dominant tree species for each type, we could not use the continuous BEF method as the whole forest stands to calculate biomass C stocks in each inventory period for natural and planted forests. Therefore, following Fang and Chen [38], we established a robust linear relationship between the mean biomass density and mean volume density at the provincial level (eq. (4), Figure 1). Using this equation, we first calculated separately the biomass C stocks of natural and planted forests for each province and each inventory period, and then obtained their corresponding proportions in each province and each inventory period. The final biomass C stocks of natural and planted forests were derived by multiplying these corresponding proportions to total biomass C stock of forest stands, which was obtained using the continuous BEF method.

$$BD = 0.704 \times VD + 19.953 \quad (R^2 = 0.968, n = 211), \quad (4)$$

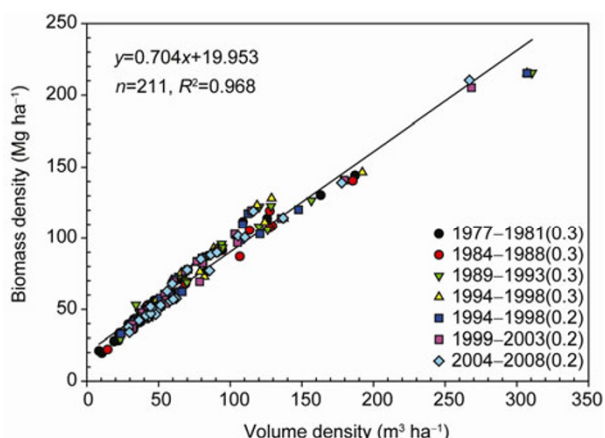


Figure 1 Relationship between the mean biomass density (Mg ha^{-1}) and the mean volume density ($\text{m}^3 \text{ha}^{-1}$) at provincial level (0.3 and 0.2 in parenthesis show the criterion of >30% canopy coverage and $\geq 20\%$ canopy coverage, respectively).

where *BD* and *VD* are the mean biomass density (Mg ha^{-1} , $1 \text{ Mg} = 10^6 \text{ g}$) and mean volume density ($\text{m}^3 \text{ha}^{-1}$) of forest stands in each province in different period, respectively.

1.2 Economic and bamboo forests

Because forest inventory only documents the area of economic forests in each province, we estimated biomass C stocks in China's economic forests between 1977 and 2008 by multiplying the mean biomass density (23.7 Mg ha^{-1} , Fang et al. [22]) with the area of economic forests in each province in each inventory period.

Forest inventory documents the areas of moso bamboo and other bamboo forests in each province. We established a literature-reviewed database of bamboo biomass, which contained 37 sets of biomass data for moso bamboo forests and 43 sets for other bamboo forests. The mean biomass density of moso bamboo and other bamboo forests was estimated as 81.9 and 53.1 Mg ha^{-1} respectively, based on the database. We therefore can obtain biomass C stocks in China's bamboo forests in the different inventory periods, using the mean biomass density for each bamboo type and areas in each province.

2 Results

2.1 Biomass C stocks and C sinks of China's forests

Total forest biomass C stock increased by 38.6% from 4972 Tg C (i.e., 4.972 Pg C) in the early 1980s (1977–1981) to 6868 Tg C in the late 2000s (2004–2008), with a net accumulation of 1914 Tg C and an overall biomass C sink rate of 70.2 Tg C a^{-1} (Table 1). As shown in Table 1, biomass C sinks varied greatly in different periods: the C sink in 1994–1998 was 10.1 Tg C a^{-1} , and then dramatically increased to $114.9 \text{ Tg C a}^{-1}$ in 2004–2008, implying a significant increase in C sink in China's forests.

Occupying 84.4%–89.4% of total forest area, forest stands have stored 93.2%–94.9% of total biomass C stock and net accumulated C of 1710 Tg , at an average rate of 63.3 Tg C a^{-1} , which accounted for 90.2% of total biomass C sink between 1977 and 2008. Net C gain was found between two sequential inventory periods, except a slight decrease of 8 Tg C in 1994–1998 probably due to the statistical error on the forest area, and the maximum biomass C sink was found in 2004–2008 with the value of $112.9 \text{ Tg C a}^{-1}$. The area-weighted mean biomass C density also increased from initial $38.2 \text{ Mg C ha}^{-1}$ to $41.3 \text{ Mg C ha}^{-1}$ in the late 2000s (2004–2008).

Over the past three decades, economic and bamboo forests have occupied 8.2%–12.9% and 2.3%–3.0% of total forest area, and stored 2.7%–4.1% and 2.4%–2.9% of total biomass C stock, respectively. During the period of 1977–2008, economic and bamboo forests have separately

Table 1 The magnitude and change of C stocks and C sinks in China's forests during 1977–2008

Period	All forests			Forest stands				Economic forests			Bamboo forests		
	Area (10 ⁴ ha)	C stock (Tg C)	C sink (Tg C a ⁻¹)	Area (10 ⁴ ha)	C stock (Tg C)	C density (Mg C ha ⁻¹)	C sink (Tg C a ⁻¹)	Area (10 ⁴ ha)	C stock (Tg C)	C sink (Tg C a ⁻¹)	Area (10 ⁴ ha)	C stock (Tg C)	C sink (Tg C a ⁻¹)
1977–1981	13798	4972	–	12350	4717	38.2	–	1128	134	–	320	121	–
1984–1988	14898	5178	29.5	13169	4885	37.1	23.9	1374	163	4.2	355	131	1.4
1989–1993	15960	5731	110.6	13971	5402	38.7	103.5	1610	191	5.6	379	138	1.5
1994–1998	15684	5781	10.1	13241	5388	40.7	–2.9	2022	240	9.8	421	154	3.1
1999–2003	16902	6293	102.3	14279	5862	41.1	94.9	2139	253	2.8	484	177	4.7
2004–2008	18138	6868	114.9	15559	6427	41.3	112.9	2041	242	–2.3	538	199	4.3
1977–2008			70.2				63.3			4.0			2.9

accumulated 108 and 78 Tg C, which equaled 5.7% and 4.1% of total biomass C sink in China's forests. Annual biomass C sink rate averaged 4.0 Tg C a⁻¹ for economic forests, ranging from –2.3 Tg C a⁻¹ in 2004–2008 to 9.8 Tg C a⁻¹ in 1994–1998, and 2.9 Tg C a⁻¹ for bamboo forests, ranging from 1.4 Tg C a⁻¹ in 1977–1981 to 4.7 Tg C a⁻¹ in 1999–2003, respectively.

2.2 Spatio-temporal distribution of biomass C sink of forest stands

Figure 2 presents provincial biomass C stock, density, and C sink in China's forest stands between the early 1980s (1977–1981) and the late 2000s (2004–2008), together with the forest area. In 1977–1981, the largest biomass C stock (801.8 Tg C) was in Heilongjiang, accounting for 17.0% of total biomass C stock in China's forest stands, followed by Tibet (621.9 Tg C, 13.2%), Yunnan (556.6 Tg C, 11.8%), Inner Mongolia (510.7 Tg C, 10.8%), and Sichuan (469.6 Tg C, 10.0%). The area-weighted biomass C density ranged from 8.8 (Shanghai) to 78.8 Mg C ha⁻¹ (Tibet), with an overall mean of 38.2 Mg C ha⁻¹ across the country's forest stands. Thirty years later, the largest biomass C stock (884.7 Tg C) was switched to Tibet, which accounted for 13.8% of total forest stand biomass C stock. Next to Tibet, it was Heilongjiang (815.5 Tg C, 12.7%), Yunnan (747.8 Tg C, 11.6%), Sichuan (719.4 Tg C, 11.2%), and Inner Mongolia (652.8 Tg C, 10.2%). In 2004–2008, the area-weighted biomass C density ranged from 16.9 Mg C ha⁻¹ (Shanghai) to 105.2 Mg C ha⁻¹ (Tibet), with an overall mean of 41.3 Mg C ha⁻¹.

As shown in Figure 2, forest stands in all provinces have functioned as C sinks during 1977–2008, and the largest C sink was in Tibet (280.8 Tg C), accounting for 15.4% of total biomass C sink in China's forest stands, followed by Sichuan (249.7 Tg C, 14.6%), Yunnan (191.2 Tg C, 11.2%), and Inner Mongolia (142.1 Tg C, 8.3%). In addition, the area of forest stands had increased in all provinces, except slight decreases in Jilin by 26.6×10⁴ ha and in Gansu by 4.7×10⁴ ha.

2.3 Biomass C stocks in planted and natural forests

The biomass C stocks and C sinks in planted and natural forests are given in Table 2. The area of planted forests had increased by 24.05×10⁶ ha since the early 1980s (1977–1981), which accounted for 74.9% of total area increment in the area of the forest stands, or 55.4% of China's forests, mainly due to afforestation and reforestation practices in the country in recent decades. Biomass C stock of planted forests had increased >3 times from 250 Tg C in the early 1980s (1977–1981) to 1063 Tg C in the late 2000s (2004–2008) and the proportion of biomass C stock in planted forests had consistently increased from initial 5.3% to 16.8% in 2004–2008. Consequently, planted forests had continuously sequestered C of 818 Tg from 1977 through 2008, at an average rate of 30.3 Tg C a⁻¹, which accounted for 47.8% of total biomass C sink in China's forest stands or 43.1% in China's forests. Compared with the planted forests, natural forests did not show a persistent C gain: it had released 132 Tg C and sequestered 1024 Tg C during the study period, with a net accumulation of 892 Tg C and at an average rate of 33.0 Tg C a⁻¹. It should be noted that since the late 1990s (1994–1998), natural forests have functioned as a persistent C sink, probably due to the implement of nation-wide Natural Forest Conservation Program starting in 1998.

Further, as shown in Table 2, the area-weighted mean biomass density of planted forests increased dramatically, from 15.6 Mg C ha⁻¹ in the early 1980s (1977–1981) to 26.7 Mg C ha⁻¹ in the late 2000s (2004–2008), which equaled to 37.7%–57.5% of that of natural forests, indicating that planted forests could still sequester more C by their regrowth in the future.

There was considerable spatial (or regional) difference in biomass C sinks in planted and natural forests over the past 30 years (Figure 3). Planted forests had acted as C sinks in all provinces during the period of 1977–2008 and the largest sink (81.1 Tg C) was in Sichuan, followed by Fujian (68.9 Tg C), Heilongjiang (60.9 Tg C), and Hunan (58.6 Tg C). For the natural forests, 25 of 30 provinces in China were C sinks and Tibet (262.1 Tg C), Sichuan (168.6 Tg C),

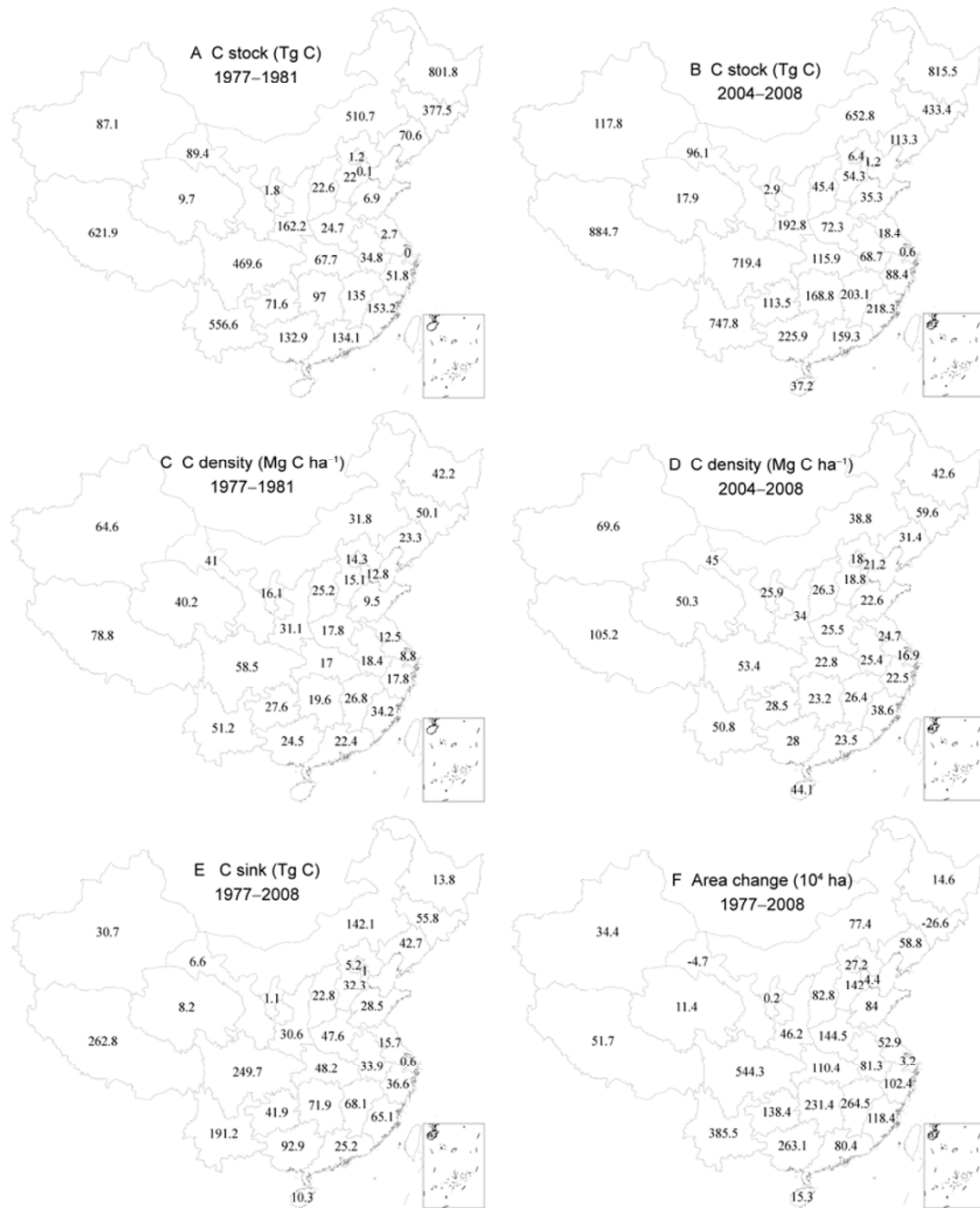


Figure 2 Spatio-temporal distributions in biomass C stocks, C densities, C sinks, and area change in forest stands during 1977–2008. A, C stock in 1977–1981. B, C stock in 2004–2008. C, C density in 1977–1981. D, C density in 2004–2008. E, C sink during 1977–2008. F, Area change during 1977–2008. There are no data in Hainan in 1977–1981, so the results of C sink and area change in Hainan are the values during 1984–2008.

Table 2 Area, C stock, C density, and C sink for planted and natural forests in China during 1977–2008

Period	Planted forests				Natural forests			
	Area (10 ⁴ ha)	C stock (Tg C)	C density (Mg C ha ⁻¹)	C sink (Tg C a ⁻¹)	Area (10 ⁴ ha)	C stock (Tg C)	C density (Mg C ha ⁻¹)	C sink (Tg C a ⁻¹)
1977–1981	1595	250	15.6	–	10755	4468	41.5	–
1984–1988	2347	418	17.8	24.1	10822	4467	41.3	–0.1
1989–1993	2675	526	19.7	21.6	11296	4876	43.2	81.9
1994–1998	2914	642	22.0	23.3	10326	4746	46.0	–26.2
1999–2003	3229	836	25.9	38.7	11049	5026	45.5	56.2
2004–2008	4000	1067	26.7	46.2	11559	5360	46.4	66.7
1977–2008				30.3				33.0

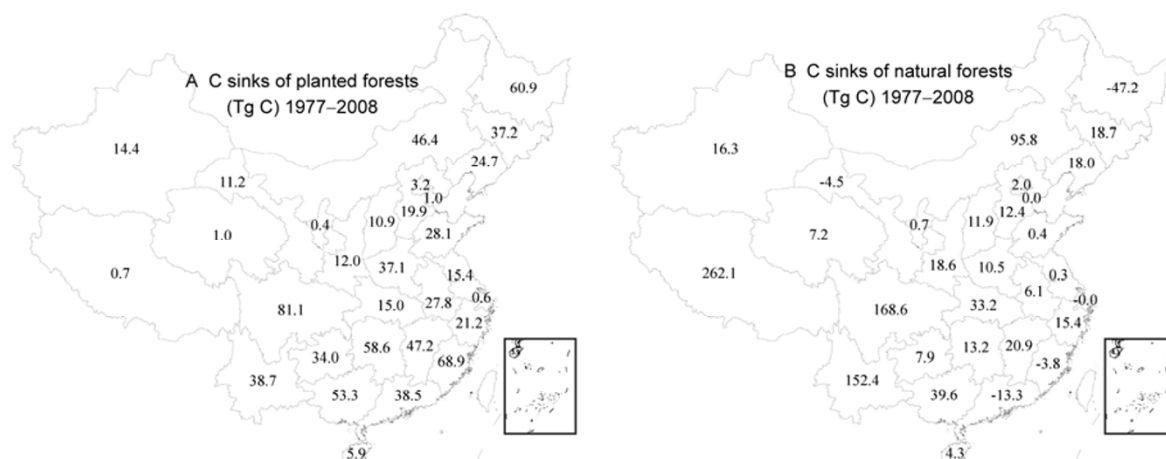


Figure 3 The regional distribution of biomass C sinks (Tg C a^{-1}) of planted and natural forest stands in each province during 1977–2008. A, C sinks of planted forests. B, C sinks of natural forests.

Yunnan (152.4 Tg C), and Inner Mongolia (95.8 Tg C) had the highest C sinks, which together accounted for 76.1% of total biomass C sink in natural forests across the country. The rest five provinces had released C and the largest biomass C source was in Heilongjiang (-47.2 Tg C), followed by Guangdong (-13.3 Tg C), Gansu (-4.5 Tg C), Fujian (-3.8 Tg C), and Shanghai (-0.02 Tg C).

2.4 Biomass C stocks of forest stands by forest type

China has almost all major forest types of Northern Hemisphere, with tropical rainforests in the south and boreal forests in the north [39]. To document the biomass C sink in various zonal forest types, we grouped China's forest stands into five zonal forest types based on dominant tree species according to Fang [40]: boreal forests, temperate coniferous forests, temperate deciduous forests, warm-temperate/subtropical mixed forests, and evergreen broadleaf forests (Table 3).

Table 3 depicts temporal change in biomass C stocks and C sinks during the period of 1977–2008 by zonal forest type. In the early 1980s (1977–1981), the largest biomass C stock (1449 Tg C) was in boreal forests, which accounted for 31.0% of total country biomass C stock of forest stands. The biomass C stocks of evergreen broadleaf forests (1282 Tg C), temperate deciduous forests (870 Tg C), temperate/subtropical mixed forests (816 Tg C), and temperate coniferous forests (282 Tg C) accounted for 27.2%, 18.5%, 17.3%, and 6.0% of China's forest stand biomass C stock, respectively. However, this order has been changed 30 years: evergreen broadleaf forests stored the largest biomass C of 1901 Tg C (29.6%), followed by temperate deciduous forests (1577 Tg C , 24.5%), boreal forests (1242 Tg C , 19.3%), temperate/subtropical mixed forests (1156 Tg C , 18.0%), and temperate coniferous forests (552 Tg C , 8.6%). Consequently, four of five zonal forest types had functioned as C sinks during the period of 1977–2008 and the net C accu-

mulation was 706 Tg C (26.2 Tg C a^{-1}) in temperate deciduous forests, 617 Tg C (22.8 Tg C a^{-1}) in evergreen broadleaf forests, 340 Tg C (12.6 Tg C a^{-1}) in temperate/subtropical mixed forests, and 268 Tg C (9.9 Tg C a^{-1}) in temperate coniferous forests. Boreal forests have functioned as a stable C sink by sequestering 220 Tg C between 1977 and 1993, while switched to a C source by releasing 441 Tg C from 1994 through 2008, with a net release of 221 Tg C over the past 30 years. These differences in biomass C sinks in these five zonal forest types were largely attributed to the change in forest area. Over the past three decades, forest area has increased by $14.11 \times 10^6 \text{ ha}$ in temperate deciduous forests, $13.09 \times 10^6 \text{ ha}$ in evergreen broadleaf forests, $7.41 \times 10^6 \text{ ha}$ in temperate coniferous forests, and $1.8 \times 10^6 \text{ ha}$ in temperate/subtropical mixed forests, but decreased by $3.69 \times 10^6 \text{ ha}$ in boreal forests.

As shown in Table 3, during the period of 1977–2008, the area-weighted mean biomass C density has increased by 35.7% (7.9 Mg C ha^{-1}) in temperate/subtropical mixed forests, 15.0% (5.4 Mg C ha^{-1}) in temperate deciduous forests, 5.8% (2.3 Mg C ha^{-1}) in evergreen broadleaf forests, 3.5% (1.2 Mg C ha^{-1}) in temperate coniferous forests, and 2.2% (1.5 Mg C ha^{-1}) in boreal forests. By the late 2000s (2004–2008), boreal forests contained the highest biomass C density of $68.8 \text{ Mg C ha}^{-1}$, due to the higher biomass C densities of dominant tree species ($125.1 \text{ Mg C ha}^{-1}$ for *Abies*, $88.6 \text{ Mg C ha}^{-1}$ for *Picea*, and $44.3 \text{ Mg C ha}^{-1}$ for *Larix* in 2004–2008). Next to this, it was $42.2 \text{ Mg C ha}^{-1}$ in evergreen broadleaf forests, $41.3 \text{ Mg C ha}^{-1}$ in temperate deciduous forests, $35.1 \text{ Mg C ha}^{-1}$ in temperate coniferous forests, and $30.0 \text{ Mg C ha}^{-1}$ in temperate/subtropical mixed forests.

2.5 Biomass C stocks of forest stands by forest age class

According to forest inventory data, China's forest stands

Table 3 The magnitude and change of biomass C stocks in different zonal forest types in China during 1977–2008

Zonal forest group ^{a)}	Period	Area (10 ⁴ ha)	C stock (Tg C)	C density (Mg C ha ⁻¹)	C sink (Tg C a ⁻¹)
Boreal forests ^{b)}	1977–1981	2174	1463	67.3	–
	1984–1988	2196	1481	67.4	2.6
	1989–1993	2350	1683	71.6	40.4
	1994–1998	2131	1637	76.8	–9.3
	1999–2003	1826	1263	69.2	–74.7
	2004–2008	1805	1242	68.8	–4.3
	1977–2008				–8.2
Temperate coniferous forests ^{c)}	1977–1981	837	284	33.9	–
	1984–1988	1055	319	30.3	5.0
	1989–1993	1114	386	34.6	13.3
	1994–1998	1025	335	32.7	–10.2
	1999–2003	1144	410	35.8	15.0
	2004–2008	1573	552	35.1	28.4
	1977–2008				9.9
Temperate deciduous forests ^{d)}	1977–1981	2425	870	35.9	–
	1984–1988	3793	1400	36.9	75.7
	1989–1993	4129	1539	37.3	27.8
	1994–1998	3694	1566	42.4	5.4
	1999–2003	3824	1626	42.5	11.8
	2004–2008	3820	1577	41.3	–9.7
	1977–2008				26.2
Temperate/subtropical mixed forests ^{e)}	1977–1981	3697	816	22.1	–
	1984–1988	3459	703	20.3	–16.2
	1989–1993	4158	901	21.7	39.7
	1994–1998	4239	1016	24.0	23.1
	1999–2003	4528	1256	27.7	47.9
	2004–2008	3855	1156	30.0	–20.0
	1977–2008				12.6
Evergreen broadleaf forests ^{f)}	1977–1981	3217	1284	39.9	–
	1984–1988	2666	982	36.8	–43.2
	1989–1993	2221	893	40.2	–17.8
	1994–1998	2151	834	38.8	–11.9
	1999–2003	2956	1308	44.3	94.9
	2004–2008	4506	1901	42.2	118.6
	1977–2008				22.8

a) The classification is based on Fang (2000). b) Dominant species: *Picea*, *Abies*, *Larix*. c) Dominant species: *Pinus koraiensis*, *P. sylvestris* var. *mongolica*, *P. tabulaeformis*, *P. armandii*, *P. densiflora*, *P. thunbergii*, *P. densata*, *P. griffithii*, and other pine forests, mixed coniferous forests, *Cupressus*, *Tsuga*, and other coniferous forests. d) Dominant species: *Populus*, *Betula*, *Tilia*, *Quercus*, *Fraxinus mandshurica*, *Juglans mandshurica*, *Phellodendron amurense*, *Paulownia*. e) Dominant species: *Pinus massoniana*, *P. yunnanensis*, *P. kesiya* var. *langbianensis*, *Cunninghamia lanceolata*, *Cryptomeria*, *Sassafras*, *Keteleeria*, mixed coniferous and deciduous forests, *Metasequoia*, nonmerchantable woods. f) Dominant species: *Cinnamomum*, *Phoebe*, *Eucalyptus*, *Casuarina*, tropical forests, mixed broad-leaved forests, other hard broad-leaved forests, other soft broad-leaved forests, and coppice forests.

were grouped into three age classes: young-aged, mid-aged, and old-aged forests (including premature forests, mature forests, and over-mature forests). The temporal change of areas, C stocks, and C densities among three age classes in China's forest stands are shown in Figure 4. About 54.4%–55.0% of total biomass C stock of forest stands occurred in the old-aged forests, which had sequestered C of 930 Tg, at an average rate of 34.5 Tg C a⁻¹. The mid-aged forests had sequestered C of 391 Tg (22.9%–32.6% of total C stock of forest stands), at an average rate of 14.5 Tg C a⁻¹. The smallest biomass C sink (388 or 14.4 Tg C a⁻¹) was in the young-aged forests (12.8%–17.0% of total C stock).

These C sinks were mainly attributed to the increase in forest area in each age class: from 1977 through 2008, forest area has increased by 12.73×10⁶ ha in the old-aged forests, 10.76×10⁶ ha in the young-aged forests, and 8.60×10⁶ ha in the mid-aged forests, which accounted for 39.7%, 33.5%, and 26.8% of total area increment in China's forest stands. In addition, forest regrowth in each age class over the past three decades was also responsible for these C sinks: compared to the initial C density in the early 1980s (14.4, 35.4 and 67.4 Mg C ha⁻¹ in the young-, mid- and old-aged forests, respectively, for 1977–1981), the area-weighted mean biomass C density has increased 31.3% in the young-aged

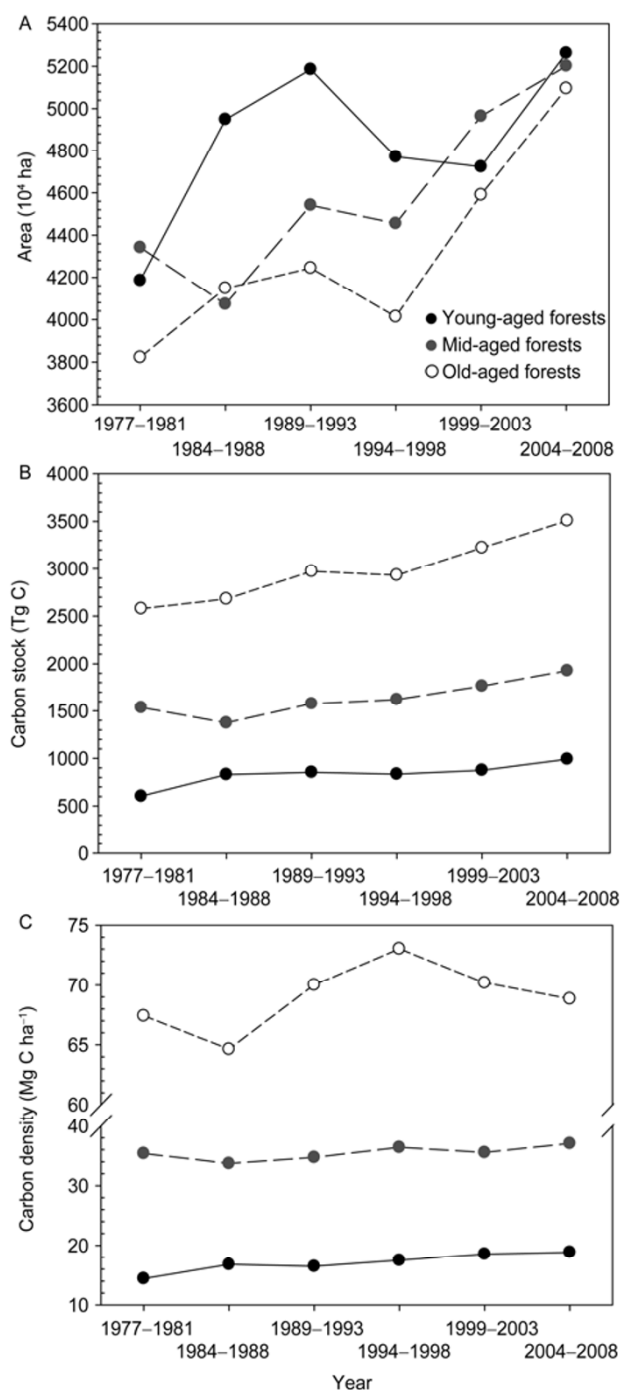


Figure 4 Area (A), C stock (B), and C density (C) of China's forest stands among different age classes during 1977–2008.

forests, 4.8% in the mid-aged forests, and 2.1% in the old-aged forests by the late 2000s (2004–2008), respectively. Note that the mean biomass C density of the old-aged forests ($68.8 \text{ Mg C ha}^{-1}$) was 1.9 and 3.7 times that of the mid-aged ($37.1 \text{ Mg C ha}^{-1}$) and young-aged forests ($18.9 \text{ Mg C ha}^{-1}$) in 2004–2008, indicating that the young- and mid-aged forests could sequester more C in the future if we keep these forests growing.

3 Discussion

3.1 Estimates of biomass C sink in China's forest

In this study, we estimated that the annual biomass C sink in China's forests averaged 70.2 Tg C during the period of 1977–2008, of which 63.3 Tg C was in the forest stands, 4.0 Tg C in the economic forests, and 2.9 Tg C in the bamboo forests. Using the linear conversion equations for estimating provincial biomass C stocks before 1994, Fang et al. [18] reported that the mean biomass C sink of China's forest stands was 75.2 Tg C a^{-1} during the period of 1977–2003, which was greater than our estimate (63.3 Tg C). This may be because the linear conversion equations could underestimate biomass C stocks of forest stands before 1994 when converting provincial biomass C stocks from the old criterion ($>30\%$ of canopy coverage) to the new criterion ($\geq 20\%$ canopy coverage). Using the new forest criterion and updated China's forest inventory data, Pan et al. [11] estimated that the mean biomass C sink of China's forests was 60 and 115 Tg C a^{-1} for the periods of 1990–1999 and 2000–2007, both were close to our estimates: in our study, the value was 60 Tg C a^{-1} during 1989–1998 (the average of the periods 1989–1993 and 1994–1998; Table 1) and 109 Tg C a^{-1} during 1999–2008 (the average of the periods 1999–2003 and 2004–2008; Table 1).

On the other hand, using the mean shoot numbers of bamboo, Pan et al. [26] estimated that the bamboo biomass C stock increased from 65 to 80 Tg C during 1977–1993, with a net C accumulation of 15 Tg C . Compared with this, we estimated that the C stock of the bamboo forests increased from 121 to 138 Tg C during 1977–1993, with a net accumulation of 17 Tg C , using the mean biomass densities of bamboos. This shows that our estimates of the C stocks were greater than those by Pan et al. [26], but the strength of the C sink was quite close to each other.

In order to compare the strength of the C sink with other countries and regions, we estimated that the average value of biomass C sink in China's forests was $0.44 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ by using an average forest area of $159.0 \times 10^6 \text{ ha}$ during 1977–2008. Similarly, we estimated the averaged value of the C sinks in China's forests between two sequential inventory periods: $0.21 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ (1977–1988), $0.72 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ (1984–1993), $0.06 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ (1989–1998), $0.63 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ (1994–2003), and $0.66 \text{ Mg C ha}^{-1} \text{ a}^{-1}$ (1999–2008). As shown in Table 4, the overall strength of China's forest biomass C sink during the past three decades was smaller than that in the Northern Hemisphere countries, but since the middle 1990s, it had increased greatly ($0.63\text{--}0.66 \text{ Mg C ha}^{-1} \text{ a}^{-1}$), and was somewhat comparable with that of the United States and Russia, although it was still lower than that of Europe and Japan.

Table 4 Estimated biomass C sink strength ($\text{Mg C ha}^{-1} \text{a}^{-1}$) by country or region in the Northern Hemisphere for the periods of 1990–1999 and 2000–2007 (modified from Pan et al. [11])

Country or region	C sink capacity ($\text{Mg C ha}^{-1} \text{a}^{-1}$)	
	1990–1999	2000–2007
China	0.43	0.77
Canada	0.03	−0.21
United States ^{a)}	0.47	0.58
Europe	1.09	1.31
Japan	1.01	0.99
Russia	0.33	0.61

a) The United States only include conterminous U.S. and Southeast Alaska.

3.2 Contribution of planted and natural forests to total biomass C sink

China holds the largest plantation area in the world, and therefore planted forests could play an important role in sequestering atmospheric CO_2 [3,27]. In our study, among a total of 1710 Tg C sequestration in China's forest stands during the year of 1977–2008, planted forests have sequestered C of 818 Tg, contributing almost half of total biomass C sink in China's forest stands (Table 2). Because the area and biomass density of planted forests have persistently increased since the early 1980s, forest expansion and regrowth are mainly responsible for this C sink. Similarly, Fang et al. [3] reported that planted forests had sequestered 0.45 Pg C during 1973–1998 because of forest expansion (from 17.4×10^6 to 23.1×10^6 ha) and regrowth (from 15.3 to $31.1 \text{ Mg C ha}^{-1}$).

Although the area of natural forests was about four times that of planted forests, natural forests have only sequestered the nearly same amount of C (892 Tg C) as did planted forests over the past 30 years. However, the mechanism responsible for the C sink in natural forests is different with that for the planted forests. During 1977–1998, although the area of natural forests decreased from 107.55×10^6 ha in 1977–1981 to 103.26×10^6 ha in 1994–1998, natural forests sequestered a net of 278 Tg C. That is to say, forest regrowth (from $41.5 \text{ Mg C ha}^{-1}$ in 1977–1981 to $46.0 \text{ Mg C ha}^{-1}$ in 1994–1998) is the main causes of the C sequestration. With the implementation of the Natural Forest Conservation Program since 1998 across the country, the area of natural forests had expanded since the late 1990s (1994–1998), and therefore forest expansion (from 103.26×10^6 ha in 1994–1998 to 115.59×10^6 ha in 2004–2008) was likely playing a major role in the increased C stocks in natural forests because of consistent biomass C density in natural forests during this period ($46.0 \text{ Mg C ha}^{-1}$ in 1994–1998 and $46.4 \text{ Mg C ha}^{-1}$ in 2004–2008).

3.3 Implication of forest C sink in China

Using the data of energy consumption and cement production recorded in “China Statistical Yearbook” and the similar methods mentioned in Fang et al. [41], we estimated

China's fossil-fuel CO_2 emissions as 27.7 Pg C during 1977–2008, at an average emission rate of 895 Tg C a^{-1} . Therefore, the biomass C sink of 70.2 Tg C a^{-1} in China's forests offsets 7.8% of the contemporary fossil CO_2 emissions in China, of which forest stands offsets about 7.1%. Pan et al. [11] reported that China's forest biomass C sink accounted for 44.4%–63.2% of total C sink in the country's whole forest sector (dead wood, harvested wood products, living biomass, litter, and soil) for the period of 1990–2007, which suggested the whole China's forests could sequester 111.1 – $158.1 \text{ Tg C a}^{-1}$, or equal 12.4%–17.1% of the contemporary fossil CO_2 emissions from 1977 through 2008, if the biomass C sink rate of 70.2 Tg C a^{-1} in China's forests was applied.

3.4 China's forest biomass C in the future

With the enhanced afforestation and reforestation practices, natural forest protection, and effective forest management, China's forest biomass C stocks will continuously increase in the next decades. Here we figure out the C sink potential for China's forest stands in the future, simply considering two factors, increase of forest area and forest regrowth.

3.4.1 Increase of forest area

The current forest coverage is 20.4% in China, of which forest stands is 16.2% (155.6×10^6 ha), with the mean biomass density of $41.3 \text{ Mg C ha}^{-1}$. According to the middle and long-term state forestry development plan, the forest coverage would be 24% in 2030 [42], which means that the area of China's forest stands (assuming the constant proportion of forest stands to total forests) will increase to 183.1×10^6 ha in the next two decades. Assuming an invariable mean biomass C density of forest stands ($41.3 \text{ Mg C ha}^{-1}$), the biomass C stock of forest stands will increase by 1135 Tg C, from 6427 Tg C at present to 7562 Tg C in 2030.

3.4.2 Forest regrowth

Within forest stands, planted forests occupy 25.7% of total forest stand area (40.0×10^6 ha) with a low mean biomass density of $26.7 \text{ Mg C ha}^{-1}$ in 2004–2008. However, the

mean biomass density of natural forests is up to 46.4 Mg C ha⁻¹ at the same period. When these planted forests grow up to the same biomass C density as natural forests did, these forests could sequester about 788 Tg C. In addition, young- and mid-aged forests occupy 67.2% of total forest stand area (52.6×10⁶ ha for the former and 52.0×10⁶ ha for the latter) with the mean biomass C density of 18.9 and 37.1 Mg C ha⁻¹, respectively. If these young- and mid-aged forests could grow up to be as old as the old-aged forests (currently its mean biomass C density is 68.8 Mg C ha⁻¹), these forests will sequester C of 4273 Tg in the future. In other words, even without an increase in forest area, China's forests will still sequester considerable amounts of CO₂ from the atmosphere because of their regrowth.

3.5 Error analyses

Error analysis of the estimates of national forest biomass C stocks based on inventory data was rarely reported, mainly because many sources of error make it complex. For the forest stands, the most important errors may come from forest inventory data and the estimation of regional biomass stocks by applying for BEFs. For example, Phillips et al. [43] analyzed the error in estimates of forest timber volumes and their changes in five southeastern states in the United States. In their study, the error sources included sampling error, measurement error and regression error, of which sampling error was mainly responsible for the total error in forest timber volumes and their changes, accounting for 90%–99% of total variance. However, the estimation of forest timber volumes is only the first step to estimate forest biomass stocks, and the regression error in subsequent steps that convert timber volume to biomass need to be further investigated [43]. Forest inventory data used in our study specified the precision requirement in sampling design: the forest area and volume precision were required to be >90% in almost each province (>85% in Beijing, Shanghai and Tianjin) [44]. In addition, the *R* square values of BEF equations used to convert timber volume to biomass were above 0.8 for most dominant tree species (Table S1). Therefore, the data and method used in this study have relatively high precision. For example, Fang et al. [22] reported that the estimation error of biomass stocks in the national level should be less than 3%.

For the economic and bamboo forests, the major error may be generated from the use of the mean biomass density. In general, the method could result in some overestimation of biomass stocks of economic and bamboo forests [34] because collected data were usually from forest stands with better growing conditions. Additional, using the constant C conversion factor of 0.5 might also yield some errors.

This study does not include other forest components recorded in China's forest inventory data, including farmland protection forest, and four-side greening trees (trees out forests). In the past decades, the farmland protection forest

and four-side greening trees have significantly increased in China, perhaps resulting in an increase of their C stocks, which need to be assessed in the future [45].

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- 1 Kauppi P E, Mielikainen K, Kusela K. Biomass and carbon budget of European forests, 1971 to 1990. *Science*, 1992, 256: 70–74
- 2 Dixon R K, Brown S, Houghton R A, et al. Carbon pools and flux of global forest ecosystems. *Science*, 1994, 263: 185–190
- 3 Fang J Y, Chen A P, Peng C H, et al. Changes in forest biomass carbon storage in China between 1949 and 1998. *Science*, 2001, 292: 2320–2322
- 4 Fang J Y, Brown S, Tang Y H, et al. Overestimated biomass carbon pools of the northern mid- and high latitude forests. *Clim Change*, 2006, 74: 355–368
- 5 Pacala S W, Hurtt G C, Baker D, et al. Consistent land- and atmosphere-based US carbon sink estimates. *Science*, 2001, 292: 2316–2320
- 6 Goodale C L, Apps M J, Birdsey R A, et al. Forest carbon sinks in the Northern Hemisphere. *Ecol Appl*, 2002, 12: 891–899
- 7 Janssens I A, Freibauer A, Ciais P, et al. Europe's terrestrial biosphere absorbs 7 to 12% of European anthropogenic CO₂ emissions. *Science*, 2003, 300: 1538–1542
- 8 Bonan G B. Forests and climate change: forcings, feedbacks, and the climate benefits of forests. *Science*, 2008, 320: 1444–1449
- 9 Yang T H, Song K, Da L J, et al. The biomass and aboveground net primary productivity of *Schima superba*-*Castanopsis carlesii* forests in east China. *Sci China Life Sci*, 2010, 53: 811–821
- 10 Zhang Q Z, Wang C K. Carbon density and distribution of six Chinese temperate forests. *Sci China Life Sci*, 2010, 53: 831–840
- 11 Pan Y, Birdsey R A, Fang J Y, et al. A large and persistent carbon sink in the world's forests. *Science*, 2011, 333: 988–993
- 12 He J S. Carbon cycling of Chinese forests: from carbon storage, dynamics to models. *Sci China Life Sci*, 2012, 55: 188–190
- 13 Brown S, Sathaye J, Cannell M, et al. Mitigation of carbon emission to the atmosphere by forest management. *Commonw Forest Rev*, 1996, 75: 80–91
- 14 Brown S L, Schroeder P, Kern J S. Spatial distribution of biomass in forests of the eastern USA. *Forest Ecol Manage*, 1999, 123: 81–90
- 15 Brown S L, Schroeder P E. Spatial patterns of aboveground production and mortality of woody biomass for eastern U.S. forests. *Ecol Appl*, 1999, 9: 968–980
- 16 Hu H F, Wang G G. Changes in forest biomass carbon storage in the South Carolina Piedmont between 1936 and 2005. *Forest Ecol Manage*, 2008, 255: 1400–1408
- 17 Fang J Y, Oikawa T, Kato T, et al. Biomass carbon accumulation by Japan's forests from 1947 to 1995. *Global Biogeochem Cycles*, 2005, 19, GB2004, doi: 10.1029/2004GB002253
- 18 Fang J Y, Guo Z D, Piao S L, et al. Terrestrial vegetation carbon sinks in China, 1981–2000. *Sci China Ser D-Earth Sci*, 2007, 50: 1341–1350
- 19 Marland G, Andres R J, Boden T A. Global, regional, and national CO₂ emissions. In: Boden T A, Kaiser D P, Sepanski R J, et al., eds. *Trends '93: a compendium of data on global change*. ORNL/CDIAC-65. Oak Ridge: Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, 1994. 505–584
- 20 Ministry of Forestry of China. *Forest Resource Report of China—the 7th National Forest Resources Inventory* (in Chinese). Beijing: China Forestry Publishing House, 2009. 2

- 21 Fang J Y, Tang Y H, Son Y. Why are East Asian ecosystems important for carbon cycle research. *Sci China Life Sci*, 2010, 53: 561–565
- 22 Fang J Y, Liu G H, Xu S L. Biomass and net production of forest vegetation in China (in Chinese). *Acta Ecol Sin*, 1996, 16: 497–508
- 23 Fang J Y, Wang G G, Liu G H, et al. Forest biomass of China: an estimation based on the biomass-volume relationship. *Ecol Appl*, 1998, 8: 1084–1091
- 24 Liu G H, Fu B J, Fang J Y. Carbon dynamics of Chinese forests and its contribution to global carbon balance (in Chinese). *Acta Ecol Sin*, 2000, 20: 733–740
- 25 Wang X K, Feng Z W, Ouyang Z Y. The impact of human disturbance on vegetative carbon storage in forest ecosystems in China. *Forest Ecol Manage*, 2001, 148: 117–123
- 26 Pan Y D, Luo T X, Birdsey R, et al. New estimates of carbon storage and sequestration in China's forests: effects of age-class and method on inventory-based carbon estimation. *Clim Change*, 2004, 67: 211–236
- 27 Piao S L, Fang J Y, Zhu B, et al. Forest biomass carbon stocks in China over the past 2 decades: estimation based on integrated inventory and satellite data. *J Geophys Res*, 2005, 110, G01006, doi:10.1029/2005JG000014
- 28 Piao S L, Fang J Y, Ciais P, et al. The carbon balance of terrestrial ecosystems in China. *Nature*, 2009, 458: 1009–1013
- 29 Piao S L, Ito A, Li S G, et al. The carbon budget of terrestrial ecosystems in East Asia over the last two decades. *Biogeosciences*, 2012, 9: 3571–3586
- 30 Xu X L, Cao M K, Li K R. Temporal-spatial dynamics of carbon storage of forest vegetation in China (in Chinese). *Progr Geogr*, 2007, 26: 1–10
- 31 Wu Q B, Wang X K, Duan X N, et al. Carbon sequestration and its potential by forest ecosystems in China (in Chinese). *Acta Ecol Sin*, 2008, 28: 517–524
- 32 Liu S N, Zhou T, Wei L Y, et al. The spatial distribution of forest carbon sinks and sources in China. *Chin Sci Bull*, 2012, 57: 1699–1707
- 33 Fang J Y, Wang Z M. Forest biomass estimation at regional and global levels, with special reference to China's forest biomass. *Ecol Res*, 2001, 16: 587–592
- 34 Guo Z D, Fang J Y, Pan Y D, et al. Inventory-based estimates of forest biomass carbon stocks in China: a comparison of three methods. *Forest Ecol Manage*, 2010, 259: 1225–1231
- 35 Brown S, Lugo A E. Aboveground biomass estimates for tropical moist forests of Brazilian Amazon. *Interciencia*, 1992, 17: 8–18
- 36 Schroeder P, Brown S, Mo J, et al. Biomass estimation for temperate broadleaf forests of the United States using inventory data. *Forest Sci*, 1997, 43: 424–434
- 37 Fang J Y, Chen A P, Zhao S Q, et al. Estimating biomass carbon of China's forests: supplementary notes on report published in *Science* (2001, 291: 2320–2322) by Fang et al. 2001 (in Chinese). *Acta Phytocool Sin*, 2002, 26: 243–249
- 38 Fang J Y, Chen A P. Dynamic forest biomass carbon pools in China and their significance (in Chinese). *Acta Botan Sin*, 2001, 43: 967–973
- 39 Fang J Y, Shen Z H, Tang Z Y, et al. Forest community survey and the structural characteristics of forests in China. *Ecography*, 2012, 35: 1059–1071
- 40 Fang J Y. Forest productivity in China and its responses to global climate change. *Acta Phytocool Sin*, 2000, 24: 513–517
- 41 Fang J Y, Liu G H, Xu S L. Carbon cycling in terrestrial ecosystems in China. In: Wang G C, Wen Y P, eds. *Emissions and Their Relevant Processes of Greenhouse Gases in China* (in Chinese). Beijing: Chinese Environmental Science Publishers, 1996. 81–149
- 42 Xu B, Guo Z D, Piao S L, et al. Biomass carbon stocks in China's forests between 2000 and 2050: a prediction based on forest biomass-age relationships. *Sci China Life Sci*, 2010, 53: 776–783
- 43 Phillips D L, Brown S, Schroeder P E, et al. Towards error analysis of large-scale forest carbon budgets. *Global Ecol Biogeogr*, 2000, 9: 305–313
- 44 Xiao X W, ed. *Forest Resource Inventory of China* (in Chinese). Beijing: China Forestry Publishing House, 2005
- 45 Guo Z D. Biomass carbon stocks and ecosystem carbon budget in China's forests (in Chinese). Dissertation for Doctoral Degree. Beijing: Peking University, 2011

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Supporting Information

Table S1 Parameters of the continuous biomass expansion factor (BEF) method for China's major forest types (modified from Guo et al. [34])

The supporting information is available online at life.scichina.com and www.springerlink.com. The supporting materials are published as submitted, without typesetting or editing. The responsibility for scientific accuracy and content remains entirely with the authors.