Circulation of Sodium Sulfate Solution Produced During NiMH Battery Waste Processing



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Abstract

Hydrometallurgical recovery of rare earth elements (REE) from NiMH battery waste can be performed using sulfuric acid leaching followed by selective precipitation as double salt (REENa(SO_4)₂·H₂O) by adding Na₂SO₄ as a precipitating agent. The formed double salts can then be further converted with NaOH solution to form REE hydroxides. However, present literature has paid little attention to how to minimize the consumption of reagents in this recovery process. In the current study, a process model was built using HSC Sim software, and metamodeling of the simulated flowsheet was conducted with the design of experiments (DOE), after which the basic functionality of the flowsheet was verified by experimental work, starting from sulfuric acid leaching of spent batteries. As the process is based on chemical circulation internally in the process, the objective was to investigate the accumulation of the elements in solution, specifically Na⁺, K⁺, $SO_4^{2^-}$ ions and impurities liberated in the NaOH conversion step. It was found out that metamodeling could be utilized in obtaining first estimate of the process behavior, prior to any experiments. Flowsheet functionality with circulated solution was confirmed, obtaining at best 96.9% La recovery.

Keywords Flowsheet metamodeling · Hydrometallurgy · Recycling · Battery waste · Rare earths · Response surface methodology

1 Introduction

Nickel metal hydride (NiMH) battery waste can be considered as an urban mineral due to the presence of valuable metals like Ni, Co, and rare earth elements (REE) [1]. In NiMH battery waste leaching experiments, sulfuric acid has often been employed as a lixiviant [2–5]. The key advantages of sulfuric acid are its stability and cost-effectiveness compared to other mineral acids such as HCl and HNO₃. However, the usage of sulfuric acid in NiMH battery waste leaching poses simultaneous challenges and advantages due to limited REE sulfate solubility. Conversely, limited REE solubility can be taken as benefit when recovering REEs double sulfates after solidliquid separation. For example, the least soluble REE sulfate, lanthanum, has the solubility of 23 g/L in water [6]. However, sodium and potassium hydroxide, present in electrolyte residues in the battery waste, can readily cause in situ precipitation during leaching due to formation of REE double sulfates, as per Eq. 1, which have a significantly more limited solubility (La 2.32 g/L at T = 20 °C in H₂O) [7, 8], thus decreasing the

$$RE^{3+} + 2SO_4^{2-} + Na^+ + H_2O \rightarrow RENa(SO_4)_2.H_2O \downarrow$$
 (1)

It must be noted that the chemistry behind the precipitation is more complex than what reaction (1) reveals, and most likely involves different complexes, such as LaSO₄⁺ and



total REE recovery [5]. However, the problem of in situ REE precipitation during the NiMH battery leaching can be circumvented by washing the Na and K containing electrolyte residues off the raw materials. Subsequently, this phenomenon can be taken of advantage in REE recovery after solid-liquid separation of battery waste and REE rich solution: sodium ions may be added into the pregnant leach solution (PLS) in order precipitate the REEs in bulk [9, 10] as a mixed double salt out from the solution at very low pH with minimal coprecipitation of Ni and other metals [11]. As shown in our previous study, the obtained precipitate is very crystalline and relatively pure and was obtained from acidic solution [5]. Before the dawn of solvent extraction in the wake of Manhattan Project, this was a commonly employed method of fractional precipitation of REEs [12]. This relatively easy separation method can still be extremely advantageous in the case of such a complex solution matrix with relatively high light REE concentrations (conc.), shown in Eq. 1:

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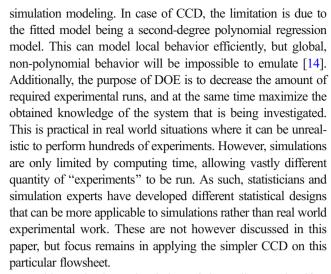
 ${\rm La(SO_4)_2}^-$ depending on ion activities and temperature. For example, the solubility of double salts has been shown to change in a complex manner based on ${\rm H_2SO_4}$ concentration and subsequent bisulfate formation [7]. Another example is how the double sulfate solubility strongly decreases in the presence of excess sodium [11]. However, for the purposes of this study, the reaction is deemed sufficient as it is presented in this manner in other studies as well [8, 11], and can be utilized in calculating the process mass balance.

Once REEs have been precipitated as double sulfates as per Eq. 1 [5], the resulting double salt precipitate lacks applications without further treatment. Direct thermal treatment would generate sulfur dioxide emissions and likely require high temperatures in order to transform double sulfates into REE oxides. Furthermore, the precipitate contains Na ions, which would need to be removed by chemical treatment from the calcined REE product.

In the context of circularity of raw materials, it can be taken as a remarkable benefit if the sodium and sulfur present in the double salt could be reused instead of losing them into the process waste streams. In order to attempt to minimize reagent consumption, an aspect ignored in present NiMH recycling literature, flowsheet modeling, metamodeling, and experiments were utilized as tools to explore this subject. In this experimental and flowsheet modeling study, the double salts are converted by dissolution and precipitation in alkaline solution to hydroxides as shown by Abreu et al. [13] (Eq. 2) which are readily treatable, e.g., with total dissolution and solvent extraction, for further purification:

$$RENa(SO4)2.H2O + 3NaOH \rightarrow RE(OH)3 + 2Na2SO4 + H2O$$
 (2)

Simulation modeling is a practice of metamodeling. HSC Sim is based on modular process flowsheet modeling, and as such, it can be hard to derive equations of the model for optimization. In metamodeling, guided by principles of design of experiments (DOE), such flowsheet model is considered to be a black box: the user is aware of different inputs and the unit process (module) will produce the possible outputs without knowing anything about the module itself. Flowsheet simulation can also be either stochastic or deterministic [14]. In this paper, metamodeling of deterministic simulation is examined by applying response surface methodology [15] in conjunction with HSC Sim, a modular process flowsheet simulation tool, which is an ideal target for such optimization investigations [16]. In this way, information of potentially complex inter-dependencies of process parameters can be attempted to explore in a more rigorous manner at process-wide level, and optimization can be attempted at localized parameter level [17]. As the original DOEs, such as the central composite design (CCD), were developed for real world applications, there are limitations in applying them to



In this work, the recirculation of the sodium and sulfate ions present in these double salts back to precipitation process was investigated. Several works have focused on recovery process, but little attention has been paid to resource efficiency issues caused by waste side streams. First, a process simulation flowsheet was created by using HSC Chemistry's Simulation module. Unit processes were created for each respective processing based on literature data, each step involving the formation and further treatment of the REE precipitate. The solutions generated during the REE double salt dissolution in NaOH were circulated back to the precipitation stage. The simulation was used to run several scenarios, where the acidity and REE contents of the pregnant leach solution (PLS) were varied, along with several other parameters that are critical to double salt precipitation, dissolution, and recirculation. By varying these parameters, observation on mass balance were performed. DOE was utilized in analyzing this data and in generating a metamodel of the flowsheet investigated. A quadratic model was fitted to a CCD based on the experimental design runs of the simulation, which also takes into account the second-degree interactions [15]. The created regression model was utilized in attempting to identify the optimal parameters for the recirculation of Na₂SO₄ solution. Subsequently, a round of chemical circulation experiments were performed in laboratory, starting from sulfuric acid leaching of NiMH battery waste. Losses borne in wash waters were recorded and dilution of the PLS measured. These experiments were performed in order to confirm the behavior observed in flowsheet modeling and metamodeling.

2 Materials and Methods

2.1 Flowsheet Modeling

Outotec's HSC 9.4.1 [18] was used in creating the flowsheet simulation model and published literature data. Minitab 17.1



[19] was used in creating the quadratic regression models based on DOE. The flowsheet simulation model was composed of several unit processes, each handling their own specific processes. The general schematic of the flowsheet is shown in Fig. 1 and the most important functions and controllable parameters for each unit process are summarized in Table 1. The flowsheet can be divided into two parts: (I) the double salt precipitation (Na₂SO₄ Prep. Tank, Precipitation Reactor #1, and Filter #1) and (II) the conversion of double salt into hydroxide (NaOH Preparation Tank, Alkaline Conversion Reactor #1, and Filter #2). In addition, there is an additional unit process (bleed) for bleeding the solution in order to achieve balanced Na concentration in the main process.

In the Precipitation Reactor #1, REEs are precipitated as double sulfates. The simulated unit process here contains an equation to simulate precipitation yield of La as a function of Na/REE ratio, presented in Fig. 2 and Eq. 3. This equation is partially based on results of Kul *et al.* [20]. However, the precipitation kinetics and equilibrium are strongly depended

on PLS ion concentrations and temperature which should be taken into consideration when modeling a unit process. The present unit process is modeled only at constant T and t. This equation is not intended as an extremely accurate reflection of reality, but to mirror precipitation behavior. For example, for the purposes of non-linear regression fitting, a synthetic data point was added with assumption that when Na/REE = 0.5, La recovery % = 50%.

Yield =
$$100-104.317 \cdot e^{-1.77958 * \left\{ \frac{Na}{REE} \right\}}$$
 (3)

Alkaline conversion Reactor #1 does not contain a model akin to Eq. 3, but the reaction was assumed to progress to completion in 1 h according to Eq. 2 based on earlier work by Abreu and Morais (2010) [13].

2.2 Flowsheet Metamodeling

Face-centered central composite design (CCF) was utilized in the creation of regression models of the responses. The design

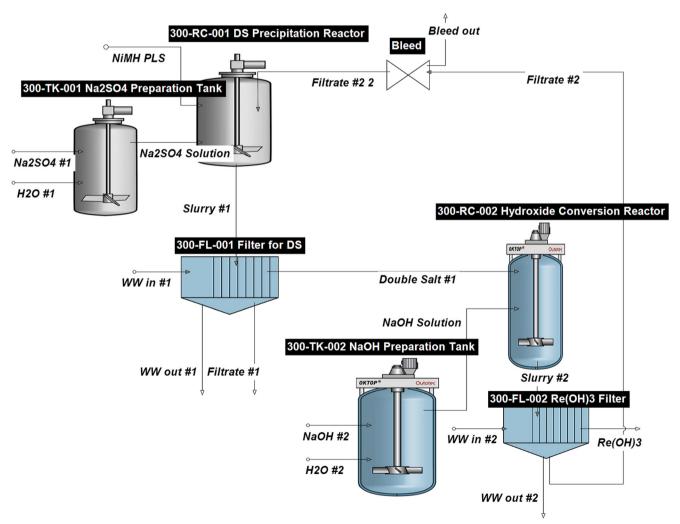


Fig. 1 Flowsheet of REE recovery step of PLS produced in leaching of spent NiMH batteries

Table 1 Short summary of the contents of the unit processes and the investigated parameters

Unit process name	Function	Controllable parameters
300-RC-001 DS Precipitation Reactor	Precipitation of double salts from the solution. Contains a regression model for REE precipitation as a function of Na/REE. In the model, a 100 L/h PLS feed control was created.	PLS REE conc. (g/L) PLS acid conc. (mol/L) PLS total volume (L/h) Initial Na/REE ratio (mol/- mol)
$\begin{array}{c} 300\text{-TK-}001 \\ Na_2SO_4 \\ Preparation \ Tank \end{array}$	Preparation of fresh Na ₂ SO ₄ solution. Controls the amount of addition of Na ⁺ in conjunction with Precipitation Reactor #1 and Filtrate #2.	Molarity of Na ₂ SO ₄ solution (mol/L)
300-FL-001 Filter for DS	Generic Outotec Filtration unit	Cake moisture (%) Washing efficiency (%)
300-RC-002 Hydroxide Conversion Reactor	Controls the dissolution of double salt and precipitation of hydroxides. The slurry moves into Filter #2.	3NaOH/REE ratio (mol/mol) L/S ratio (mL/g)
300-TK-002 NaOH Preparation tank	Preparation tank for NaOH solutions. It will adjust the NaOH volume based on feedback from Alkaline Conversion Reactor #1.	
300-FL-002 Filter for RE(OH) ₃	Generic Outotec Filtration unit	Cake moisture (%) Washing efficiency (%)
Bleed junction	Responsible for bleeding a fraction of the recirculated solution out of the process. This is included in the model in order to investigate how bleeding affects the mass balances of other unit processes.	Bleed fraction (%)

was composed of full two-level factorial design (2^7). It is possible to use fractional factorial designs to create a CCD with resolution V; however, there was no reason for doing so in the computer modeling in this particular instance. The design is augmented with star points at equidistance α from the center of the design. In CCF models, the distance of $\alpha = 1$. The center points were replicated (18 times); however, the rest of the model was not as the replication of a deterministic model would not introduce any benefits due to the fact that there is no variation between replicates. The data obtained from the simulation was fitted into CCD in Minitab. The

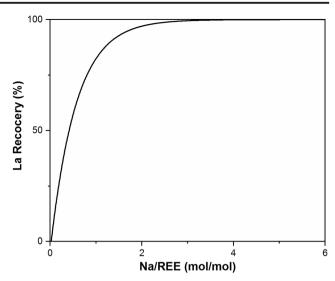


Fig. 2 An emulated precipitation curve fitted to experimental and synthetic data [20]

flowsheet model parameters, and hence model constraints, used to create the design of experiments are presented in Table 2. The table only represents model constraints, not all the possible combinations of experiments. The cube points are experimental points of standard x^2 full factorial design with center points. Axial points are the extension of the factorial design into CCD [15].

2.3 Experimental Verification of the Process Flowsheet

Circulation experiments were performed in laboratory in order to demonstrate its viability in precipitation of REEs from the PLS with the recyclant solution. Leaching of sieved NiMH battery waste of - 0.5 mm size fraction was performed at room temperature in 2 M H₂SO₄. A 1 L glass reactor was utilized with mechanical agitation (VWR VOS16, ca. 300 rpm). The sieved underflow sample weighed 45.8 g. The density of the material was determined via water displacement method and was measured to be ca. 4.7 g/cm³. S/L ratio of 1/10 was utilized. The S/L was calculated as a ratio of sample mass and total volume of initial slurry: 448 mL of 2 M acid was added into reactor. The sample was added carefully into the acidic solution due to potentially dangerous gas formation of H₂ gas and runaway exothermic reaction. Once added, leaching was let to continue for 2 h after which PLS was filtered for the circulation experiments. The leach residue was washed with 100 mL distilled water (100 mL vs 4.02 g). Both the PLS and wash water composition are shown in Table 3. All the solutions obtained were analyzed by inductively coupled plasma-optical emission spectroscopy (ICP-



Parameter code	Parameter	Levels					
		Cube – 1	Cube + 1	Center	Axial -1	Axial + 1	
P1	[Na ₂ SO ₄]	0.5	2	1.25	0.5	2	
P2	Targeted Na/REE mol ratio	1	3	2	1	3	
P3	PLS [acid] (mol/h)	0.5	4	2.25	0.5	4	
P4	PLS [REE] (mol/h)	12	48	30	12	48	
P5	L/S ratio	12.5	100	56.25	12.5	100	
P6	3NaOH/REE (mol ratio)	1	3	2	1	3	
P7	Bleed fraction (%)	0	100	50	0	100	

Table 2 Parameter levels utilized to model REE recovery and conversion to hydroxides. The levels shown were used in the response surface design to create a fully quadratic model with second-degree interaction terms

OES, Perkin Elmer Optima 7100DV). Sodium was analyzed with flame atomic absorption spectroscopy (FAAS, Varian A240).

Initially, the maximum content of REEs was estimated by adding excessive quantity of Na₂SO₄. Eighty milliliters of PLS was mixed with 20 mL of 355 g/L Na₂SO₄ solution. The collected, dry precipitate was weighed (4.14 g) and taken as a baseline (4.14 g of DS per 80 mL of PLS) for a complete REE separation from the PLS, as REEs are known to precipitate in the presence of excessive amount of Na and SO₄ [5]. This was later confirmed with the chemical analysis. Based on the mass of obtained double salt, the solution concentration of total REEs was calculated to be roughly 19.5 g/L.

In the circulation experiments, 40 mL of the recyclant was added to 60 mL of PLS. Leach residue and all precipitates were washed with distilled water. The wash water was heated up to 60 °C. Vacuum filtration and filtration assembly was utilized in precipitate recovery and washing. Recycling was performed twice. In both double salt precipitation and rare earth hydroxide conversion, T=60 °C was utilized for 2 h in a jacketed 100 mL glass reactor. Magnetic stirring was supplied to the reactor. The calculated, Na/REE ratio in precipitation was roughly 2.32 and 2.04 in round 1 and round 2, respectively. The ratio was calculated based on known mass of the double sulfate and known quantity of Na in NaOH solution. In rare earth hydroxide conversion, L/S was not controlled, but 100 mL

NaOH was mixed with the produced double salts. Mole ratio of 3NaOH/REE = 1.6 was utilized, equaling to ca. 0.35 M 100 mL NaOH solution.

3 Results and Discussion

It can be seen from Eqs. 1 and 2 that there is an imbalance between the stoichiometric requirements of Na⁺, OH⁻, RE³⁺, and SO₄²⁻ ions. In each process cycle, according to Eq. 1, 1 mol of Na was required to produce 1 mol of double salt. Subsequently, additional 3 mol of Na per 1 mol of REE were added into the system during hydroxide conversion stage, resulting in 4 mol of Na per 1 mol of REE ratio in the process. However, it is commonly known that REE double salts are slightly soluble in sulfuric acid [7], and their precipitation efficiency can be improved via common ion effect by having an excess of sodium sulfate in the solution. It has been reported that to achieve > 95% double salt precipitation efficiency, a stoichiometric ratio (Na/REE) of 1-3 is required when operating at metal concentrations typical to NiMH battery recycling [20]. It is therefore advantageous for the process that ratio of 4 mol Na per 1 mol of REE is maintained in the process circulation system.

Initially, fresh Na₂SO₄ is injected in the double salt precipitation tank to start the process. Based on Eqs. 1 and 2, Na/REE molar ratio of 1:4 can be achieved in one complete circulation cycle of the process. After the

Table 3 PLS and leach residue wash water analysis. The results are reported in mg/L

Solution	Ni	Со	Mn	Al	Zn	Fe	La	Na
LR wash water (mg/L)	3363	512	123	38.6	47.1	29.5	944	115
PLS (mg/L)	57752	7386	3528	1259	1857	601	9796	86

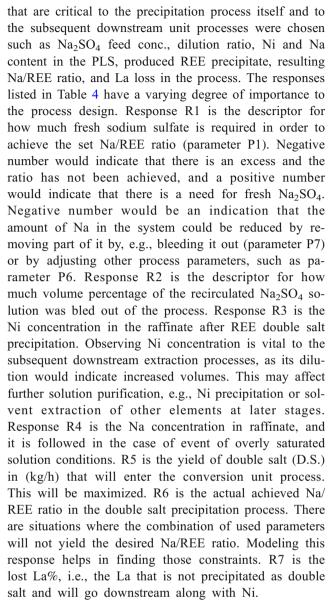


initial batch of DS is produced, the sodium sulfate originating from the added NaOH in the double salt conversion to hydroxide is enough to fulfil the requirement for the double salt precipitation because of presence of the excess stoichiometry of Na with respect to REE (4:1 Na/REE ratio). However, this will come at a cost of dilution of the main PLS. This is undesirable, as it will reduce the concentrations of valuable metals in the leach solution and ultimately, push the downstream processes to handle larger volumes. Furthermore, La loss will also increase in more diluted PLS due to its relatively high solubility in water and sulfuric acid, as shown by Lokshin et al. (2.32 g/L in H₂O at 20 °C) [7]. However, in this process flowsheet model, only stoichiometric ratio of Na and REE was taken into consideration in DS precipitation efficiency, not REE solubility, and same precipitation efficiency was assumed regardless of the REE concentration in the solution. By adjusting the L/S ratio (liters of NaOH solution per kg of double salt) used in the double salt conversion to hydroxide, the magnitude of dilution factor can be affected. This trade-off of optimal L/S ratio will be investigated in this flowsheet simulation, and a regression model was built in order to find an optimal bleed fraction for different situations.

3.1 Flowsheet Metamodeling Results

The flowsheet model presented here has been simplified with several assumptions. (1) La is selected to represent all REEs as a group. (2) Ni is representing other metals that are soluble and will not be recovered in double salt precipitation stage. Ni concentration will act as an indicator of dilution due to recirculation. (3) Due to lack of electrolyte density data in the literature, the effect that the dissolved La will have on the density of the electrolyte will be estimated by using the value of Al₂(SO₄)₃ [21]. (4) Energy balance will not be considered in the model since the thermodynamic formation values for lanthanum double salt (LaNa(SO₄)₂·H₂O) were not found in the literature. It needs to be noted that the fitted regression model equations are not based on chemistry confirmed by the experimental work, but the literature.

The parameters given in Table 2 were used in Minitab to generate a DOE composed of CCD. The experimental matrix was entered into HSC SIM, and the resulting data originated from HSC simulation using process flowsheet shown in Fig. 1 was again entered back to Minitab. As a result, several models were created, each corresponding to a response. These were fully quadratic models that could be used to predict the responses in the whole precipitation system. Responses



Firstly, quadratic regression models for each response (R1–R7) were created. CCD provides a quadratic model of following form:

$$Y = \beta_0 + \sum_{i=1}^{k=7} \beta_i \cdot P_i + \sum_{i=1}^{k=7} \beta_{ii} \cdot P_i^2 + \sum_{1 \le i \le j}^{k=21} \beta_{ij} \cdot P_i P_j + \epsilon$$
 (4)

where β_i are the regression coefficients for the parameters P_i [15]. The first term is the constant coefficient β_0 , the second term represents the linear effects (7 in total), third term represents the polynomial effects (7 in total), and fourth term represents the second-order interactions (21 in total). In total, 35 terms consisting of linear, binomial, and second-order interactions must be considered. This is exemplified by the uncoded regression equation for raffinate Ni concentration, shown in Eq.5:



(5)

Table 4 The selected responses and the stream or unit process from where the value of the response was recorded

Response number	R1	R2	R3	R4	R5	R6	R7
Response content	Fresh Na ₂ SO ₄ needed (kg/h)	Dilution Factor (multiplier)	Raff. Ni (g/L)	Raff. Na (g/L)	D.S. produced (kg/h)	Equilibrium Na/REE (mol/mol)	La lost (%)
Unit process or stream where measured	Precipitation Reactor #1	Filtrate #1 Precipitation Reactor #1	Filtrate #1	Filtrate #1	Double salt	Precipitation Reactor #1	Input vs. output

29.33 - 7.5 Na2SO4 Conc. - 2.7 Na/REE - 3.3 3NaOH/REE + 0.434 Bleed% - 1.15 PLS [A] - 0.858 PLS [REE] + 0.590 L/S Ratio + 1.07 Na2SO4 Conc.*Na2SO4 Conc. + 0.48 Na/REE*Na/REE + 0.29 3NaOH/REE*3NaOH/REE - 0.00167 Bleed%*Bleed% + 0.171 PLS [A]*PLS [A] + 0.00895 PLS [REE]*PLS [REE] - 0.00293 L/S Ratio*L/S Ratio + 0.338 Na2SO4 Conc.*Na/REE + 0.522 Na2SO4 Conc.*3NaOH/REE + 0.10635 Na2SO4 Conc.*Bleed% + 0.540 Na2SO4 Conc.*PLS [A] + 0.0582 Na2SO4 Conc.*PLS [REE] - 0.0036 Na2SO4 Conc.*L/S Ratio + 0.586 Na/REE*3NaOH/REE - 0.06271 Na/REE*Bleed% + 0.159 Na/REE*PLS [A] - 0.0362 Na/REE*PLS [REE] + 0.00207 Na/REE*L/S Ratio + 0.00527 3NaOH/REE*Bleed% - 0.098 3NaOH/REE*PLS [A] + 0.0111 3NaOH/REE*PLS [REE] - 0.00520 3NaOH/REE*L/S Ratio - 0.00063 Bleed%*PLS [A] - 0.000667 Bleed%*PLS [REE] - 0.002207 Bleed%*L/S Ratio - 0.0103 PLS [A]*PLS [REE] + 0.00141 PLS [A]*L/S Ratio - 0.000943 PLS [REE]*L/S Ratio (5)

Due to the large amount of terms in the equations, the models for the other responses (R1–R2 and R4–R7), are not presented, instead, only results statistical analysis are shown.

In Table 5, the results are summarized for the created models for each investigated response. S is the standard error of regression and R^2 (pred) is the predicted value. R^2 (pred) is obtained by the software removing observed data points, after which the model is refitted and the removed data point is compared to the refitted model. It can be seen that the regression and predictive power is relatively high ($\geq 95\%$) for each response except for responses R1 (fresh Na₂SO₄ needed), R2 (dilution factor), R3 (Ni concentration in raffinate), and R5 (D.S. produced). This may be due to the fact that a quadratic model is not sufficient to predict La% losses due to many factors having simultaneous effect on the amount of La% being lost. This is an inherent weakness in DOE-only

Table 5 Regression coefficients of the each response surface

Response	S	R ² (pred)
R1	1.39967	94.63%
R2	0.875865	85.30%
R3	2.7286	87.83%
R4	0.682155	97.25%
R5	2.38869	78.59%
R6	0.219010	99.62%
R7	0.0649982	99.87%

approach to process simulation optimization [22]. The use of only cube and axial points will cause smoothing of the data due to small number of levels (3) each parameter has. It is also possible that there are higher than second-degree interactions, causing flawed regression fitting. The face-centered CCD employed in this study has issues in detecting quadratic effects. Further improvements may be achieved by using other designs than CCD [23].

Once fitted, the models can be analyzed. As shown in Table 4, there are several responses that need to be optimized in the flowsheet model. Optimization of these parameters may be difficult manually due to many interdependencies of the different feed parameters. This will be especially true in more complex flowsheets. Minitab allows multi-parameter optimization, however as a result, there may be a large set of optimal parameters corresponding to the goals. It is up to the human operator to choose the criteria and weights for the multiparameter optimization and its results. The goal (G1-G8) of the optimization in this study is summarized in Table 6. Certain responses are preferred to be minimized (raffinate volume. Na concentration in raffinate) or maximized (double salt yield, raffinate Ni concentration) or not optimized at all, based on the user preference. The optimization runs G1-G7 were single response optimizations which predict at what parameters values, it is possible to optimize one response at time. In optimization run G8, a multi-parameter optimization was performed. Simultaneously, two conflicting parameters were



Table 6 Response optimization targets. Not opt. = not optimized. Max. = maximized. Min. =
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Experimental code	Fresh Na ₂ SO ₄ Needed (kg/h)	Dilution factor (multiplier)	Raff. Ni (g/L)	Raff. Na (g/L)	D.S. produced (kg/h)	Equilibrium Na/REE (mol/mol)	La lost (%)
G1	0	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.
G2	Not opt.	Min.	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.
G3	Not opt.	Not opt.	Max.	Not opt.	Not opt.	Not opt.	Not opt.
G4	Not opt.	Not opt.	Not opt.	Min.	Not opt.	Not opt.	Not opt.
G5	Not opt.	Not opt.	Not opt.	Not opt.	Max.	Not opt.	Not opt.
G6	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.	2	Not opt.
G7	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.	Not opt.	Min.
G8	0	Not opt.	Max.	Not opt.	Max.	2	Not opt.

optimized: Ni concentration in raffinate and the yield of double salt, both were maximized. These two goals are in conflict: in order to maximize the double salt yield, more Na is required. However, the use of recirculated solution as a source of Na will produce more dilute Ni-containing raffinate. In this optimization run, the need for fresh Na₂SO₄ balance was also set to target a value of 0 t/h. Negative value would indicate an excess of Na originating from the recirculation, and a positive value would indicate the needed quantity of fresh Na₂SO₄.

These data originated in experiments G1–G8 was then input in Minitab in order to find optimal solutions for desired responses, as shown in Tables 7 and 8. The results based on modeling are shown in Fig. 3. This optimization mainly explains that 74% of the solution must be bled out in order to satisfy the responses involving G8. The reason for model to predict that the optimal PLS acid concentration is high (4 M), is that the acid concentration would slightly affect the density of the solution, and hence, the Ni concentration present in equal volumes. *D* is the desirability (0–1), and shows how well the set goals were achieved in terms of response optimization. Since G1–G7 were single response optimizations, the desired result was always achieved. In Fig. 3, the optimization scenario (exp. G8) is described as seen in Minitab. The results

clearly indicate that significant bleeding (47.6%) of the recirculated solution will be required in order to limit excessive dilution of the raffinate, while simultaneously maintaining high enough Na (Na/REE = 2.0) content by only using the recirculated solution (fresh Na = -0.294 kg/h) to cause satisfactory recovery of REEs as double salts (12.44 kg/h). It can also be seen that increase in REE concentration of PLS will cause decrease in raffinate Ni concentration. The [Na₂SO₄] has mostly no effect on anything as the source of Na is the recirculated solution (fresh Na₂SO₄ needed = 0).

The results showed that combination of DOE and process flowsheet modeling can be a powerful tool to quickly identify possible bottlenecks in process design. However, response surface methodology in itself is not enough to create a robust regression model of the entire flowsheet. The future work on this model should involve extension and addition of downstream processes for the raffinate and the mixed rare earth hydroxide. Furthermore, the double salt conversion to hydroxide and other parts of the simulation can be further refined with experimental data on equilibration and kinetics to complete the full process design. Additionally, more appropriate designs should be considered in creating models that better conform to the higher order responses.

Table 7 Optimized parameters for investigated G1–G8

Experimental code	D	[Na ₂ SO ₄] (M)	Target Na/REE (mol/mol)	3NaOH/REE (mol/mol)	Bleed %	PLS [acid] (M)	PLS REE (g/L)	L/S ratio (L/kg)
G1	1.00	1.25	2.00	2.00	71.28	0.5	12.00	100.00
G2	1.00	2.00	1.00	1.63	100.0	4.00	12.00	40.78
G3	1.00	2.00	1.00	1.00	100.0	4.00	12.00	60.22
G4	1.00	2.00	1.00	1.00	100.0	0.5	12.00	100.00
G5	1.00	1.2576	3	2.03	2.02	2.27	48	55.81
G6	1.00	0.5	3.00	2.80	79.29	0.5	48	12.50
G7	1.00	1.2576	1.87	2.03	0.00	2.23	30.18	55.81
G8	0.88	2.00	1.9652	1.00	47.62	4.00	48	74.36



Table 8 Results: the optimized responses, obtained by using parameters, Table 7

Experimental code	Fresh Na ₂ SO ₄ needed (kg/ h)	Dilution factor (multiplier)	Raff. Ni (g/L)	Raff. Na (g/L)	D.S. produced (kg/h)	Equilibrium Na/REE (mol/ mol)	La lost (%)
G1	0.00						
G2		- 0.4159					
G3			43.2563				
G4				_			
				1.7- 92			
G5				92	13.2413		
G6						2	
G7							_
							2 7- 54
G8	- 0.2939		24.7581		12.4446	2.0055	

3.2 Experimental Verification of the Process Flowsheet

Experimental work was conducted in order to verify that the chemistry suggested by Eqs. 1 and 2 and flowsheet is valid and technically feasible. Before circulation experiments, 20 mL of 355 g/L Na₂SO₄ was added to 80 mL of PLS. The obtained precipitate was the baseline mass (4.11 g) for the REE recovery. It was discovered with ICP-OES that the La yield was 99.1%. Recirculation experiments were performed in two rounds. In the first round, the recyclant obtained from the initial conversion was used. The results clearly

indicate that the circulation experiments can be performed with no carryover of impurities to recyclant as shown in Table 9 and Fig. 4. However, the wash waters of rare earth double sulfate both contained large quantities of base metals and rare earths. It is clear that vacuum was not sufficient in removing the remaining PLS from the double sulfate precipitate. Another key issue in the washing of the rare earth double sulfates is that they are slightly soluble in water [11]. It may be necessary to consider this by utilizing washing reagent where the solubility is minimized, e.g., hot water or hot sodium sulfate solution. It would also be beneficial to recycle the wash waters back to precipitation stage;

Fig. 3 Optimization results for the process flowsheet according to optimization settings of G8 as outlined in Table 6

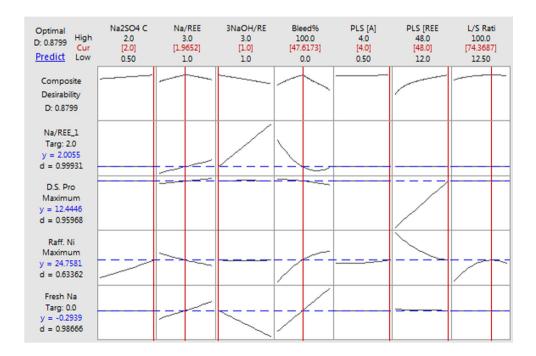




Table 9 Experimental input and output concentrations of elements in precipitation reactor (300-RC-001) and alkaline conversion reactor (300-RC-002) when circulating the solution internally (round 1, round 2). n.a. = not analyzed

		er #1				
ROUND 1		Feed In			Flow Out	
	(A)	(B)	(C)	(D)	(E)	(F)
Vol. (mL)	60	40	50	50	94	-
Mass (g)	-	-	-	-	-	2.74
Ni (mg/L)	57752	<1	n.a.	979	35416	n.a.
Co (mg/L)	7386	<1	n.a.	111	3951	n.a.
Mn (mg/L)	3528	<1	n.a.	53.7	2627	n.a.
Al (mg/L)	1259	<1	n.a.	20.3	797	n.a.
Zn (mg/L)	1857	<1	n.a.	26.6	1330	n.a.
Fe (mg/L)	601	<1	n.a.	8.3	410	n.a.
La (mg/L)	9796	<1	n.a.	462	195	n.a.
Na (mg/L)	86	>0.53 M	n.a.	240	4702	n.a.
S (mg/L)	n.a.	6740	n.a.	n.a.	n.a.	n.a.

		Alkaline Conversion Reactor #1 & Filter #2							
ROUND 1		Feed In		Flow Out					
	(F)	(G)	(H)	(I)	(K)	(J)			
Vol. (mL)	-	100	50	50	94	-			
Mass (g)	2.74	-	-	-	-	2.09			
Ni (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Co (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Mn (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Al (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Zn (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Fe (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
La (mg/L)	n.a.	n.a.	n.a.	<1	<1	n.a.			
Na (mg/L)	n.a.	0.35 M	n.a.	n.a.	>0.35 M	n.a.			
S (mg/L)	n.a.	n.a.	n.a.	n.a.	4580	n.a.			

ROUND 2		Feed In			Flow Out	
	(A)	(B)	(C)	(D)	(E)	(F)
Vol. (mL)	60	40	50	50	94	-
Mass (g)	-	-	-	-	-	2.56
Ni (mg/L)	57752	<1	n.a.	1147	35576	n.a.
Co (mg/L)	7386	<1	n.a.	134	3958	n.a.
Mn (mg/L)	3528	<1	n.a.	49.4	2312	n.a.
Al (mg/L)	1259	<1	n.a.	18.5	694	n.a.
Zn (mg/L)	1857	<1	n.a.	23.5	1163	n.a.
Fe (mg/L)	601	<1	n.a.	5.2	354	n.a.
La (mg/L)	9796	<1	n.a.	463	449	n.a.
Na (mg/L)	86	>0.35 M	n.a.	197	1903	n.a.
S (mg/L)	n.a.	4580	n.a.	n.a.	n.a.	n.a.



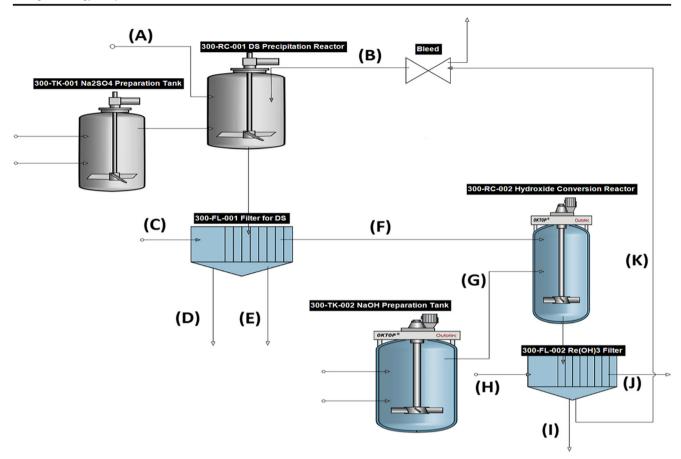


Fig. 4 Stream locations that are being depicted in Table 9

however, this would amount to dilution of the PLS which might be undesirable. 60 °C wash water was utilized in this study; however, some losses still occurred. The contents of La in the wash water is higher in relation to other metals than in the leach residue

wash water (0.17 vs. 0.47 La/Ni)—an indication that some of the double salts may have dissolved, regardless of the 60 °C wash water. Unsurprisingly, NaOH solutions were devoid of metals as the measured pH of these solutions was 13 after the conversion. No

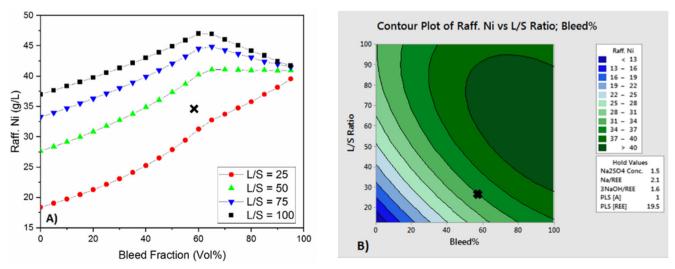


Fig. 5 The flowsheet model results (a) and metamodel results (b) at identical hold values. X marks the real experimental results



dissolved aluminum was found either. The sulfur content found in these solutions mirrors the dissociation of the REE double salt and its subsequent hydrolytic precipitation. By calculating the mole quantity of S in round 1 recyclant (14 mmol), one can find that the amount of S mirrors the amount of double sulfate (2.74 g of feed D.S. vs. 2.66 g D.S. based on sulfur content) that was fed to the experiments. The yield of lanthanum changed between round 1 and round 2 (96.9% and 92.8% respectively). This can be attributed to Eqs. (1) and (2): the amount of double sulfate obtained was less, and since rigid 3NaOH/REE = 1.6 ratio was used, the Na concentration in NaOH solution was less as it was subsequently too in the recyclant between the two experiments, decreasing the saturation degree in precipitation. This shows that the principle chemistry of flowsheet using the Na and SO₄ containing solution originating from the double salt conversion as a precipitation agent is valid. Utilization of this method will have potential to drastically reduce the consumption of Na in the processing of NiMH battery waste.

3.3 Observed, Modeled, and Metamodeled

The results showed that combination of DOE and process flowsheet modeling can be a tool to quickly identify possible bottlenecks in process design. However, response surface methodology in itself is not enough to create a robust regression model of the flowsheet. This can be used for early evaluation of the process optimization, to be further verified by experimental work in the desired conditions. Experimental work conducted verified that the modeled chemistry works in the laboratory scale and that recirculation of the process solution can result in internal usage of Na as precipitation agent for REE. To illustrate this, Fig. 5 shows the comparison of data from flowsheet model (Fig. 5a) and metamodel (Fig. 5b) at process held process condition. In this investigation, the effect of L/S ratio and bleed fraction (%) on raffinate Ni concentration is investigated. The experiments conducted at shown hold conditions show that it is in good agreement with the model however with slight variation. The experiments were performed at L/S = \sim 28, and Ni in raffinate was 35 g/L after precipitation experiments as shown in Table 9.

The future work on this model should involve extension and addition of downstream processes for the raffinate and the mixed rare earth hydroxide. Furthermore, the double salt conversion to hydroxide and other parts of the simulation can be further refined with experimental data on equilibration and kinetics to complete the full process design. Additionally, more appropriate designs should be considered in creating models that better conform to the higher order responses.

Furthermore, the optimization work can be improved by iteratively looking at ever smaller parameter space.

4 Conclusions

The current study was conducted in order to optimize chemical consumption in future NiMH recycling process: NiMH leaching, REE DS precipitation—NaOH conversion of REEs. Tools used to study the process, specifically internal solution circulation were HSC Sim and Minitab. Further the model was compared to experimental data.

Firstly, a process flowsheet was successfully designed for recirculation of sodium-rich solution originated in REE hydroxide conversion step during NiMH battery wastes treatment using HSC Sim software. Design of experiments was utilized in analyzing the variability in several responses (fresh Na₂SO₄ balance (kg/h), dilution factor, raffinate Ni conc. (g/L), raffinate Na conc. (g/L), double salt produced (kg/h), and La lost%) in the created flowsheet simulation model. By combining these methods, it was possible to acquire information on how to optimize the re-use of sodium and sulfate ions present in the rare earth double sulfate. CCD was employed in forming quadratic regression models for the responses. Generally, good regression ($R^2 > 95\%$) was achieved for most of the responses, but results for some response models (La lost%, R7) were not optimal, indicating that quadratic models were not able to model all the responses. Furthermore, the researcher must understand the limitation of applying CCD, namely, the model being (a) polynomial in form and (b) its tendency to smooth out the data. It was shown that combination of process simulation, DOE, and statistical software can be useful to identify bottlenecks in a simple process model. In order to maximize the Ni concentration in the raffinate, along with maximal double salt production and optimal Na usage, 47% of the recirculated solution must be bled out of the loop.

Secondly, experiments were performed in observing whether recycling the solutions would be technically viable and to compare one experimental point/result with the modeled conditions. Solutions could be successfully recirculated, and the yield of La was between 90-95%. The yield depended upon the NaOH concentration and S/L ratio used in double salt conversion as it directly affects the Na concentration of the recyclant. The results suggest that metamodeling of flowsheet model and process flowsheet modeling can be a powerful tool to quickly identify possible behavior in the process design. However, in the development of real hydrometallurgical recycling process, it is essential to conduct real experiments in the defined conditions in order to verify the functionality of the said flowsheet. It was shown in this work that circulation flowsheet was functional, albeit at the cost of dilution.



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Compliance with Ethical Standards

Conflict of Interest The authors declare that they have no conflict of interest

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